



## Current and future prospects for heat recovery from waste in European district heating systems: A literature and data review

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*Published in:*  
Energy

*Link to article, DOI:*  
[10.1016/j.energy.2015.12.074](https://doi.org/10.1016/j.energy.2015.12.074)

*Publication date:*  
2016

*Document Version*  
Peer reviewed version

[Link back to DTU Orbit](#)

*Citation (APA):*  
Persson, U., & Münster, M. (2016). Current and future prospects for heat recovery from waste in European district heating systems: A literature and data review. *Energy*, 110, 116–128.  
<https://doi.org/10.1016/j.energy.2015.12.074>

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35 also an alternative fuel for power and heat generation; energy recovery from waste represents  
36 an effective measure to reduce landfilling and avoid disposal emissions while simultaneously  
37 reducing the equivalent demand for primary energy supply. A key factor for obtaining the full  
38 synergetic benefits of this energy recovery is the presence of local heat distribution  
39 infrastructures, without which no large-scale recovery and utilisation of excess heat is  
40 possible. In this paper, which aims to estimate municipal solid waste volumes available for  
41 heat recovery in European district heating systems in 2030, a literature and data review is  
42 performed to establish and assess current and future EU waste generation and management.  
43 Main conclusions are that more heat can be recovered from current Waste-to-Energy facilities  
44 operating at low average heat recovery efficiencies, that efficient incineration capacity is  
45 geographically concentrated, and that waste available for heat recovery in 2030 is equally  
46 determined by total generation volumes by this year as by future EU deployment levels of  
47 district heating.

48

## 49 **1 Introduction**

50 Generation of waste is a symbiotic, unfortunate, and potentially detrimental consequence  
51 of producing and consuming material goods and services in the world today. Waste,  
52 furthermore – if not treated properly – endangers the sanity of local environments, affects  
53 the equilibrium of regional and global ecosystems, and may eventually threaten the very  
54 foundation for healthy and sustainable life conditions in any community [1-4]. For this reason,  
55 waste management represents a critical societal function needed to meet the challenges  
56 raised by generated waste flows, be they municipal solid wastes (MSW), industrial wastes, or  
57 any other kind of material residues.

58 In the European Union (EU) context, these circumstances has been apprehended for at  
59 least half a century and plausible approaches and solutions to reduce total generation  
60 volumes, as well as to arrange proper management measures to treat these, have been  
61 conceived and formulated in environmental action programmes [5, 6], in legislation [7-11],  
62 and in numerous projects and studies [12-19]. Most urgent, as it represents the least attractive  
63 waste management option due to e.g. uncontrolled greenhouse gas emissions, has been  
64 efforts to reduce landfilling and land deposits (expressed as early as 1999 in the so called  
65 Landfill Directive [20]), an unfortunate practice among EU Member States (MS) still accounting  
66 for nearly 40% of MSW treatment in 2010 [21]. More profoundly, in the recognition of latent  
67 interdependencies between waste generation, population growth, and economic activity, the  
68 idea of decoupling human well-being from virgin resource use (resource decoupling) have  
69 constituted the ground foundation of EU waste policy for the last couple of decades.

70 If not by landfilling then, according to the principles outlined in most recent EU waste  
71 legislation, the Waste Framework Directive [22], waste management should consist of not

72 one, but many different treatment options reflecting the characteristics and properties of  
73 different waste fractions, and should ideally be initially prevented or minimised. In order of  
74 preference, treatment options should moreover be arranged in the following priority order:  
75 re-use, recycling, composting, incineration (with and without energy (and heat) recovery), and  
76 – when all other options are exhausted - landfilling (deposit). To underline this fundamental  
77 strategy for the coming years, a revision of current EU waste legislation has been suggested in  
78 2013 [23] and a proposal for a new coherent waste management policy aligned with the  
79 overarching concepts of a circular economy [24, 25] and an energy union [26] is now under  
80 discussion [27].

81 With this clear aim for future European waste management to move “up the waste  
82 hierarchy” one immanent, topical, and unneglectable issue concerns the forthcoming fate of  
83 energy recovery from waste, or Waste-to-Energy (WTE) [28-30]. It is understood that, in a  
84 realised circular economy future, no or little recyclable waste fractions should remain  
85 available for this definitive conversion of materials into electricity, heat, flue gas and ashes (at  
86 least not non-hazardous fractions), so it is fair to wonder what waste volumes eventually will  
87 be left for energy recovery in view of these ambitions?

88 As will be further investigated and presented in this paper, this question may however not  
89 be entirely answered merely by correctly forecasting future European waste generation  
90 volumes (which in itself is an arduous and, by nature, speculative undertaking), nor by justly  
91 interpreting the scope and credibility of future waste legislation targets. The equitable answer  
92 need as well to consider the future development of more energy efficient supply structures  
93 for the provision of space heating and domestic hot water in European buildings. If this  
94 development is to be characterised by a comprehensive and continued expansion of district

95 heating systems (mainly in urban areas), as outlined and discussed amongst others in [31-34]  
96 and modelled inter alia in [35-37], the European community will have gained additional access  
97 to necessary infrastructures by which to facilitate heat recoveries from waste designated to  
98 energy recovery.

99       Such a development, however, is by no means given for Europe of tomorrow, despite  
100 ambitious legislation to bring forth national heating and cooling plans, heat and cold synergy  
101 mapping, and improved structural energy efficiency, as mediated inter alia in the Energy  
102 Efficiency Directive of 2012 [38, 39]. As will be discussed further in this paper, European  
103 district heating has, on average, evolved markedly in absolute terms during the last 20 years,  
104 especially so in commercial and public service sectors, but not significantly so in relative terms  
105 – and not coherently so in terms of geographical distribution. This latter circumstance itself  
106 has an indirect influence on future prospects for heat recovery from waste since a continued  
107 disproportionate distribution of heat recovery infrastructures may reflect in an as well skewed  
108 distribution of efficient WTE facilities, hereby perhaps intensifying an already untenable  
109 situation of continental exports and imports of waste. Possible magnitudes of future European  
110 heat recoveries from WTE activities must therefore be conceived as a consequence of several  
111 prospective developments, where e.g. behavioural, demographical, economic, political, but,  
112 not least, infrastructural dimensions of society all appear significant.

113       The main aim of this paper is to add perspective and some clarity as to what can be  
114 expected of European heat recovery from waste in the future, mainly by illustrating the  
115 historic development and plausible future progress of MSW generation, MSW management,  
116 and district heating deployment in Europe, as well as to discuss some contextual concerns and  
117 issues. As with all complex matters, characterised by multiple influences and mutual

118 interdependencies, of which seldom all are identifiable and perhaps even less so quantifiable,  
119 it is appropriate first to outline an initial general overview. The purpose of this study is  
120 essentially to establish such a holistic understanding at EU MS level, hence principally devoting  
121 little or no attention to local-, technology-, or regional policy-specific issues undoubtedly  
122 related to the continental scope topics at hand. The structural nature of the questions raised  
123 in this study is rather; what general tendencies are observable in past records? What are  
124 current MS national and EU continental state of affairs? What can be expected in the future,  
125 given these historic trends, current conditions, model predictions, and outspoken ambitions  
126 for the years to come? More specifically, this paper aims to answer the following three  
127 research questions:

128

- 129 • What are the historical trends of MSW generation and management among EU MS  
130 from 1995 to 2012?
- 131 • What models have been developed and used in recent years to assess future EU MSW  
132 generation and district heating deployment, and what are their predictions for 2030?
- 133 • According to model predictions and scenario targets, how much MSW is likely to be  
134 available for thermal energy conversion and heat recovery in EU district heating  
135 systems in 2030?

136

137 Given the general uncertainty of future assessments, the answers presented to the latter  
138 two of these questions must be interpreted as indicative and approximate only. The purpose  
139 here is not claiming to have identified the most likely future development for European waste  
140 generation and heat recoveries from these waste flows, a development subject to several

141 additional sector influences not considered here (e.g. changing characteristics of waste,  
142 technical changes and modifications of WTE technologies), but merely to make explicit what  
143 past and present modellers have assessed plausible. In this respect, reviewed models and  
144 associated assumptions (in the case of regression models, chosen explanatory variables), are  
145 viewed and accepted as is, and no model analyses are performed in this context. The study  
146 limitations, further, exclude any technical or commercial feasibility assessment of future WTE  
147 technologies or heat recoveries from waste in European district heating systems, estimates  
148 which despite their general relevance and interest are beyond the study scope and objective.  
149 Additional study limitations concern related waste incineration topics, such as ash deposits,  
150 air emissions (e.g. greenhouse gases), and issues concerning the production of refuse derived  
151 fuels (RDF), as well as waste trade related issues, such as gate fees, public opinion, and  
152 economic incentives. Finally, given the immense publication rate within the field of waste  
153 management during the last 25 years (a Scopus search 2015-09-15, 1990 to present, returned  
154 4,369 articles with “waste management” in the paper title!), the literary sources referred to  
155 in this work represent a sample reflective of the study objectives rather than constituting a  
156 complete review of this ample flow.

## 157 **2 Materials and Methods**

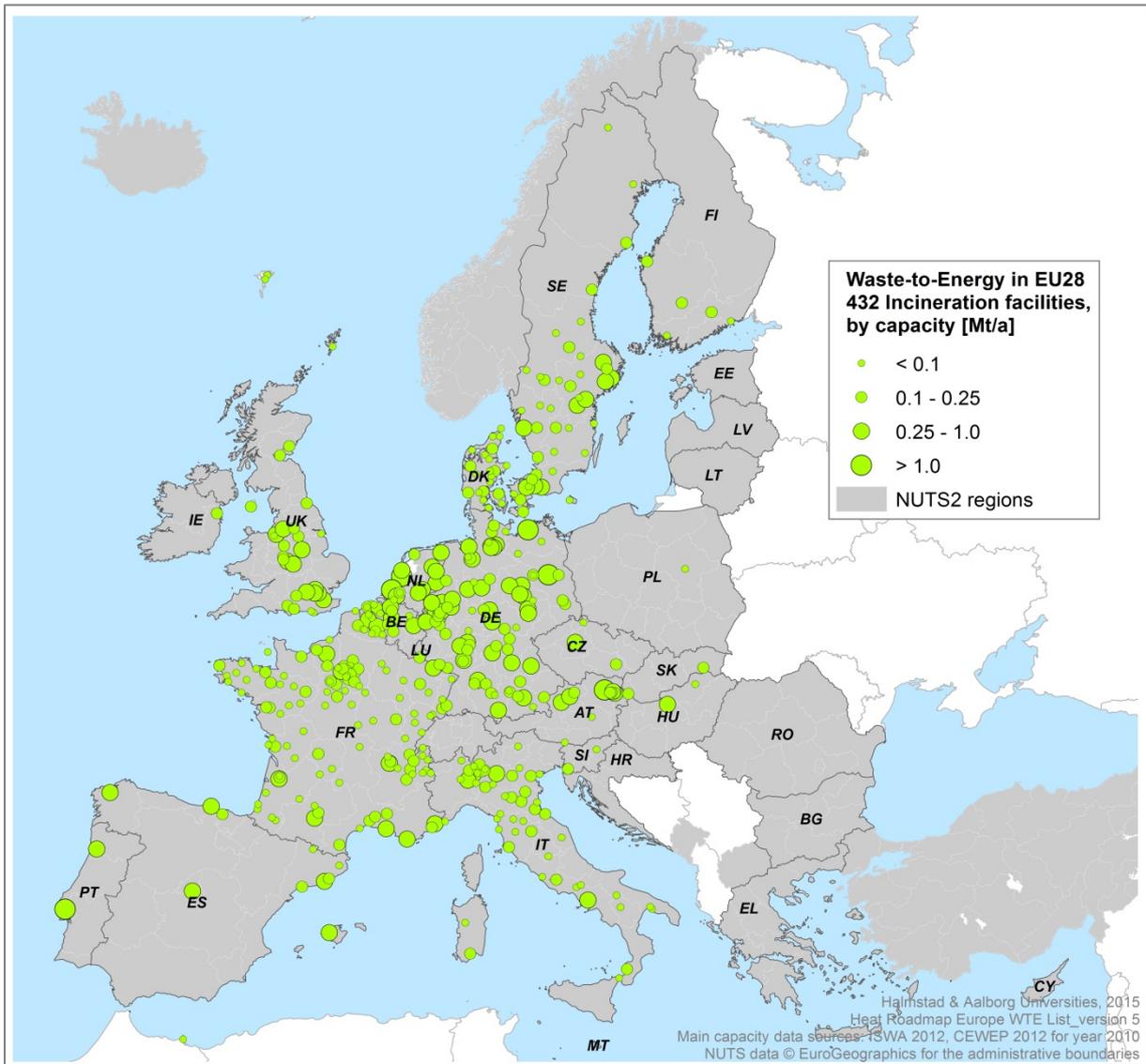
158 The main focus in this paper is EU MSW, which in terms of total generation in recent years  
159 has hovered at around some 230 to 250 million metric tonnes (Mt) per year (EU28) [40]. For  
160 the sake of reference, however, it should be noted that the total sum of all generated waste  
161 in EU28, i.e. mineral, metal, animal and vegetal, chemical and medical, textile, glass, plastic,  
162 and sludge wastes etc., amounted to no less than 2.51 billion metric tonnes (Bt) in 2012,

163 according to [41]. From a perspective of MSW, hence, the annual volumes considered here  
164 represents merely one approximate tenth of total generation volumes, and waste destined  
165 for energy recovery herein represents itself merely a fraction of total MSW generation. It  
166 should also be noted that current heat recoveries from MSW incineration represent less than  
167 10% of total heat supplies to European district heating systems, which are dominated by  
168 natural gas, other bituminous coal, and primary solid biofuels.

169 At current, if neglecting anaerobe digestion processes, European energy recovery from  
170 WTE activities is more or less solely performed by Rankine steam cycle incineration processes,  
171 where waste is – either directly or by co-firing – burnt to generate steam and hot flue gases.  
172 Gasification of waste, associated with both pros and cons compared to incineration, e.g. an  
173 intermediate product (syngas) with a wider array of applications (e.g. fuel production),  
174 potential for higher conversion efficiency (e.g. integrated gasifier combined cycles), but syngas  
175 being toxic and potentially explosive [42], has so far seen very limited commercial use in  
176 Europe. For the future, however, gasification (and pyrolysis) of waste may very well become  
177 a challenger in European WTE, considering additional benefits such as reduced generation of  
178 pollutants (dioxins and NO<sub>x</sub>), lower operation temperatures, and more efficient material  
179 recovery, e.g. metals [43-45].

180 Consisting in essence of incineration facilities only then, the stock of designated EU WTE  
181 plants currently in operation (2011) was assessed in 2013 as a sub-task during the second EU  
182 pre-study of the Heat Roadmap Europe project [46, 47]. As outlined in Fig. 1, this estimation,  
183 which aligned all reported EU28 plants in the 2012 ISWA State-of-the-Art Report (397) [48]  
184 with georeferenced facilities reported in the European Pollutant Release and Transfer Register  
185 [49], while using as well some additional sources [50, 51], identified 432 facilities with

186 geographical location and annual incineration capacity. The total annual incineration capacity  
187 of these facilities was assessed to ~86 Mt, and, as can be seen in Fig. 1, the majority of these  
188 facilities are located in Central and North-Western EU Members States.



189  
190 **Fig. 1. 432 designated EU28 waste incineration facilities in operation during 2011, by location and by assessed**  
191 **annual capacity. Result from the 2<sup>nd</sup> European pre-study in the Heat Roadmap Europe project (2013). Sources:**  
192 **[46-51].**

193 As established in the Confederation of European Waste-to-Energy Plants (CEWEP) report  
194 [52], the average energy content of European MSW is ~10 GJ/t, expressed as net calorific heat  
195 value. If accepting this value as a general conversion factor, the Heat Roadmap Europe

196 assessment is well in consonance with 2010 primary energy and energy recovery volumes  
197 from waste reported in the extended energy balances of the International Energy Agency  
198 (IEA). Herein [53], total primary energy supplies from industrial and municipal (renewable and  
199 non-renewable) waste flows destined for energy conversion summed up to 800 PJ in this year,  
200 hence equivalent to ~80 Mt. If neglecting industrial waste (152 PJ), the two fractions of  
201 municipal waste represented all together 648 PJ (333 PJ renewable and 315 PJ non-  
202 renewable), which would correspond to some 65 Mt in total under given assumptions. By a  
203 municipal waste electricity output of 119 PJ and a heat output of 159 PJ, it is fair to anticipate  
204 an average overall conversion efficiency of 43% and a heat recovery efficiency of  $159 \text{ PJ} / 648$   
205  $\text{PJ} = 0.245 \approx 25\%$  from EU MSW incineration (quota of recovered heat and primary input, see  
206 [32, 36] for further references).

207       And the concept of conversion efficiency truly lies at the heart of European WTE today,  
208 since the R1 formula (a performance indicator of waste incineration facilities), was introduced  
209 in the Waste Framework Directive in 2008. Taking into consideration both electricity and heat  
210 output from WTE conversions, and accentuating generated electricity by a factor 2.6 (and  
211 generated heat by a factor 1.1), the total energy yield is weighted against the energy content  
212 of the combusted waste [54]. By this regulation it has become possible for efficient WTE plants  
213 to be classified as "energy recovery" operations (R1) rather than "waste disposal" activities  
214 (D10), by achieving an efficiency higher than 0.6 (for plants in operation before 2009-01-01)  
215 and higher than 0.65 (after 2008-12-31). In 2009, around 66% of European WTE plants were  
216 classified as recovery plants, according to the formula [52], and among multiple benefits for  
217 these facilities can be mentioned e.g. lower taxes, access to imported waste (without violation  
218 of proximity and self-sufficiency principles), and eased admission to bank finance [55]. Albeit

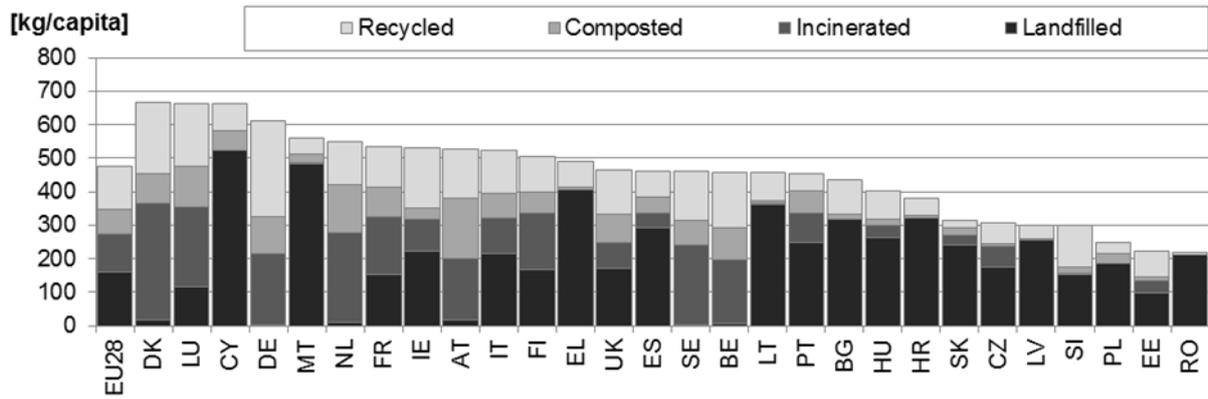
219 the pronounced promotion of generated electricity, and as well allowance to include some  
220 on-site used energy output in the calculation, achieving the R1 efficiency factor is more viable  
221 if having cogeneration options available compared to electricity only generation.

## 222 **2.1 Past and current waste generation and management**

223 The data used to assess historic developments and current conditions in this study are all  
224 publicly available statistics from Eurostat. These include time series data from 1995 to 2012  
225 on EU MS MSW generation and management [40] and national and NUTS2 regional total  
226 population counts [56]. On NUTS2 regional level, corresponding MSW generation data has  
227 been made partly available by a promising – however yet to reach full continental stretch –  
228 pilot project covering the years from 2000 to 2011 [57], which also comprises precious data  
229 on number and capacities of recovery and disposal facilities at this regional level [58].

230 By use of this information, specific MSW volumes for EU28 and its MS during 2012 are  
231 illustrated in Fig. 2. This figure is similar in arrangement to the 2008 projection presented by  
232 Persson and Werner in 2012 [59], and by comparison it can be observed that average EU MSW  
233 per-capita volumes has decreased from approximately 520 kg/capita (EU27, 2008) to ~475  
234 kg/capita (EU28) during these five years. In terms of waste management, however, landfilling  
235 still constitutes the largest treatment option share in relative terms (34%), but it is noteworthy  
236 that the corresponding share was 40% in 2008.

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**Fig. 2. Distribution of municipal solid waste treatment in EU28 Member States in 2012, by waste hierarchy order categories. Source: [40, 56].**

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## 2.2 Future waste generation and model scenarios

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To facilitate an assessment of future European MSW generation and management, mainly so for the year 2030, a literature review was performed by which to learn more about what typical approaches, models, and projections, have been conceived and developed in a devotion to this purpose during recent years. Dating back to the early 1990s, aggregated US industry sector waste generation was modelled by Ayres et al. [60, 61] using material balances, a methodological approach attentive of the mass difference between commodity inputs (materials, substances etc.) and corresponding commodity outputs from considered sectors (see also the Leontief input-output economic model [62]). Simultaneously, in Europe, extending from the input-output approach, general equilibrium model applications by, for example, Alfsen et al. [63, 64] and Conrad [65], embraced environmental, emission, and energy system dimensions into the modelling (hence considerably widening the analytical scope), a discourse essentially originating in earlier works of e.g. Johansen [66, 67], Edmonds and Reilly [68, 69], and others [70-73].

In 1997, a novel study by Bruvoll and Ibenholt [74], targeting Norwegian manufacturing industry, presented national projections of solid waste generation by use of a macroeconomic

257 model, i.e. a computational multi-sectoral equilibrium model. Considering aspects such as  
258 technological change, price substitution (value of resources vs labour), and other key  
259 economic variables, they anticipated a total 64% increase of waste generation between 1994  
260 and 2010. In Sweden, computational general equilibrium models were later developed and  
261 used by Östblom et al. [75-77], principally based on assuming direct linkage between waste  
262 generation and economic activity of firms and households; future waste generation hence  
263 being reflected in economic growth and the relative use of production factors. This approach,  
264 the so-called EMEC model, was also used by the Swedish Environmental Research Institute  
265 (IVL) in 2010 to model future national waste generation volumes in 26 industrial sectors, the  
266 public sector, and in Swedish households [78].

267       Still within, and inherently rooted to, classical economic theories, another generation of  
268 models adhered to the principal idea of relating waste generation to e.g. economic activity  
269 and population growth, but, as they utilise econometrics, with a conceptually different  
270 methodological approach. Statistical investigations and models, by means of simple or  
271 multiple regression analysis, characterised this new vein of European studies emerging in the  
272 early 2000s – and especially so constant elasticity models, characterised by generating relative  
273 parameters and outcomes [79]. Through a series of reports and book chapters [80-82], the  
274 Copenhagen based European Topic Centre (ETC) developed and presented, in 2007, an  
275 econometric European model for waste and material flows [83], later also used to produce  
276 EU27 MSW generation and treatment predictions for 2020; the ETC 2008 [84] and ETC 2011  
277 [85] projections.

278       By 2014, after an institutional name change to the Copenhagen Research Institute, the  
279 ETC model was incorporated in the mass flow module of the “European Reference Model on

280 Municipal Waste Generation and Management”, developed for the European Commission and  
 281 the European Environment Agency (EEA) [86, 87]. This model (hereafter referred to as the “EU  
 282 Ref. model”), projects EU28 MSW waste generation up to year 2030 and is structured as a set  
 283 of operational modules considering a wide range of influential conditions (e.g. prevention,  
 284 collection, employment, financial costs, environmental impact etc.). A thematic overview of  
 285 notable model approaches reviewed in this paper are presented in Table 1, which also includes  
 286 the 2010 Arcadis/Eunomia bio-waste projection for EU27 up to year 2020 (including as well a  
 287 projection of total MSW generation by this year) [88].

288 **Table 1. Examples of notable waste generation model approaches developed and used during the period 1994**  
 289 **to 2014, by principal model type, flow, and projection year**

Author/Name	Published	Model type	Flow	Target	Year	Reference(s)
Ayres et al.	1994	Input/Output	Ind.	USA	-	[60, 61]
Bruvoll & Ibenholt	1997-2003	Gen. equilibrium	Ind.	Norway	2010	[74, 89]
Östblom et al.	2006-2010	Gen. equilibrium	Ind./MSW	Sweden	2030	[75-77]
Sundquist et al.	2010	Gen. equilibrium	Ind./MSW	Sweden	2030	[78]
Møller Andersen et al.	2005-2011	Econometrics	MSW	EU27	2020	[80-85]
Arcadis/Eunomia	2010	Mathematical	MSW	EU27	2020	[88]
EU Ref. model	2014	Econometrics/Modules	MSW	EU28	2030	[86]

290

291 As the most recent, comprehensive, and only EU model assessment extending to 2030,  
 292 the EU Ref. model is used in this work as the key benchmark trajectory by which to evaluate  
 293 plausible prospects for future European heat recovery from waste. Appropriately, for this aim,  
 294 the baseline and five scenario projections elaborated in the model [86], as tabulated in Table  
 295 2, provide a suitable framework for the study analysis. Current legislation, i.e. the Waste  
 296 Framework Directive, stipulates a 50% recycling, or preparation for re-use, target for EU MS  
 297 by 2020, and, correspondingly, the majority of MS shall have accomplished a 35% landfill limit  
 298 by 2016 according to the Landfill Directive. It is likely that current revisions of these regulations  
 299 will propose much more stringent targets; perhaps as high as 80% recycling shares and  
 300 complete landfill bans for all recyclable and biologically degradable wastes. Given these

301 circumstances, Scenario 4 from the EU Ref. model assessments, i.e. 70% recycling and a 5%  
 302 landfill limit by 2030 (highest ambition level), constitute the reference case used here.

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305 **Table 2. Baseline (Business-as-Usual) and five scenarios evaluated by the EU Ref. model on waste generation**  
 306 **and management, 2014. Sources: [20, 22, 86, 90]**

	Description	Recycling/Preparation for reuse target	Landfill limit	By year
Baseline	Steady state waste management	-	-	-
Scenario 1	Full implementation of existing targets	50% <sup>a</sup>	35% <sup>b</sup>	2020/2016
Scenario 2.1	Increased recycling, no limit on landfill	60%	-	2030
Scenario 2.2	Further increased recycling, no limit on landfill	70%	-	2030
Scenario 3	Full implementation of existing targets, landfill limit	50% <sup>a</sup>	5%	2020/2030
Scenario 4	Further increased recycling, landfill limit	70%	5%	2030

307 <sup>a</sup> Referring mainly to household waste as expressed in the Waste Framework Directive.

308 <sup>b</sup> Referring to biodegradable municipal waste fractions as expressed in the Landfill Directive. For MS that had >80% landfilling in 1995, the  
 309 target year is extended to 2020.

### 310 **2.3 European district heating**

311 Until recently, there have been very few continental assessments regarding number,  
 312 distribution, and energy magnitudes of European district heating systems. In 2006, the  
 313 Intelligent Energy Europe (IEE) supported project Ecoheatcool [91], presented a  
 314 comprehensive estimate based on energy statistics from the IEA, where reported sold heat  
 315 was interpreted, principally, as district heat. For 1992 and 2003, the study concluded that  
 316 district heat deliveries amounted to 1.86 EJ and 1.76 EJ (EU25), respectively, indicating a slight  
 317 decrease – in absolute terms – during the time span considered. In the case of number and  
 318 distribution, it has been apprehended during the development of the Halmstad University  
 319 District Heating and Cooling database [92] that some 6000 systems currently are in operation  
 320 (see also [91]), and that these systems are fairly wide spread among MS (for a map on  
 321 geographical locations of EU district heating systems, see [36]).

322 More recently, as detailed in e.g. [32, 93], district heating has been anticipated to  
323 represent ~12% of the EU residential and service sector heat market (still heavily dominated  
324 by fossil fuels, e.g. natural gas), and consisted of ~1.6 EJ out of a total final heat demand of  
325 13.1 EJ in 2010, according to the up-to-date Stratego assessment [37]. The winter season in  
326 2010, however, was considerably colder than an average year, why a weather-corrected  
327 comparison of 1995 and 2012 data was performed. Assuming, as in [91], that reported heat  
328 sales in the IEA energy balances [94] represent district heat deliveries, and by use of Eurostat  
329 reported heating degree-day statistics for the time period 1980 to 2009 [95-97], a more  
330 comparable assessment of EU28 and MS district heating developments is presented in Table  
331 3. In relative terms, and keeping in mind some minor deviances in terms of MS heating degree-  
332 day calculation practices, EU district heating has on average expanded by 20% during the  
333 considered time-period, especially so in service sectors, albeit with large national variations.  
334 Since some constituents of the heat demand (domestic hot water and industrial process heat)  
335 are weather independent (not compensated for here), this comparison serves however  
336 merely an indicatory purpose. Noteworthy, the average drop in heating degree-days has been  
337 16-18 units per year since 1980.

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350 **Table 3. Weather corrected district heat deliveries to residential, service and industrial sector buildings in 1995**  
351 **and 2012, by EU28 Member States. Energy volumes in [PJ/a]. Sources: [94-97]**

MS	HDD <sub>A</sub> <sup>a</sup>	1995					2012					Difference [%]			
		f <sub>HDD</sub> <sup>b</sup>	Res <sup>c</sup>	Ser	Ind <sup>d</sup>	Tot	f <sub>HDD</sub> <sup>e</sup>	Res <sup>c</sup>	Ser	Ind <sup>d</sup>	Tot	Res	Ser	Ind	Tot
AT	3,540	0.98	15	15	4	34	1.08	37	32	13	82	141	119	192	138
BE	2,830	1.04	0.9	0.6	8	10	1.13	2	4	24	30	130	671	187	210
BG	2,654	0.97	31	1	80	112	1.07	16	4	23	44	-47	268	-71	-61
CY	762	0.90	0	0	0	0	1.17	0	0	0	0	-	-	-	-
CZ	3,533	0.98	41	18	91	150	1.09	47	20	29	96	14	10	-68	-36
DE	3,199	0.98	291	0	69	360	1.10	156	89	234	478	-46	-	238	33
DK	3,438	0.97	59	26	4	89	1.13	78	36	5	120	32	42	18	34
EE	4,393	1.04	23	1	2	26	1.08	15	5	2	22	-32	264	-8	-14
EL	1,642	1.00	0	0	0	0	1.09	2	0	0	2	-	-	-	-
ES	1,831	1.20	0	0	0	0	1.09	0	0	0	0	-	-	-	-
FI	5,774	1.02	82	0	9	91	1.09	124	0	71	195	52	-	655	115
FR	2,459	1.03	24	0	0	24	1.11	75	34	8	116	214	-	-	389
HR	2,561	0.96	5	0.8	4	10	1.11	6	2	2	11	17	117	-37	5
HU	2,886	1.00	32	17	5	54	1.08	24	7	14	45	-25	-61	205	-17
IE	2,871	1.07	0	0	0	0	1.12	0	0	0	0	-	-	-	-
IT	1,949	1.01	0	0	0	0	1.15	35	5	124	164	-	-	-	-
LT	4,048	1.02	35	8	7	51	1.08	22	9	10	41	-36	8	31	-19
LU	3,164	1.01	0	0	0	0	1.13	0	3	1	3	-	-	-	-
LV	4,220	1.03	26	11	2	39	1.08	17	6	0.3	24	-33	-42	-86	-38
MT	543	1.21	0	0	0	0	1.40	0	0	0.1	0.1	-	-	-	-
NL	2,854	1.01	7	12	55	74	1.14	12	24	51	87	71	107	-9	17
PL	3,574	0.99	263	23	77	363	1.08	195	43	32	269	-26	89	-59	-26
PT	1,278	1.39	0	0	2	2	1.09	0.3	1	15	16	-	-	588	645
RO	3,092	0.98	131	0	51	183	1.08	43	11	13	66	-67	-	-75	-64
SE	5,387	0.98	80	51	14	145	1.09	125	62	19	205	56	22	30	41
SI	3,024	0.99	4	4	1	9	1.09	4	2	2	8	-2.8	-42	109	-3.7
SK	3,416	1.00	16	12	1	29	1.08	22	6	7	34	33	-53	475	17
UK	3,081	1.03	0	0	0	0	1.10	2	18	37	57	-	-	-	-
<b>Tot</b>	<b>3,000</b>	<b>0.99<sup>f</sup></b>	<b>1,168</b>	<b>198</b>	<b>488</b>	<b>1,854<sup>g</sup></b>	<b>1.10<sup>f</sup></b>	<b>1,062</b>	<b>421</b>	<b>734</b>	<b>2,217<sup>h</sup></b>	<b>-9</b>	<b>112</b>	<b>50</b>	<b>20</b>

352 <sup>a</sup> Average annual Heating Degree Days (HDD<sub>A</sub>) per MS based on 1980 to 2009 time series data reported in [95, 97].  
353 <sup>b</sup> Multiplier, f<sub>HDD</sub>, is the ratio of HDD<sub>A</sub> over actual HDD for the year at hand. f<sub>HDD</sub> > 1 indicates a year warmer than the normal year.  
354 <sup>c</sup> Residential sector volumes include Non-specified (other) sector heat.  
355 <sup>d</sup> Reported industrial heat volumes may include some heat used on-site for internal purposes.  
356 <sup>e</sup> Multipliers for 2012 established by simple linear regression of MS 1980 to 2009 time series data and extrapolated to this year.  
357 <sup>f</sup> Total multipliers are fictive (reflecting the sum of MS weather corrected district heat volumes relative the sum of corresponding MS raw  
358 data heat volumes) and does not necessarily represent actual EU28 HDD factors for 1995 or 2012.  
359 <sup>g</sup> Raw data for 1995: 1,874 PJ.  
360 <sup>h</sup> Raw data for 2012: 2,014 PJ  
361

362 In terms of future European district heating deployment levels, no formal targets, as for  
363 waste management, have so far been established. Evidence, however, that district heating  
364 could expand cost-effectively by three times compared to current levels (essentially in inner-  
365 city areas) have been provided in [31], and – partly by this rationale – future expansion levels  
366 considering 30% (2030) and 50% (2050) shares of the total EU27 residential and service sector  
367 heat market have been modelled within the Heat Roadmap Europe context [35]. As perceived  
368 herein, the potential for WTE heat recovery in district heating systems was assessed at 330 PJ  
369 (~33 Mt) in 2030 and 585 PJ (~59 Mt) in 2050. At the current, average, EU28 heat recovery  
370 efficiency (25%), such heat recovery volumes would need to correspond to total MSW  
371 volumes designated for energy conversion of approximately 132 Mt and 236 Mt, respectively.  
372 If, hypothetically, average heat recovery efficiencies were to increase to 30%, 40% or 50% in  
373 the future, corresponding MSW volumes destined for energy conversion would be  
374 respectively; 110, 83, and 66 Mt by 2030, and 197, 148, 118 Mt by 2050.

### 375 **3 Results**

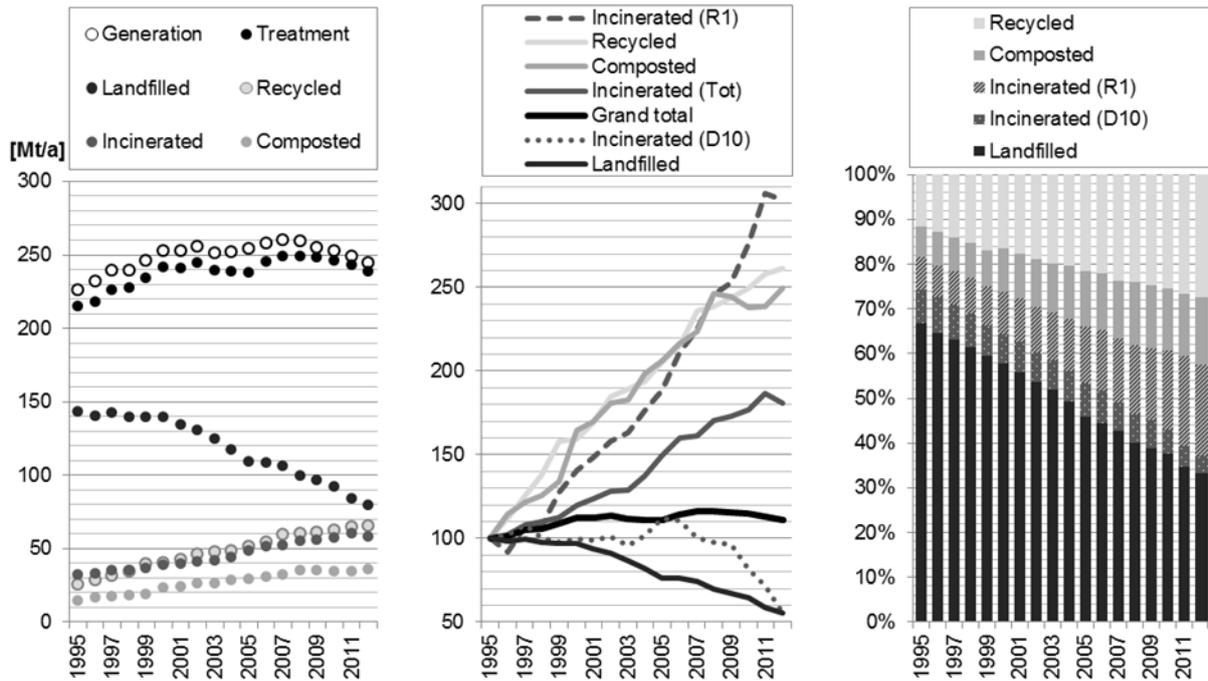
376 The results presented in the following are all based on data, literature, models, and  
377 methodological approaches, as described and accounted for above, and further ordered in  
378 alignment with the initial research questions asked. Hereby, study results refer firstly to  
379 historical and current EU MS MSW generation and treatment volumes from 1995 to 2012,  
380 secondly to model predictions of future MSW generation volumes for 2020 and 2030, and  
381 thirdly to assessed MSW volumes available for heat recovery in European district heating  
382 systems by 2030. To complement the latter, an arbitrary sensitivity analysis, where the two

383 variables; (i) predicted total EU MSW generation in 2030 and (ii) average EU heat recovery  
384 efficiency by this year, are allowed to take on alternative values.

### 385 **3.1 Waste generation and treatment – 1995 to 2012**

386 MSW generation in EU27 saw steady annual increases during the late 1990s and marked  
387 a thitherto all-time high of 256 Mt in 2002. Succeeded by a temporary decline during 2003 to  
388 2005, new record-breaking generation volumes were once again recorded in 2006 (258 Mt)  
389 and 2007 (260 Mt), to be followed by marginal, but consecutive, annual declines leading up to  
390 year 2012 (see Fig. 3, at left). Whether the apparently incessant decrease from 2008 and  
391 onwards reflects genuine behavioural (or structural) changes among European citizens and  
392 communities (and if so with decisive implications for future generation volumes), is an issue  
393 resolvable only at the access of more recent data. It is clear, however, that European efforts  
394 to reduce landfilling has been expedient during the considered period; representing a drop  
395 from 143 Mt (1995) to 79 Mt (2012). This significant achievement is reflected in corresponding  
396 and coherent increases in annual MSW volumes designated for incineration, recycling, and  
397 composting. Since Eurostat statistics subdivide incineration volumes in two categories; with  
398 energy recovery (R1) and without energy recovery (D10), they provide insight into the  
399 distribution of incineration volumes by these two treatment options, see Fig. 3 at centre.  
400 Among all treatment categories, WTE with energy recovery represents the fastest growing  
401 alternative, indexing above 300 during the considered time interval.

402



403  
 404 **Fig. 3. EU27 MSW generation and management from 1995 to 2012. At left, total generation, treatment, and**  
 405 **category volumes. At centre, indexed developmental trends by category with base year 1995=100. At right,**  
 406 **relative distribution of treatment categories. Source: [40].**

407 As illustrated in Fig. 3, at right, the relative reduction of EU landfilling further corresponds  
 408 to a 34% decrease for this treatment option out of total MSW treatment volumes (67% in 1995  
 409 and 33% in 2012), which effectively confirms an average fulfilment of the Landfill Directive  
 410 targets for 2016. As for the relative change of MSW volumes incinerated with energy recovery,  
 411 equivalent shares out of total MSW treatment volumes were 7% (1995) and 20% (2012), while  
 412 total incineration increased from 15% to 24%. Validation crosschecking of Eurostat reported  
 413 MSW incineration volumes for 2010 with IEA data [53] identified smaller deviations. Possible  
 414 explanations might be varying MSW definitions and counting routines among MS and perhaps  
 415 incineration of MSW in e.g. cement kilns not accounted for in the Eurostat data. While  
 416 summing up total incineration volumes to 57 Mt (44 Mt with energy recovery and 13 Mt  
 417 without), the Eurostat data was somewhat lower than the anticipated 65 Mt (648 PJ and by  
 418 conversion factor of 10 GJ/t) reported by the IEA. Since the used Eurostat data includes no

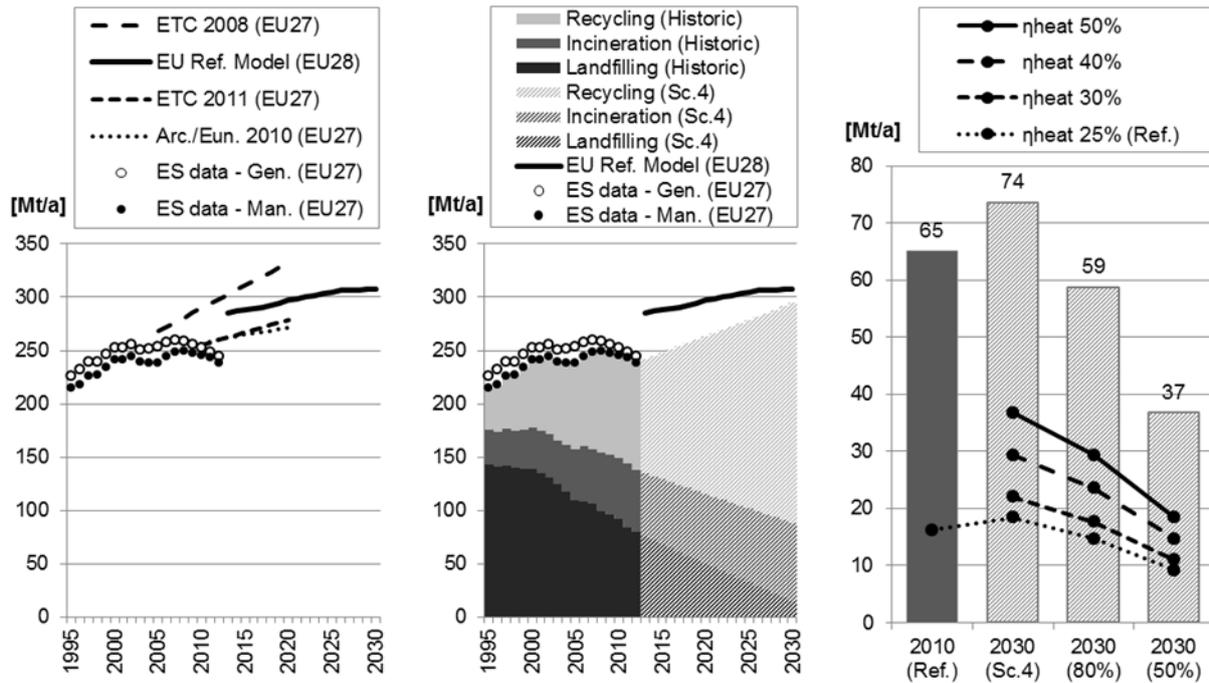
419 information on generated heat output, hence inhibiting an estimate of heat recovery  
420 efficiency, the IEA data was chosen for this purpose.

### 421 **3.2 Waste generation and treatment – in 2030**

422 The first of four considered model predictions assessing EU MSW generation was the ETC  
423 2008 projection. As shown in Fig. 4, at left, the trajectory proposed at this stage (projection  
424 start year 2005, EU27) anticipated total MSW generation volumes well above 330 Mt by 2020,  
425 a probable over-shoot three years later curtailed to approximately 280 Mt in the ETC 2011  
426 version (start year 2007). The 2010 assessment of Arcadis/Eunomia arrived at a similar level  
427 (270 Mt). As if outlining maximum and minimum conditions, the three 2020 assessments  
428 seemingly mark the confines of most plausible developments for 2030, as the EU Ref. model  
429 projection stretches on to a total MSW generation volume of 308 Mt by this year (294 Mt  
430 available for management, given a historic 95.6% average ratio between generation and  
431 management in EU27 during 1995 to 2012).

432

433



434  
 435 **Fig. 4.** At left, EU27 MSW generation and management volumes from 1995 to 2012 and four model predictions  
 436 of future EU MSW generation. At centre, EU27 MSW volumes by treatment categories from 1995 to 2012  
 437 (“Recycling” including composting) linearly interpolated from 2012 to 2030 according the EU Ref. model  
 438 (Scenario 4). At right, MSW volumes designated for incineration in 2010 and current heat recovery efficiency  
 439 (reference) compared by sensitivity analysis to the EU Ref. model in 2030 (Scenario 4: 100%, 80% and 50%)  
 440 and three levels of plausible future EU heat recovery efficiencies. Sources: [40, 53, 84-86, 88].

### 441 3.3 Heat recovery in European district heating systems in 2030

442 If in 2030, the EU Ref. model projections will have proven accurate, approximately 294 Mt  
 443 of MSW will be available for waste management in EU28. If the European community,  
 444 furthermore, manages to comply with the 4<sup>th</sup> scenario in the EU Ref. model context, 70% (206  
 445 Mt) of this manageable MSW volume will be recycled and only 5% (~15 Mt) will be landfilled.  
 446 The remaining 25%, equalling 74 Mt by this year, should consequently be available for energy  
 447 and heat recovery. Hereby, as presented in Fig. 4, at centre, data based 2012 levels of MSW  
 448 treatment categories may be linearly interpolated to meet these anticipated 2030 volumes,  
 449 hence permitting a visual representation and comprehension of the plausible distribution of  
 450 EU MSW management in the years to come.

451        Supposing that realised conditions in 2030 arrive at only 80% (235 Mt), or merely 50%  
452 (147 Mt) of the model projection, correspondingly less MSW will be available for energy  
453 conversions (59 Mt (80%) and 37 Mt (50%), assuming constant shares), as outlined in Fig. 4, at  
454 right. Depending, however, on what average EU heat recovery efficiencies that will have been  
455 attained by this year, which itself, in essence, will be determined by materialised levels of  
456 future EU district heating deployment, actual heat recovery volumes may increase although  
457 available MSW volumes are reduced. If, in 2030, only half of projected MSW volumes will be  
458 generated (thereby leaving only 37 Mt to energy conversions), then 184 PJ may still be  
459 recovered as heat – given an average EU heat recovery efficiency of 50% (corresponding to  
460 18.4 Mt at conversion factor 10 GJ/t).

#### 461    **4 Discussion**

462        From this, it is likely that future heat recoveries from waste in European district heating  
463 systems will be determined by (at least) two independent processes; (i) the success by which  
464 decoupling of human well-being from virgin resource use is materialised, and (ii) forthcoming  
465 transition levels towards serial supply structures for the provision of building heat demands.

466        As for the first of these detached developments, there are today only vague indications  
467 that absolute decoupling, i.e. reduced waste generation parallel with continued economic  
468 growth, is occurring in the EU context. Mazzanti and Zoboli, who, by non-linear regression  
469 modelling (considering as well structural and socio-economic variables), evaluated the EU  
470 decoupling progress in 2008, concluded that “for waste generation there is still no absolute  
471 delinking trend” [98], see also [99]. In a Swedish study [100], Sjöström and Östblom suggested  
472 that, for absolute decoupling by 2030 (start year 2006), waste intensities linked to the three

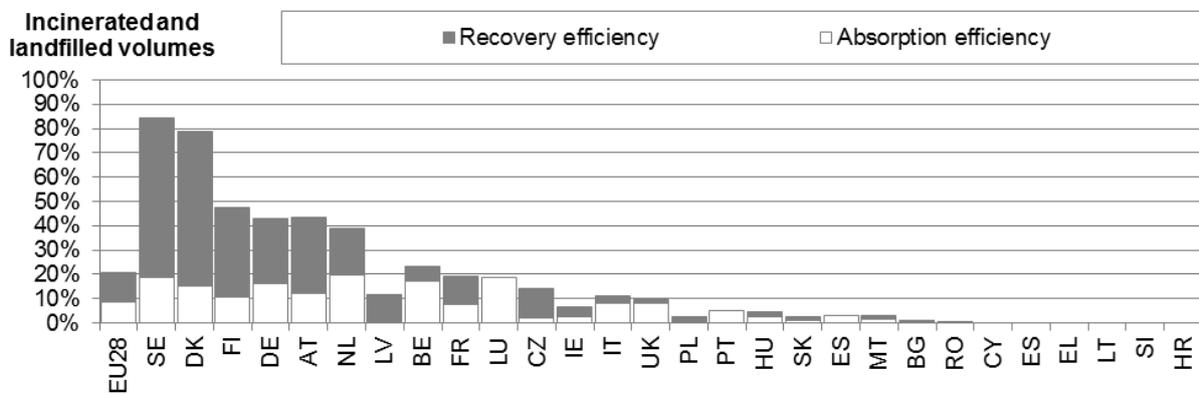
473 drivers technical change, economic growth, and household consumption, will need to  
474 decrease at a rate twice that of historical reduction rates, which appears not to be the case in  
475 contemporary Europe. Some indices, however, of both absolute and relative MS level  
476 decoupling of MSW generation relative economic growth, has been reported more recently in  
477 [101], in turn referring to [102-104]. Nevertheless, despite marginally more efficient domestic  
478 and raw material consumption during 2000 to 2012 (mainly in coincidence with the financial  
479 crisis of 2008), as reported in [105], European consumption patterns have in general remained  
480 highly resource intensive, which is why continued absolute decoupling must be acknowledged  
481 as posing both fundamental and structural challenges for the European community.

482 In this respect, let aside prevention and minimisation efforts (as, for a radical example,  
483 taxation of resources rather than taxation of labour, proposed by Bruvoll in 1998 [106]),  
484 absolute decoupling needs to imply either a shift towards more immaterial consumption, i.e.  
485 less resource intensive products and commodities, or a principally new regime in terms of  
486 European re-use and recycling. In view of current recycling levels (~27% recycling and ~15%  
487 composting, totalling at ~42% in 2012, see Fig. 3, at right), a proposition of absolute  
488 decoupling of waste generation from economic growth will in itself be contradictory unless  
489 supported by the permeate arrangement of effective, operational, and sustainable recycling  
490 technologies and infrastructures. It can therefore be assumed that a realisation of current EU  
491 decoupling ambitions will require levels of political and economic devotion, of state and  
492 municipal commitment, and of collective and individual discipline, intrinsically higher than  
493 current ones, and that this needs to be addressed and accompanied by appropriate policy and  
494 procurement measures.

495 As for the second process, of which much less can be said in terms of expected  
496 developmental progress since no formal targets so far has been set for EU district heating  
497 deployment, a transition towards serial supply structures on the European building heat  
498 market is as well associated with considerable economic investment. Moreover, a genuine  
499 transition of this kind, i.e. a reform towards improved structural energy efficiency, is likely to  
500 have a profound influence on traditional energy system perceptions, as well as on a wide array  
501 of technical, industrial, social, and infrastructural dimensions. In relation to WTE conversions,  
502 given their appropriate application, waste incineration with heat recovery in district heating  
503 systems is likely to represent, also in ambitious recycling and circular economy contexts, a  
504 necessary, multifunctional technology solution by which to bridge and enhance resource and  
505 energy efficiency improvements. The peculiar circumstance that (non-recyclable) combustible  
506 waste fractions, simultaneously being both a burden and an asset (residue and fuel), find their  
507 most rewarding application in such WTE conversions needs to be kept in mind when  
508 formulating future EU waste management and energy system policies.

509 In a recent European study by Sundberg [107], asking what role energy recovery will have  
510 in the context of increased material recycling, main conclusions are that continued access to  
511 WTE facilities are essential also at high recycling rates; namely by the ability to treat  
512 deteriorated combustible residues inevitably extending from recycling processes.  
513 Additionally, let aside generation of electricity and heat, incineration (and gasification)  
514 processes make viable the extraction of metals contained in the original waste flow, as well as  
515 the destruction of contaminated and non-recyclable fractions – all eventually contributing to  
516 a reduced demand for landfill deposits. Another key message from Sundberg, also illustrated  
517 in Fig. 2, is that MS with lowest landfilling rates at current; all have highest levels of both

518 recycling and incineration (e.g. AT, BE, DE, DK, NL, and SE). It is clear, however, that waste  
 519 incineration processes without heat recovery, not to mention without energy recovery all  
 520 together, imply significantly reduced synergetic qualities of the respective treatment  
 521 operations performed. To illustrate this further, see Fig. 5 (similar in arrangement to the 2008  
 522 projection presented by Persson and Werner in [108]), average 2012 EU28 MS heat recovery  
 523 and absorption (electricity) efficiencies from all non-recycled MSW fractions (i.e. both  
 524 incineration and landfill volumes) are depicted together with EU28 average values.

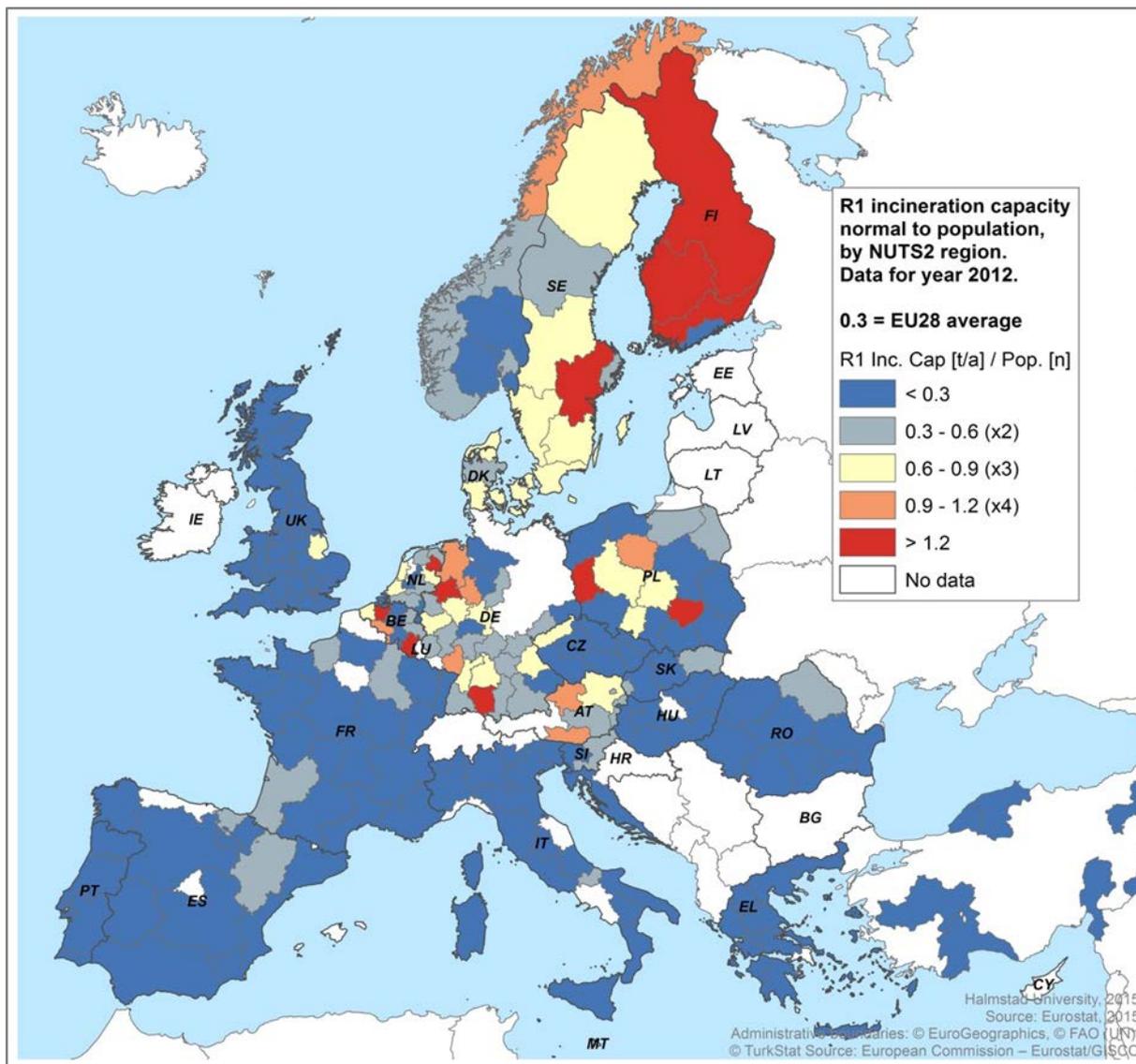


525  
 526 **Fig. 5. Distribution of recovery and absorption efficiencies for incinerated and landfilled (non-recycled)**  
 527 **volumes of MSW in EU28 Member States during 2012. Sources: [40, 53]. Absorption efficiency is equal to**  
 528 **electrical efficiency.**

529 The Eurostat data [40] used in Fig. 5 corresponded to 58.1 Mt (incineration) and 80.7 Mt  
 530 (landfilling) in 2012, hence totalling at 139 Mt in this year. At the used study conversion factor  
 531 of 10 GJ/t, this would then correspond to energy magnitudes of approximately 0.58 EJ and  
 532 0.81 EJ, respectively, revealing a total 1.39 EJ of energy embedded in these waste fractions.  
 533 After cross-referencing with IEA data for the same year [53], the total energy content was  
 534 eventually corrected to 1.51 EJ. By reference to this total volume, the 0.13 EJ of generated  
 535 electricity rendered an absorption efficiency of 9% and the total heat output of 0.19 EJ (sum  
 536 of reported heat output from energy sector activities (0.17 EJ) and additional MSW heat  
 537 output in total final consumption outside the industry sector (0.02 EJ)) rendered an average

538 EU28 heat recovery efficiency of 12%. From this, an average overall conversion efficiency of  
539 only 21% from European MSW fractions currently not recycled or composted indicates that  
540 79% (1.19 EJ) of the total energy content remained unharvested during this year. Thus, it is  
541 fair to conclude that more energy can be recovered from current EU MSW not recycled, mainly  
542 by redirection of current landfill volumes, and that installation of more incineration capacity,  
543 as well as retrofitting of existing units, seems generally viable.

544 Another decisive aspect to consider in this respect, perhaps also representing a third  
545 independent process to monitor, concerns the geographical distribution of current  
546 incineration capacities, which by use of pilot project Eurostat data on NUTS2 level is visualised  
547 in Fig. 6 [56, 58]. Concerning installed R1 capacity (which notably is summed up to a staggering  
548 149.6 Mt in this dataset, not commented further here) this map illustrates the current  
549 distribution of recovery capacity in Europe by normalisation to regional population counts and  
550 referring to an anticipated EU28 average ratio (0.3 by relating to the abovementioned total  
551 capacity). Since, in practice, a more even distribution of R1 capacities is achievable mainly by  
552 increased access to local heat distribution infrastructures, this image indeed underlines the  
553 need to expand European district heating if to obtain spatially coherent heat recoveries from  
554 present and future WTE conversions.



555  
 556 **Fig. 6. R1 incineration capacity per NUTS2 region, normal to population. Average EU28 value established as**  
 557 **ratio of reported total R1 capacity (149.6 Mt) and total EU28 population in 2012. Sources: [56, 58].**

558        Additionally, contextual developmental trends such as demographically expressed in  
 559 migrations and urbanisation, do in themselves influence future European waste management  
 560 conditions. According to used study data, including time series of regional population counts  
 561 from 1990 to 2014 [56], clear indications of population movements mainly from former east-  
 562 European MS westward, and as well continued urbanisation processes in practically all large  
 563 urban zones, should constitute useful information in the construction of future European  
 564 waste management structures. It is noteworthy, since urban areas represents most beneficial

565 conditions for cost-effective heat distribution [31], while also hosting a majority of energy and  
566 industry sector excess heat [36], that Mazzanti and Bozoli [98] found evidence of increased  
567 costs for recycling infrastructures (sorting, separation) in city areas compared to rural areas,  
568 hereby imposing relatively stronger economic constraints on urban recycling. Finally,  
569 illustrative of the many approaches and preferences plausible, current waste imports and  
570 exports may reversely be viewed as a potent means by which to achieve a faster reduction of  
571 current landfill volumes, as discussed amongst others in [109], which may serve here as a final  
572 example of the general complexity characterising European waste management.

## 573 **5 Conclusions**

574 To conclude, in this paper a literature and data review has been presented by which to  
575 answer three initial research questions concerning current and future prospects for heat  
576 recovery from waste in European district heating systems. The study has focused on MSW in  
577 the EU context with the purpose to identify aggregated general tendencies and trends by  
578 which to assess plausible future developments for EU waste generation and management. The  
579 main and overarching conclusion is that efficient, i.e. recovery classed WTE conversions (as  
580 defined by the R1 formula), in principal requires access to heat distribution infrastructures by  
581 which to utilise recovered excess heat. In this respect, current and future deployment levels  
582 of district heating systems throughout the European continent are and will be directly  
583 reflected in the geographical distribution and spatial spread of efficient waste incineration  
584 capacity. For obvious reasons, therefore, EU waste management policies should align and  
585 interact to a considerable degree with corresponding energy system related regulations and  
586 concerns.

587 As for the first dedicated research question, historical trends of MSW generation and  
588 management among EU MS from 1995 to 2012 reveal, on average, clear evidence of markedly  
589 reduced annual volumes designated for landfilling, effectively reflected in increasing  
590 corresponding shares for recovery incineration, recycling, and composting, respectively. On  
591 MS level, however, landfilling is still excessive in some instances and one third of all generated  
592 MSW in Europe is still deposited to land. Apart from representing a valuable source of energy  
593 not exploited, this practise is associated with greenhouse gas emissions and stress on local  
594 ecosystems. It is further observable, that MS which have successfully implemented waste  
595 management routines including all non-landfilling option, i.e. recycling, composting, and  
596 recovery incineration, also have lowest landfilling volumes. Additionally, WTE incineration  
597 with heat recovery has increased three-fold during the considered period in terms of energy  
598 volumes, but recovery capacity is poorly distributed over the European continent.

599 For the second research question, the study answer is that several different models and  
600 approaches have been developed and used in recent years to assess future EU MSW  
601 generation and, to a lesser extent, to estimate future deployment levels of European district  
602 heating. Originating principally in classical economics, be they characterised by material  
603 balances (input/output), multi-sectoral equilibrium models, or statistical econometric  
604 approaches, four up-to-date MSW model predictions targeting the EU were selected during  
605 the review process. While data leading up to 2012 indicate continuously reduced total MSW  
606 generation volumes in Europe, as to which the influence of the financial crisis in 2008 – or  
607 perhaps a genuine behavioural shift in consumption patterns – is too early to determine  
608 statistically, all considered models predict increased MSW generation volumes for the years  
609 to come. The most recent of these, the EU Ref. model, being the only one extending to 2030,

610 foresees a total MSW generation volume of 308 Mt by this year, which, by a historical average  
611 ratio between generation and treatment volumes, should correspond to some 294 Mt  
612 available for waste management options. Through the availability of regional waste data in  
613 later years, future modelling approaches should benefit from applications such as  
614 geographically weighted regression analysis and other contemporary Geographical  
615 Information Systems (GIS) applications.

616 Thirdly, based upon the most ambitious and stringent of five scenarios for 2030  
617 elaborated in the context of the EU Ref. model, stipulating 70% recycling and a 5% landfill limit  
618 by this year, the study results suggest that the remaining one-quarter could be conceived as  
619 available for energy recovery. The key conclusion in this respect, however, is not what exact  
620 magnitudes of MSW that will be available for thermal energy conversion in a given year, but  
621 rather at what heat recovery efficiencies that available waste volumes will be harnessed in  
622 forthcoming WTE conversions. Depending thus, in a deeper sense, on the transitional progress  
623 towards more efficient supply structures on future European heat markets (key features in  
624 smart energy systems), the presence of local heat distribution infrastructures, as conceived  
625 here, is an equally determinant factor as that of total generation volumes regarding prospects  
626 for future heat recoveries from waste.

627 Hereby, district heating systems represent an important infrastructural technology,  
628 essential for not only providing energy and environmentally efficient heat supplies to  
629 residential, service, and industry sectors, but also for facilitating increased total conversion  
630 efficiencies of WTE plants. In such contexts, energy recovery from waste, itself then a driver  
631 for future district heating deployment in Europe, represents an enabling technology whereby  
632 to achieve improved energy system efficiency while simultaneously reducing the demand for

633 landfilling and deposits, again reminding us of the synergetic opportunities inherent to WTE  
634 system technologies.

## 635 **6 Acknowledgements**

636 The work presented in this paper is a result of the research activities of the Strategic  
637 Research Centre for 4th Generation District Heating (4DH), which has received funding from  
638 The Innovation Fund Denmark. The study was performed under WP2 and initiated as item 19  
639 in the 2014/2015 work programme. The authors also wish to extend their grateful thanks to  
640 professor Frits Møller Andersen at the Technical University of Denmark for two interviews  
641 during the spring of 2015.

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