

# Using artificial-reef knowledge to enhance the ecological function of offshore wind turbine foundations: implications for fish abundance and diversity

Glarou, Maria; Zrust, Martina; Svendsen, Jon C.

Published in: Journal of Marine Science and Engineering

Link to article, DOI: 10.3390/jmse8050332

Publication date: 2020

Document Version Peer reviewed version

Link back to DTU Orbit

Citation (APA):

Glarou, M., Zrust, M., & Svendsen, J. C. (2020). Using artificial-reef knowledge to enhance the ecological function of offshore wind turbine foundations: implications for fish abundance and diversity. *Journal of Marine Science and Engineering*, *8*(5), Article 332. https://doi.org/10.3390/jmse8050332

#### **General rights**

Copyright and moral rights for the publications made accessible in the public portal are retained by the authors and/or other copyright owners and it is a condition of accessing publications that users recognise and abide by the legal requirements associated with these rights.

• Users may download and print one copy of any publication from the public portal for the purpose of private study or research.

- You may not further distribute the material or use it for any profit-making activity or commercial gain
- You may freely distribute the URL identifying the publication in the public portal

If you believe that this document breaches copyright please contact us providing details, and we will remove access to the work immediately and investigate your claim.





1 Review

#### Using artificial-reef knowledge to enhance the 2

- ecological function of offshore wind turbine 3
- foundations: implications for fish abundance and 4
- diversity. 5

#### 6 Maria Glarou 1,2,\*, Martina Zrust 1 and Jon C. Svendsen 1

- 7 <sup>1</sup> DTU Aqua, Technical University of Denmark (DTU), Kemitorvet, Building 202, 2800 Kongens Lyngby,
- 8 Denmark; martinazrust@icloud.com (M.Z.); jos@aqua.dtu.dk (J.C.S.)
- 9 Department of Ecology, Environment and Plant Sciences, Stockholm University, Svante Arrhenius väg 20 10 A (or F), 114 18 Stockholm, Sweden
- 11 Correspondence: maria\_glarou@outlook.com; Tel.: +45 50174014
- 12 Received: date; Accepted: date; Published: date

13 Abstract: As the development of large-scale offshore wind farms (OWFs) amplifies due to 14 technological progress and a growing demand for renewable energy, associated footprints on the 15 seabed are becoming increasingly common within soft-bottom environments. A large part of the 16 footprint is the scour protection, often consisting of rocks that are positioned on the seabed to 17 prevent erosion. As such, scour protection may resemble a marine rocky reef and could have 18 important ecosystem functions. While acknowledging that OWFs disrupt the marine environment, 19 the aim of this systematic review was to examine the effects of scour protection on fish assemblages, 20 relate them to the effects of designated artificial reefs (ARs) and, ultimately, reveal how future scour 21 protection may be tailored to support abundance and diversity of marine species. The results 22 revealed frequent increases in abundances of species associated with hard substrata after the 23 establishment of artificial structures (i.e. both OWFs and ARs) in the marine environment. Literature 24 indicated that scour protection meets the requirements to function as an AR, often providing shelter, 25 nursery, reproduction and/or feeding opportunities. Using knowledge from AR models, this review suggests methodology for ecological improvements of future scour protections, aiming towards a 26 27 more successful integration into the marine environment.

- 28 Keywords: scour protection; offshore wind farm; renewable energy; artificial reef; ecosystem 29 restoration; ecological engineering
- 30

#### 31 1. Introduction

32 Humans rely upon and utilize the marine environment for a variety of ecosystem services and 33 valuable resources, leaving no parts of the oceans unaffected [1,2]. The majority of people live in the 34 coastal zones, where anthropogenic activities have been progressively altering the seascape [3–5], 35 resulting in direct and indirect negative impacts on marine biota [6]. The introduction and 36 proliferation of artificial (i.e. engineered) structures to marine ecosystems, driven by commercial and 37 residential demands, causes a phenomenon known as "ocean sprawl" [7]. "Ocean sprawl" refers to 38 a marine environment that is increasingly becoming dominated by artificial structures, including 39 artificial reefs, breakwaters, seawalls, piers, oil platforms and marine renewable energy installations 40 [7], often causing various effects on the surrounding ecosystems [8].

41 An artificial reef (hereafter abbreviated as "AR") may be defined as a structure of "natural or 42 human origin deployed purposefully on the seafloor to influence physical, biological, or socio-43 economic processes related to living marine resources" [9]. A designated AR can be employed for a 44

45 also to ecosystem restoration, as well as socioeconomic development [10,11]. Specific AR purposes 46 may include aquaculture/sea-ranching, biomass increase, biodiversity enrichment, fisheries 47 production, ecosystem management, prevention of coastal erosion, recreational activities (e.g. scuba 48 diving, ecotourism, fishing), and research [10,12,13]. These engineered structures are typically 49 constructed to resemble natural reefs as much as possible, with the ultimate goal to produce similar 50 effects. There are also other structures acting as ARs that (a) have entered the marine environment 51 accidentally (e.g. shipwrecks, lost containers), (b) have been repurposed (e.g. sunken ships for 52 recreational activities) or (c) serve other functions (e.g. offshore oil and gas platforms) [14–17].

53 Another example of man-made infrastructure are offshore wind farms (hereafter abbreviated as 54 "OWFs"). OWFs are increasingly established in marine areas with the purpose to meet the rising 55 global demand for renewable energy [18,19]. Apart from the obvious benefits provided by renewable 56 energy, OWFs may have several ecological effects on the marine environment. Ecological disruptions 57 caused by OWF include avian collisions [20,21], underwater noise [22–24] and electromagnetic fields 58 [18,25,26], as well as loss of soft-bottom habitat with the introduction of hard substrata [27]. Other 59 changes caused by OWFs include increased abundance and biodiversity of hard-bottom species due 60 to reef effects and creation of no-take zones within the OWF, with possible spill-over effects to 61 neighbouring areas [28-30].

62 OWFs typically consist of an array of individual wind turbines placed on soft-bottom habitats. 63 Since offshore wind turbines are located on soft sediment, scouring inevitably occurs around their 64 foundations. Scour is created when a steady current (e.g. tide, wave activity) encounters a vertical 65 structure on the seabed, causing local increases in flow speeds and turbulence levels and ultimately 66 leading to the creation of a scour pit around the structure [31]. Scouring may compromise the stability 67 and dynamic behaviour of wind turbine foundations [32]. The magnitude of scouring is affected by 68 the current speeds, the water depth and sediment types [31].

69 Scour protection is a measure used to prevent the erosion of seabed sediment around individual 70 foundations of offshore wind turbines [33]. Typically, scour protection consists of a filter layer made 71 of gravel, shielded by a rock armour layer [34] (Figure 1). The material is placed around the 72 foundation of the turbine (e.g. around the monopile) with a radius typically reaching up to 20 meters 73 [34,35]. The size and design of the scour protection is determined by a range of environmental factors 74 (i.e. wave and current activity, water depth, sediment characteristics), as well as structural factors 75 (i.e. monopile diameter and design) [36,37] (Figure 1). Generally, the size of the scour protection area 76 reflects the area of the scour pit that would arise due to the environmental factors (e.g. current speed), 77 if the foundation was left unprotected [31], which is usually four to five times the monopile diameter 78 [38].



80 Figure 1. Illustration of a common scour protection design, modified from Whitehouse et al. [38]: 81 Scour protection around an offshore wind turbine, laid on the seafloor. The vertical bar represents the 82 monopile that holds the windmill rotor and turbine. The scour protection includes a preinstalled filter 83 layer of small sized rocks (e.g. 5 cm in diameter) covered by an armour layer of larger rocks (e.g. 40 84 cm in diameter). The filter layer prevents erosion of sand through the upper layer of larger rocks. The 85 filter layer is usually about 0.5 m high, whereas the armour layer is about 1 m high. Edge scour (e.g. 86 0.5 m deep) may develop in the periphery of the scour protection. Further details on scour protection 87 designs are provided by Whitehouse et al. [38]. In the present review, the foundation is defined as the 88 monopile and the surrounding scour protection as illustrated in the figure.

89 It is increasingly recognized that the changes introduced during the Anthropocene, including 90 ocean sprawl and coastal hardening, have transformative effects on natural ecosystems [7,39,40]. 91 Hence, the establishment of a mutually beneficial relationship between biota and man-made 92 infrastructure is crucial. The concept of "renewal ecology", introduced by Bowman et al. [41], falls 93 directly under this scope. Renewal ecology is defined as "a solutions-focused discipline aimed at 94 creating and managing ecosystems designed to maximize both biodiversity and human well-being 95 in the face of rapid environmental change" [41]. Although renewal ecology remains to be embraced 96 widely, especially in the marine environment, the concept is progressively gaining support. 97 Importantly, renewal ecology resembles IUCN's principles of "Nature-based solutions", recently 98 embraced by the commission of the European Union [42]. For example, coastal defence structures are 99 increasingly constructed not only to meet engineering requirements, but also to enhance biodiversity 100 and provide ecosystem services [11]. Moreover, in terms of OWFs, The Netherlands recently 101 introduced new permit obligations, requiring engineers to "make demonstrable efforts to design and 102 build the wind farm in such a way that it actively enhances the sea's ecosystem, helping to foster 103 conservation efforts and goals relating to sustainable use of species and habitats" [43]. These new 104 obligations recognize the increasing presence of anthropogenic infrastructures in the marine 105 environment and therefore seek to design OWFs in a fashion where positive effects exceed negative 106 effects for the marine biota.

# 107 *1.1 Objectives*

108 The concept of eco-engineering man-made infrastructure is increasingly discussed in the 109 existing literature [44–46], however, to date, a review investigating how scour protection in OWFs

110 may be tailored to create productive habitats for various fish species has not been carried out. The

111 present review aims to fill this void by synthesizing current evidence on changes in the abundance 112 and diversity of fish and commercially exploited species in response to the establishment of OWFs in 113 temperate seas and relating this knowledge to existing AR research. The overarching goal is to extract 114 current knowledge of AR designs and use it to suggest scour protection refinements aimed at 115 supporting fish abundance and diversity. Ultimately, improved fish production attributed to scour 116 protection may support commercially exploited fish stocks and contribute to sustainable fisheries. 117 The review concludes by identifying knowledge gaps and suggesting future research directions. Only 118 ARs designed with a purpose (e.g. increasing species abundance and diversity) and designed 119 similarly to scour protection (e.g. gravel, rocks, stones, granite, boulders etc.) were examined in this 120 review, while accidental ARs (e.g. shipwrecks) [12] were disregarded. This allowed for comparisons 121 between ARs and scour protections with the purpose of revealing various AR design refinements 122 that could be transferable to future scour protection construction. Additionally, studies occurring in 123 non-temperate regions were excluded, because the majority of OWFs occur in temperate marine 124 waters.

# 125 2. Materials and Methods

126 The present study was conducted following the systematic literature review protocol developed 127 by Pullin and Stewart [47]. The guidelines have been used by numerous systematic reviews related 128 to offshore energy, including reviews concerning the potential of OWFs to act as marine protected 129 areas [30] and the impacts of energy systems on marine ecosystem services [48].

130 2.1 Search Terms

A wide range of search terms were selected to incorporate all the components of the study objectives. The specific search terms were formulated to gather all relevant literature from the databases. Similar to previous systematic reviews [49], negative terms were included to increase the efficacy of the string and to remove impertinent search results. Applied search terms are listed in Table 1, where terms with an asterisk (\*) represent a search engine wild-card [49]. These search terms were applied to two different databases: "Web of Science" and "Scopus".

137 **Table 1.** Selected search terms used in "Web of Science" and "Scopus".

*Technical terms:* artificial reef \*, reef effect\*, fish attraction device\*, artificial structure\*, scour\*, scour\* protection\*, scour control\*, wind power foundation\*, offshore wind, wind farm foundation\*, turbine foundation\*, offshore energy, offshore wind energy structure\*, artificial offshore construction\*, offshore wind farm\*, wind turbine\*, rock armour\* *Ecosystem-related terms:* ecological effect\*, impact\*, sanctuary\*, spillover, habitat\*, habitat change\*, habitat restor\*, habitat creat\*, species abundance\*, biomass\*, biodiversity, species composition\*, species densit\*, nurser\*, recruitment, coloni\*, migrat\*, food availab\*, ecological function\*, aggregat\*, productiv\*, enhanced habitat\*, heterogeneity, feeding, spawn\* OR habitat creat\*, habitat connectivity, habitat complexity, habitat enhancement\*, no trawling zone\*, hard bottom, invasive species, alien species, nonindigenous, habitat fragment\*, habitat degrad\*, habitat loss\*, hard substrate\*, substrate\* *Target species group term:* fish\*

Negative terms: tropical, subtropical, Caribbean, Indian Ocean

138 2.2 Screening Process

139 The papers resulting from the searches were assessed for relevance at three sequential levels:

- 140 title, abstract, and full text [47]. Papers included in the analyses fulfilled the inclusion criteria (Table
- 141 2).

Criteria	Include	Exclude	
Peer-reviewing	Peer-reviewed studies	Reviews & everything else	
Text Language	English	Everything else	
Years	All years (1900-2018)	-	
Location	Temperate seas	Everything else	
Intervention	OWFs and designated ARs	Accidental ARs, such as shipwrecks	
Subject	Fish and fisheries	Everything else	

## 142 **Table 2.** Inclusion and exclusion criteria for the systematic review.

#### 143 2.3 Data Extraction

Study characteristics and the findings of the resulting papers were recorded (Table 2). Specifically, documented details included: (a) the type of structure (e.g. AR, OWF and the associated scour protection), (b) the spatial scale of the study (e.g. area of AR or scour protection), (c) the temporal scale of the study (e.g. time elapsed after the deployment of structure), (d) the physical characteristics of the structure (e.g. material, shape) and (e) the examined variables (e.g. fish abundance and diversity).

# 150 2.4 Data Analysis

A quantitative analysis was conducted based on the data extracted by parsing the literature over 152 1900 – 2018. Relevant literature was limited within a 37-year period (1982 – 2018), and for each year, 153 the total number of AR and OWF papers was recorded. Similar to previous reviews [49], the 154 accumulated number of papers for both subjects was plotted across time and general linear regression 155 was applied to identify research trends. In addition, further analyses compared various parameters 156 related to scour protection and ARs (e.g. material and water depth). Statistical analyses were carried 157 out using the free software R Studio (R Core Team, 2019).

# 158 **3. Results**

## 159 3.1 Selection Process

160 The selection process resulted in 7,027 papers, with a total of 6,537 remaining after duplicate 161 removal. Following the removal of irrelevant subjects, locations, and reviewing the material through

162 title, abstract, or full-text analysis, a total of 115 peer-reviewed papers met the inclusion criteria (Table

163 2; Figure 2). Of those, 89 and 26 papers pertained to ARs and OWFs, respectively.



Figure 2. A "PRISMA" diagram showing the flow of information through the different phases of thereview [50].

167 *3.2 Temporal research trends* 

168 Published AR research started in the early 1980s, whereas OWF research started around the mid

169 2000s (Figure 3). Despite this time difference, both topics reveal increasing trends (ARs: R<sup>2</sup> = 0.34, p <

170 0.001; OWFs:  $R^2 = 0.39$ , p < 0.001). Earlier peer-reviewed literature could possibly be non-accessible

171 due to historical limitations of certain search databases (e.g. Web of Science), creating potential bias

172 in regard to the commencement of research on both topics.



174Figure 3. Trends of research output (articles per year) for artificial reef (AR) and offshore wind farm175(OWF) studies through the years 1982 – 2018 (source: Web of Science, Scopus). For clarity, data are176presented in six years intervals, with the exception of the 2012-2018 period, which includes seven177years. This unequal division was necessary to address the total period of 37 years that contained all178articles fulfilling the inclusion criteria. Regardless of the duration of the time interval (six or seven179years), data are presented as the mean number of articles per year from each time interval.

180 3.3 Location of available studies

Many AR studies originated from Southern Europe, Southwest USA and East Asia (Figure 4). In
contrast, OWF studies mainly originated from northern Europe, particularly within the North Sea
(including Belgium, Denmark, Germany, the Netherlands and the UK) (Figure 4). This indicates
geographical differences between AR research and OWF research. Only Germany and the USA have
covered both topics (Figure 4).



187Figure 4. Number of artificial reef (AR) and offshore wind farm (OWF) studies per country188throughout the years 1982 – 2018 (source: Web of Science, Scopus).

#### 189 3.4 Inconsistent study details

190 The amount of study details varied between papers (Table 3). All studies reported the sampling 191 method, whereas fewer studies included information about the spatial scale and the material used 192 for the structure (Table 3).

193**Table 3.** A list of 17 variables reported in the articles that met the inclusion criteria (Table 2), similar194to Whitmarsh et al. [51]. The table provides examples of the variables that were reported in each195article, as well as the percentage of articles that mentioned each specific variable. For example,196regarding the spatial scale of an AR or OWF (variable 8 from the top), an example of 700m<sup>2</sup> is197provided. This means that at least one of the reviewed articles examined an AR or OWF area covering198700m<sup>2</sup>. The rest of the included examples follow the same logic and are not general findings.

Variable	Examples	% of AR studies reported in (% out of 89)	% of OWF studies reported in (% out of 26)
General			
Type of Study	Scientific Paper, Report	100	100
Year published	2000, 2018	100	100
Location	Italy, North Sea, UK	100	100
Study design			
Sampling method	Visual census, Fishing	100	100
Study design	BACI*, Impact only	92	96

Temporal Scale	10 years after deployment	79	69
Sampling Season	Summer, Winter	82	92
Structure			
Spatial Scale	700m <sup>2</sup>	44	54
Depth	30m	76	61
Material	Concrete, rock	84	61
Shape of material	Cube, boulder	79	0
Volume of material	1m <sup>3</sup>	52	4
Effects on fish			
Overall impact	+, +/-, -	94	88
Biomass/Abundance	+, +/-, -	72	73
Diversity	+, +/-, -	48	38
Shelter (Complexity)	Yes, No, Possibly	34	85
Spawning/Settlement	Yes, No, Possibly	22	31

199 + Increase

200 - Decrease

201 +/- No noticeable effect

202 \* Before-after-control-impact experimental design

# 203 3.5 Overlapping use of materials for ARs and OWFs

204 Variable materials have been used for ARs and OWFs (Figure 5). In some occasions, ARs were 205 created using various types of rocks [52,53], but the majority of ARs were made using designed 206 concrete units [12]. Other AR components included metal and scrap materials [54,55] (Figure 5). Some 207 AR studies compared different materials, sizes and shapes used for the construction of ARs and 208 included more eco-friendly materials [56,57]. In contrast, scour protection structures commonly 209 consisted of various rock types (stones, pebbles etc.), with 61% of the literature reporting on at least 210 one of these materials (Figure 5). For the rest of the OWF studies (39%), the material used for scour 211 protection was not described. Thus, evidence suggested that scour protection is mainly made of 212 rocks.



214Figure 5. Materials used for artificial reefs (ARs) and scour protection associated with offshore wind215farms (OWFs). Some papers are included in multiple categories, because they studied more than one216material. "Other" includes scrap materials, wood, tires and PVC materials. "N/A" means that the217material used was not specified in the articles.

# 218 3.6 Overlapping water depths used for ARs and OWFs

219 Water depths of many AR and OWF studies are overlapping (Figure 6). The majority of AR and

OWF studies concern structures that are located at water depths between 15 and 30m. Studies related

to OWFs are, however, absent at water depths exceeding 30 m. This is largely due to OWF foundation

typology, as monopiles (Figure 1) are rarely constructed at water depths exceeding 30m [58].

![](_page_11_Figure_2.jpeg)

Figure 6. Mean water depths of artificial reef (AR) and offshore wind farm (OWF) structures. Some papers are included in multiple categories, because they studied more than one structure, located at different water depths. "N/A" means that the water depth was not specified in the articles.

# 227 3.7 Effects of ARs and OWFs on fish abundance

228 Provision of food [59,60], spawning [61] and shelter opportunities [62] by artificial structures are 229 among the main themes dominating the literature, causing changes in fish abundance, biodiversity 230 and distribution. In general, the vast majority of the AR studies (94%) reported that fish abundance 231 and biodiversity increased or was unaffected by AR deployment. More specifically, 49% of the 232 literature reported locally increased fish abundances after AR deployment, while 31% found an 233 increase in species richness in the AR area. Remaining papers did not report on the matter or recorded 234 non-significant differences. Similarly, about half of the OWF literature (46%) also reported increases 235 in fish abundance. Specifically, fish abundance was consistently higher near the OWF foundations 236 compared to reference areas [63–67]. This was particularly the case for species associated with rocky 237 substratum [68], but occasionally it also extended to soft-bottom species when sampling the sandy 238 areas within an OWF [69]. There was evidence of locally increased fish abundance for at least eight 239 fish species in the OWF literature: Atlantic cod (Gadus morhua), European eel (Anguilla anguilla), 240 goldsinny wrasse (Ctenolabrus rupestris), pouting (Trisopterus lucas), rock gunnel (Pholis gunnellus), 241 shorthorn sculpin (Myoxocephalus scorpius), sole (Solea solea), striped red mullet (Mullus surmuletus), 242 and whiting (Merlangius merlangus) [35,65,68–71]. Reubens et al. [70] found an average density of 14 243 pouting individuals per m<sup>2</sup> on the scour protection, yielding an estimated local population of 22,000 244 pouting individuals around one wind turbine foundation. It is important to note, however, that the 245 results reflect case studies and not necessarily general findings.

Only 38% of OWF papers reported on changes in fish diversity, however 40% of those (i.e. 15% of total) indicated that scour protection elevated fish diversity compared to neighbouring control areas [72,73]. Only one OWF study reported lower species diversity. Specifically, Wilhelmsson et al. [64] found lower diversity of demersal fish around the turbine foundations compared to the seabed 1 to 20 m away. Conversely, a few studies indicated that a range of soft-bottom species were unaffected by the scour protection [65,74]. For example, Langhamer et al. [75] found that the abundance of viviparous eelpout (*Zoarces viviparous*) was unaffected by an OWF in Sweden. Similarly, the abundances of adult individuals of several pelagic species, including horse mackerel (*Trachurus trachurus*), mackerel (*Scomber scombrus*), herring (*Clupea harengus*) and sprat (*Sprattus*) sprattus), were unaffected by the scour protection at a Dutch OWF [69].

256 A limited number of AR studies examined possible effects on soft-bottom species (e.g. flatfishes). 257 For example, Fabi and Fiorentini [76] reported lower catch rates of the soft-bottom fish species 258 (e.g. Trigla lucerna and Soleidae spp.), compared to a control site, similar to Bombace et al. [77]. In 259 relation to OWF, 19% of the studies included data on the impact on soft-bottom species. At Block 260 Island Wind Farm, off Rhode Island in US, a study of seven flatfish species revealed no negative 261 impacts [78]. On the contrary, Lindeboom et al. [65] reported a significant decrease in the soft-bottom 262 species lesser weever (Echiichthys vipera) two years after construction of a Dutch OWF. Krone et al. 263 [68] also reported low abundances of species associated with soft-bottom habitat, including species 264 of Gobiidae and *Callionymus* spp., near the scour protection of a German OWF. Similar OWF related 265 trends have also been observed for flatfish species, including dab (Limanda limanda), sole, and 266 solenette (Buglossidium luteum) [69].

## 267 4. Discussion

268 Although few countries have investigated both AR and OWF, this review identified overlap 269 between AR studies and OWF studies in terms of water depth and materials, indicating that 270 knowledge from temperate AR studies may be used to guide designs of future OWFs. Collectively, 271 AR and OWF studies often observed an increase in abundance and diversity of fish species associated 272 with hard-bottom habitats. Hard-bottom habitats provide food, shelter and habitats for reproduction, 273 supporting the hypothesis that scour protection may create a reef effect [65]. Whether the increase in 274 abundance is due to attraction or production is yet to be determined, however there is growing 275 evidence of new production associated with OWFs [64,72,79,80]. By modifying the designs of future 276 scour protection, a range of habitats may be supported, increasing the abundance of target species 277 [81] and possibly influencing fisheries via various mechanisms, including spill-over effects [28–30].

## 278 4.1 Fish abundance, biodiversity and distribution associated with ARs and OWFs

279 Artificial structures (i.e. both ARs and OWFs) typically hold higher fish density and biomass 280 [82–84], largely attributed to the form, complexity, area coverage [62] and/or food abundance [85]. 281 The findings of this review collectively confirm that artificial structures often increase the abundance 282 of hard-bottom species as well as fish diversity in the local area. In some instances, however, artificial 283 structures have no apparent reef effect for certain fish species, meaning that there is no noticeable 284 increase or decrease for a given species. For example, abundances of adult individuals of several 285 pelagic species, including horse mackerel, mackerel, herring and sprat were unaffected by the 286 presence of scour protection [69]. While adult individuals of pelagic species seemed unaffected by 287 scour protection, they may still utilize scour protection as spawning or rearing habitat. Hard-bottom 288 substrates are often important for the spawning of herring [86–89], suggesting that scour protection 289 could have similar function. The loss of soft-bottom substrate arising from OWF installations may 290 occasionally decrease the abundance of soft-bottom fish species [65,68,69]; however, due to the small 291 amount of area covered by scour protection within an OWF (approximately 0.8%), possible negative 292 impacts are considered insignificant at the population level as indicated by previous studies [69,73]. 293 Even though scour protection may have negative effects on soft-bottom species at the local scale, the 294 effects should be evaluated at larger spatial scales and related to fish population sizes and 295 movements.

#### 296 4.2 Attraction versus production

There are mainly two hypotheses aiming to explain the increased fish abundance in artificial structures [90]. Firstly, the attraction hypothesis presumes that fish simply aggregate at the installations from the surrounding environment, without a net increase in the population of the larger area. Secondly, the production hypothesis suggests that the carrying capacity of the ecosystem increases because of the artificial structures. According to the second hypothesis, fish growth, reproduction and/or survival is elevated in the area, resulting in population enhancement, eventually contributing to an increase in the net production, both for biomass and abundance [90,91].

304 While there is growing evidence of increased fish abundance associated with OWFs, it remains 305 unclear whether fish are simply attracted from surrounding areas or if the OWF facilitates fish 306 production. Recent studies have highlighted the importance of increased production associated with 307 ARs [92–94]. For example, Roa-Ureta et al. [94] predicted a 35% increase in the carrying capacity for 308 Diplodus vulgaris four years after ARs deployment, indicating new production. In terms of OWFs, 309 several studies have suggested that OWFs act as nursery grounds and therefore, facilitate new fish 310 production at the local scale [64,72,79,80]. In contrast, Wilson et al. [95] suggested that OWF 311 foundations attract species from neighbouring areas, and Bergström et al. [35] concluded that 312 elevated local abundance or diversity of fish near OWF foundations was due to a change in fish 313 distribution, rather than a result of increased productivity.

The attraction versus production debate remains unresolved in most cases and the outcome of the debate relies upon an array of factors. The responses towards introduced hard substrata may vary among species, environments, locations and age-specific requirements [90], while reef design characteristics, such as location, materials, size and number of units, also influence fish responses [91]. It is increasingly recognized that attraction and production effects are not mutually exclusive but should be considered as the two ends of a continuum [93,96–98].

# 320 4.3 Community changes associated with ARs and OWFs

Adding hard substrate may alter ecosystems in ways that benefit some species more than others [35,99]. For example, based on hydrodynamic modelling, introduced substrate from OWFs and ARs may increase the settlement of jellyfish polyps, potentially leading to jellyfish blooms [100,101]. This can affect fish species, as jellyfish competes with planktivorous fish and may forage on fish larvae [100]. OWFs may also attract fish species that would not naturally reside in the area [64]. For example, goldsinny wrasse and grey triggerfish (*Balistes carolinensis*) were observed in Dutch OWFs, where they had not been recorded previously [69].

328 Non-indigenous fouling species, possibly invasive, may also utilize hard substrate associated 329 with OWFs and ARs in soft-bottom environments [102]. Accordingly, OWFs may act as stepping-330 stones [103], enabling species to spread over large distances through a series of short distance 331 colonization events [63,102,104]; this may be particularly relevant for species like Jassa spp. that lack 332 a distinct planktonic larval stage [104]. Due to the stepping-stone effect, sequential establishment of 333 non-indigenous fouling species may occur rapidly on newly established OWF foundations [65,102]. 334 Considering ocean warming, the placement of new hard substrata on the seascape may also facilitate 335 the poleward expansion of non-indigenous species [105–107]. Observed non-indigenous species on 336 OWF foundations include the Japanese oyster (Crassostrea gigas) and amphipods Jassa marmorata, 337 Caprella mutica, and Caprella linearis [63,65,108,109]. In the Belgian part of the North Sea, ten non-338 indigenous fouling macrobenthic species were identified on OWFs foundations [102]. OWF sites may 339 also provide settlement habitat for pelagic larval particles that would otherwise have been 340 unsuccessful in the area [110].

341 4.4 Scour protection functioning as an AR

Reviewed literature suggests that OWFs provide similar functions for marine organisms as ARs [18,64,65,69–71,79,111,112], and OWF foundations have even been termed Windmill Artificial Reefs (WARs) [71,113]. These structures act as ARs by providing habitat, food, shelter and spawning opportunities, leading to the aggregation of various fish species around the foundations [65]. Importantly, scour protection enhances the habitat complexity and thereby augments the reef effect, as highlighted by previous studies [29,70,95,112,113].

348 4.4.1 Food provision

349 OWFs often increase the local food availability for fish. For example, Leonhard and Pedersen 350 [109] estimated that the scour protection at a Danish OWF provided a 50-fold increase in the local 351 food availability for fish, compared to the previously sandy area. The uneven and coarse surfaces of 352 scour protection and foundations allow sessile, fouling organisms to settle and provide food for 353 various fish species [66,114,115]. To examine the provision of prey items by scour protection, De 354 Troch et al. [66] profiled the energy levels of Atlantic cod and pouting caught just above the scour 355 protection. The study found that the scour protection provided both species with sufficient energy 356 reserves to grow and/or reproduce [66]. Moreover, stomach analyses performed on pouting 357 aggregating near scour protection found a preference for hard substrate prey, amphipod Jassa 358 herdmani and porcelain crab Pisidia longicornis, the dominating epifauna on the OWF foundations 359 [70,113]. Although further studies are needed, these findings suggest that scour protection may 360 provide feeding opportunities for a variety of fish species.

#### 361 4.4.2 Shelter

362 The layer of boulders and rocks commonly used for scour protection creates crevices that 363 provide hiding opportunities for fish and other fauna [113], consistent with similar mechanisms 364 occurring in natural boulder reefs [116]. Video footage and diving surveys revealed commercial 365 species such as Atlantic cod and pouting utilizing crevices associated with scour protection for shelter 366 [65,70]. Moreover, acoustic telemetry demonstrated that Atlantic cod exhibit strong residency, high 367 site fidelity and habitat selectivity towards OWF foundations with scour protection [71]. Therefore, 368 scour protection may serve as a potential fish refuge and shelter from predators and strong currents, 369 similar to ARs.

#### 370 4.4.3 Nursery

Several studies have observed nursery functions of scour protection [64,72,79,80]. For example, Krone et al. [68] revealed that OWF foundations function as nursery grounds for the edible crab, *Cancer pagurus*. Andersson and Öhman [72] detected pregnant two-spotted goby females, as well as increased numbers of several size-classes associated with scour protection. On this basis, the authors hypothesized that the habitat created by the OWF facilitated recruitment and increased production. These indications signify that OWFs may function as nursery grounds, although the spatial extent is

- 377 probably limited to the scour protection and the monopile alone.
- 378 4.4.4 Distance between artificial habitats

379 The location of OWFs may affect the composition and diversity of species assemblages that 380 colonize and utilize the structures. A shorter distance to neighbouring hard substrata provides a 381 greater likelihood of larvae and juveniles associated with local hard substrata to arrive and colonize 382 [18]. A reef in a given area has a specific carrying capacity; often determined by reef size, in 383 combination with the community composition (e.g. size of individuals and their foraging habits). To 384 create the most productive ARs, it is important to consider the potential overlap with forage areas of 385 neighbouring reef species. If the forage areas of two proximate ARs are overlapping, then the food 386 availability and the potential yield from each AR may decrease [117]. Recently, Rosemond et al. [118] 387 suggested a minimum buffer zone of 60m to 120m between neighbouring ARs to optimize the fish 388 habitat utilization. The proposed buffer zone aims to minimize attraction from proximate reefs, 389 maximize food resource availability in soft-bottom habitats around the reef and increase the area for 390 typical fish behaviours, such as reproduction and foraging. Similar buffer zones between scour 391 protections within OWF are *de facto* in place, since design guidelines highlight that the optimal 392 distance between adjacent monopiles and scour protections should be seven times the rotor diameter 393 of each turbine, resulting in distances exceeding ca. 430m between individual scour protections [119]. 394 These observations suggest that individual scour protections function as individual reefs.

395 4.4.5 Reef size

396 Numerous studies have indicated that reef volume and coverage area are crucial design 397 requirements for a successful AR. A typical scour protection has a high perimeter-to-area ratio, with 398 a radius up to 20 meters from the monopile [34,35], elevated about 1.5m above the seabed [38] (Figure 399 1); this is equivalent to a reef with a volume of 471m<sup>3</sup>. Smaller ARs (i.e. similar to the size of a scour 400 protection) often have increased fish densities in comparison to larger ARs [82,120,121]. This 401 observation, however, applies mainly for attraction and not necessarily for new production [121]. 402 Champion et al. [122] developed a model to examine the relationship between AR size and foraging 403 capacity. Their model predicted that highest per-capita food availability occurs when the AR is 404 relatively small. Ogawa et al. [123] reported that productivity of an AR increases directly with reef 405 size from 400 m<sup>3</sup> and up to 4000 m<sup>3</sup>, indicating that the minimum effective AR size is 400 m<sup>3</sup> [123,124]. 406 These AR findings suggest that the volume of an individual scour protection may be adequate to 407 function as a reef.

## 408 4.5 Optimizing scour protection

A fundamental objective of this review is to extract knowledge from AR designs and apply it to scour protection research with the aim to potentially enhance favourable ecological functions. Even though the ultimate purpose of scour protection is to prevent the scouring of sediment (Figure 1), scour protection may also provide preferred habitats for several species [112]. The level of complexity, the distance between artificial structures, as well as the building material and water depth are the primary characteristics that determine the efficacy of an artificial structure in terms of supporting fish abundance and biodiversity [12,112,118].

# 416 4.5.1 Structural complexity

417 Previous studies have observed a positive correlation between the structural complexity of ARs 418 and fish diversity and abundance [62,125]. Typically, larger fish require larger crevices, while smaller 419 fish prefer smaller crevices [126]. A heterogeneous structure with a variety of crevice sizes and shapes 420 may augment desired reef effects by elevating habitat heterogeneity, allowing a variety of fish species 421 and life stages to utilize the area [34]. Feeding efficiency and growth of fish are maximized at 422 intermediate levels of structural complexity, since very dense structures may hinder foraging [127]. 423 Required crevice variety may be achieved with the use of mixed-sized blocks or boulders that are 424 adjusted to the preferred shelter size for targeted species. In this fashion, scour protection design may 425 be tailored to certain species or life stages.

426 The level of complexity can also be adjusted by the number of holes and crevices present in the 427 applied material. For example, concrete modules with two holes may have a higher species 428 abundance than simpler modules without holes [62]. Purposefully designed concrete reef units, also 429 known as Reef Balls<sup>®</sup>, have 25 to 40 holes into a hollow centre [112]. Reef Balls<sup>®</sup> are used as coastal 430 defence structures and may also function as scour protection structures [112]. For a scour protection 431 area of ca. 440m<sup>2</sup>, a single layer of 169 Reef Balls<sup>®</sup> is required to protect one monopile [112]. The 432 projected carrying capacity of a Reef Ball<sup>®</sup> is approximately 385 kg of fish within a year, suggesting 433 that the annual carrying capacity of a scour protection may approach 65,000 kg [29,112]. Although 434 further research is required, these studies indicate that the structural complexity of scour protections 435 may be elevated to enhance fish abundance and may even be tailored to certain fish sizes and life 436 stages.

# 437 4.5.2 Materials

Different species settle on disparate substrates [128], suggesting that substrates may be tailored to suit target species. Rock is the most commonly used material for scour protection (Figure 5), because it is strong, stable, cost-efficient, erosion resistant and suitable for benthic flora and fauna settlement. Nonetheless, concrete-gravel aggregate may be more suitable for design manipulations [129,130]. In addition, a single scour protection may include various materials, such as boulders, gravel and synthetic fronds to replicate a natural range of habitats and further increase habitat heterogeneity [81,112]. While synthetic fronds alone provide less surface area and space for
colonization, they mimic vegetation and may provide additional ecological functions when combined
with other types of scour protection [29].

- 447 4.5.3 Enhancing ecological function
- 448 Although the large-scale effects of OWFs on marine ecosystems remain uncertain [69,73] and
- several studies have revealed negative effects of OWFs [131–133], the present review identified a
- 450 range of ecosystem functions that might be enhanced by modifying the foundation of OWFs and the
- 451 scour protection in particular (Figure 7). The conceptual illustration (Figure 7) is not exhaustive, and
- it remains crucial to develop and test structural manipulations of scour protection designs to achieve
- 453 favourable outcomes.

![](_page_16_Figure_10.jpeg)

454

Figure 7. Conceptual illustration of offshore wind turbine and proposed ecological functions of
scour protection. Reviewed evidence reveals that scour protection may provide food, shelter and
reproduction grounds for fish, as well as settlement grounds for bivalves and macro algae. Examples
show species that could benefit from improved scour protection designs: (A) Marine mammal feeding
grounds [134]; (B) Atlantic cod, utilizing scour protection structures [71]; (C) Atlantic herring,
utilizing scour protection for spawning [89]; (D) Mussel and oyster banks [135]; (E) Macro algal
restoration [136,137]; (F) Shelter for shellfish [91].

462 Mammalian and fish apex predators may utilize OWFs for foraging; GPS tracking of two seal 463 species (Phoca vitulina and Halichoerus grypus) in the North Sea indicated foraging close to the OWF 464 foundations [134] (Figure 7A). Similarly, significantly higher occurrence of harbour porpoises was 465 recorded inside a Dutch OWF compared to reference areas, partly attributed to increased food 466 availability [138]. Various fish species may exhibit strong residency at OWF foundations with scour 467 protection [71] (e.g. Atlantic cod; Figure 7B), which are also included in the diet of the two seal species 468 [139,140]. This indicates that modified scour protection designs enhancing fish abundance could 469 translate into benefits for a range of apex predators.

470 Rough surface texture in the materials used for reef construction can enhance benthic settlement 471 [115,141,142] (Figure 7D, 7E). Orientation of surfaces also plays an important role; bivalves, hydroids 472 and barnacles mainly prefer colonizing vertical surfaces [135] (Figure 7D). High-vertical relief 473 structures often support fish recruitment [143] and planktivore abundances, possibly by creating 474 suitable hydrodynamic conditions [64]. Therefore, combined manipulations of materials and 475 orientations may incite the growth of specific assemblages (Figure 7D, 7E). This may even be 476 extended to setting up multi-use systems on OWFs, including, for example, mussel cultivation 477 (Figure 7D).

478 Moreover, the macroalgal colonization of scour protection may be augmented using macroalgae 479 restoration techniques [136,137,144], as macroalgae has been shown to influence the composition and 480 abundance of fish species on rocky habitats [145]. Verdura et al. [136] developed a relevant novel 481 technique to restore patches of the canopy-forming alga Cystoseira barbata, while Fredriksen et al. 482 [137] developed a method to restore kelp forests using rocks (Figure 7E). Macro algae growth may 483 also support fish reproduction (Figure 7C). For example, in the Baltic Sea, herring select red algae 484 Furcellaria lumbricalis and Polysiphonia fucoides as spawning substrate [89] (Figure 7C, 7E). This 485 function requires that scour protection is in the photic zone, requiring shallow water OWFs. Such 486 cases exist, however, including the "Horns Rev 1" OWF in the North Sea, which is situated at water 487 depths between 6.5 - 13.5 m [73].

488 Shelter availability for shellfish species, such as the European lobster, Homarus gammarus, and 489 C. pagurus, may create a demographic bottleneck [91]. Scour protection optimizations in favour of the 490 European lobster could include size, shape, number of shelter entrances and substrate [146,147] 491 (Figure 7F). Typically, adult lobsters and crayfish select crevices that have entrances analogous to 492 their body size [148], while juveniles (carapace length, CL < 35mm) are active burrowers and find 493 shelter in soft substrates [149]. Barry and Wickins [149] proposed a mathematical model to estimate 494 the best rock size combination to produce holes suitable for individuals with CL between 38 - 145mm. 495 The model estimated that a 10-size rock solution (ranging from 40 to 3000 mm radius) produced the 496 best results. Even though their model is a simplified version of reality, because the modelled rocks 497 are spherical and therefore result in a denser packing than what might occur naturally, the model 498 may still provide rough recommendations for the optimal rock size combination for scour protection 499 designs (Figure 7F). If specific lobster sizes are targeted, the following simplified equation (1) could 500 be used to calculate the appropriate rock size radius (r) that will create the required target hole radius 501 (*x*) [149]:

502

$$x = 0.15r \tag{1}$$

503 For example, assuming that the rocks are spherical and homogenous in size, rocks with a radius 504 of 1078 mm should be used to benefit lobsters with a CL of 145 mm [149]. To the best of our 505 knowledge, preferred shelter sizes have not been quantified for fish, although such data could prove 506 valuable to tailor scour protection designs for specific fish species and life stages.

# 507 4.6 Advancing the field

508 Soft-bottom species might be vulnerable when hard substrata (e.g. foundations) are introduced 509 in their ecosystem, because soft-bottom habitat is eliminated. Due to conflicting results, further 510 studies are required to determine the impact of scour protection on soft-bottom species. Likewise, 511 edge and transition zone effects [150–155], as well as community recovery models [156–158] should 512 also be taken into consideration. Given the complexity of patchy habitats induced by scour protection, 513 such studies may reveal unique ecological conditions existing in the transition zones between scour 514 protection and soft-bottom habitats, as well as how soft-bottom species recover after partial loss of 515 their habitat. 516 To date, no empirical studies have examined ecological effects of different scour protection

517 configurations. It is crucial to identify the size, shape, height, orientation and material that provide 518 most support for targeted ecosystem functions and services. The use of assorted scour protection 519 materials should also receive further attention. Future studies should test concrete (possibly in Reef Ball<sup>®</sup> configuration) as an alternative material for scour protection. Further studies on preferred shelter dimensions for fishes are also needed to predict the optimal rock size for scour protections to provide ideal habitats for target species. Coastal defence structures have also been shown to offer opportunities for supporting ecological functions [11,159]. Even though these are not offshore structures, foundation designs of coastal OWFs could possibly benefit from coastal-defence knowledge [11,159]. This knowledge remains to be tested and integrated into cases where OWFs are being deployed in coastal areas.

527 Many OWF studies concern areas that are *de facto* no fishing zones, since trawling and other 528 fishing activities often are prohibited within a radius from the structures. These no-take conditions 529 essentially create a marine protected area that may increase the abundance of fish in neighbouring 530 areas through spill-over effect [28–30,160]. The spill-over could be harvested by sustainable fisheries, 531 but the topic has received limited attention. The main knowledge gap is estimating how much 532 additional fish biomass is added to the ecosystem (at a larger scale) due to the presence and design 533 of scour protection. Mounting evidence suggest spill-over effects of marine protected areas [161,162], 534 but it remains unknown to what degree the effects depend on scour protections inside OWFs.

535 Existing evidence indicates that OWFs cause little to no effects on fish communities on a large 536 spatial scale; however, due to the expected massive increase in OWFs [18,19], the scale of OWF effects 537 is likely to expand over the coming years. With increased OWF development, effects of scour 538 protection could reach a tipping point (i.e. induce a marine regime shift) [163,164], where a certain 539 density of scour protections may have significant effects on large-scale fish communities.

540 Apart from the scour protection, effects of the monopiles (or equivalent structures) should also 541 be considered when determining changes in biodiversity and species abundance, as the community 542 composition of the fouling fauna may differ along a depth gradient on a monopile [165]. Monopiles 543 are, however, considered difficult to manipulate, opposite to the scour protection and were therefore 544 excluded from the present review. Interestingly, monopiles may attract mid-water forage species 545 [64,69] and associated predators, pointing to similar effects recorded at oil and gas platforms [15,166– 546 169]. There are clearly interactions between these mid-water species and the species linked to scour 547 protection, which should be further explored for a better understanding of community changes.

548 Lastly, when the production life of offshore infrastructure is over, current international policies 549 call for removal of the structures. This type of decommissioning, however, could result in loss of 550 marine habitat, especially considering species benefitting from OWFs [16,99,170]. Before 551 decommissioning of the first OWF in the world (Vindeby, Denmark), the associated environmental 552 impact assessment indicated that the removal of the foundation could reduce the local Atlantic cod 553 abundance [171]. The impact on Atlantic cod was not attributed to the foundation removal activities 554 per se, but rather to the loss of reef habitat in the area, after eliminating the foundations [171]. 555 Therefore, before removing OWF foundations, reefing options may be considered [172,173]. An 556 alternative to removal could be the "renewables-to-reefs" program, in which the foundation is 557 partially removed, leaving the scour protection intact on the seafloor [174,175]. Modifying scour 558 protection designs to favour fish abundance and diversity may amplify interests in keeping parts of 559 the foundation intact during decommissioning.

# 560 5. Conclusions

561 This review is the first to transfer knowledge from AR studies to scour protection designs. 562 Specifically, the review synthesized the effects of ARs and scour protection on fish biota and 563 suggested approaches for enhancing the ecological function of scour protection. Overall, the vast 564 majority of both AR and OWF studies report increases in fish abundance and diversity within the 565 local area. Current evidence suggests that fish species of both high and low commercial value utilize 566 the hard substrata that scour protection adds to the marine environment. This review has identified 567 several scour protection manipulations that could influence abundance and diversity of fish species. 568 By modifying future scour protection designs, fish abundance and diversity may be enhanced. As 569 such, scour protection designs may be tailored to commercial species or threatened species. Further 670 empirical research is needed to explore and test new and innovative scour protection designs, with671 the ultimate goal of creating scour protections that support marine goods and ecosystem services.

572 **Author Contributions:** Conceptualization, J.C.S.; methodology, J.C.S. and M.Z.; validation, J.C.S. and M.Z.; 573 formal analysis, M.G.; investigation, M.G. and M.Z.; resources, J.C.S.; data curation, M.G.; writing—original

draft preparation, M.G. and M.Z.; writing – review and editing, M.G. and J.C.S.; visualization, M.G.; supervision,

575 J.C.S.; project administration, J.C.S.; funding acquisition, J.C.S. All authors have read and agreed to the published

- 576 version of the manuscript.
- Funding: This research was partly funded by the European Maritime and Fisheries Fund (grant number: 33113 B-16-057) and Vattenfall AB.

579 **Acknowledgments:** We sincerely thank the anonymous reviewers for their constructive and valuable comments.

580 **Conflicts of Interest:** The authors declare no conflict of interest. The funders had no role in the design of the

581 study; in the collection, analyses, or interpretation of data; in the writing of the manuscript, or in the decision to 582 publish the results.

#### 583 References

- Halpern, B.S.; Walbridge, S.; Selkoe, K.A.; Kappel, C. V.; Micheli, F.; D'Agrosa, C.; Bruno, J.F.; Casey,
   K.S.; Elbert, C.; Fox, H.E.; et al. A Global Map of Human Impact on Marine Ecosystems. *Science* 2008,
   319, 948–953.
- Halpern, B.S.; Longo, C.; Hardy, D.; McLeod, K.L.; Samhouri, J.F.; Katona, S.K.; Kleisner, K.; Lester, S.E.;
   Oleary, J.; Ranelletti, M.; et al. An index to assess the health and benefits of the global ocean. *Nature* 2012,
   488, 615–620.
- Waltham, N.J.; Connolly, R.M. Global extent and distribution of artificial, residential waterways in
  estuaries. *Estuar. Coast. Shelf Sci.* 2011, 94, 192–197.
- Dafforn, K.A.; Glasby, T.M.; Airoldi, L.; Rivero, N.K.; Mayer-Pinto, M.; Johnston, E.L. Marine
   urbanization: An ecological framework for designing multifunctional artificial structures. *Front. Ecol. Environ.* 2015, *13*, 82–90.
- 595 5. Heery, E.C.; Hoeksema, B.W.; Browne, N.K.; Reimer, J.D.; Ang, P.O.; Huang, D.; Friess, D.A.; Chou, L.M.;
  596 Loke, L.H.L.; Saksena-Taylor, P.; et al. Urban coral reefs: Degradation and resilience of hard coral
  597 assemblages in coastal cities of East and Southeast Asia. *Mar. Pollut. Bull.* 2018, 135, 654–681.
- 598 6. Bulleri, F.; Chapman, M.G. The introduction of coastal infrastructure as a driver of change in marine
  599 environments. J. Appl. Ecol. 2010, 47, 26–35.
- Bishop, M.J.; Mayer-Pinto, M.; Airoldi, L.; Firth, L.B.; Morris, R.L.; Loke, L.H.L.; Hawkins, S.J.; Naylor,
  L.A.; Coleman, R.A.; Chee, S.Y.; et al. Effects of ocean sprawl on ecological connectivity: impacts and
  solutions. J. Exp. Mar. Bio. Ecol. 2017, 492, 7–30.
- 8. Heery, E.C.; Bishop, M.J.; Critchley, L.P.; Bugnot, A.B.; Airoldi, L.; Mayer-Pinto, M.; Sheehan, E. V.;
  Coleman, R.A.; Loke, L.H.L.; Johnston, E.L.; et al. Identifying the consequences of ocean sprawl for
  sedimentary habitats. *J. Exp. Mar. Bio. Ecol.* 2017, 492, 31–48.
- 606 9. Bortone, S.A.; Seaman, W. Technology for the creation of aquatic habitats and their evaluation in 607 fisheries ecosystems. *Fish. Sci.* **2002**, *68*, 1677–1682.
- 608 10. Seaman, W. Artificial habitats and the restoration of degraded marine ecosystems and fisheries.
  609 *Hydrobiologia* 2007, 580, 143–155.
- Firth, L.B.; Thompson, R.C.; Bohn, K.; Abbiati, M.; Airoldi, L.; Bouma, T.J.; Bozzeda, F.; Ceccherelli, V.U.;
  Colangelo, M.A.; Evans, A.; et al. Between a rock and a hard place: Environmental and engineering
  considerations when designing coastal defence structures. *Coast. Eng.* 2014, *87*, 122–135.
- 613 12. Baine, M. Artificial reefs: a review of their design, application, management and performance. *Ocean*

614		Coast. Manag. 2001, 44, 241–259.
615	13.	Tessier, A.; Francour, P.; Charbonnel, E.; Dalias, N.; Bodilis, P.; Seaman, W.; Lenfant, P. Assessment of
616		French artificial reefs: due to limitations of research, trends may be misleading. Hydrobiologia 2015, 753,
617		1–29.
618	14.	Reubens, J.T.; Braeckman, U.; Vanaverbeke, J.; Van Colen, C.; Degraer, S.; Vincx, M. Aggregation at
619		windmill artificial reefs: CPUE of Atlantic cod (Gadus morhua) and pouting (Trisopterus luscus) at
620		different habitats in the Belgian part of the North Sea. Fish. Res. 2013, 139, 28–34.
621	15.	Fujii, T. Potential influence of offshore oil and gas platforms on the feeding ecology of fish assemblages
622		in the North Sea. Mar. Ecol. Prog. Ser. 2016, 542, 167–186.
623	16.	Fowler, A.M.; Jørgensen, A.; Coolen, J.W.P.; Brabant, R.; Rumes, B.; Degraer, S. The ecology of
624		infrastructure decommissioning in the North Sea: what we need to know and how to achieve it. ICES J.
625		Mar. Sci. 2019.
626	17.	Edney, J.; Spennemann, D.H.R. Can Artificial Reef Wrecks Reduce Diver Impacts on Shipwrecks? The
627		Management Dimension. J. Marit. Archaeol. 2015, 10, 141–157.
628	18.	Petersen, J.K.; Malm, T. Offshore Windmill Farms: Threats to or Possibilities for the Marine
629		Environment. Ambio 2006, 35, 75–80.
630	19.	Ramírez, L.; Fraile, D.; Brindley, G. Offshore wind in Europe - Key trends and statistics 2019. WindEurope
631		<b>2020</b> , 1–38.
632	20.	Drewitt, A.L.; Langston, R.H.W. Assessing the impacts of wind farms on birds. Ibis (Lond. 1859). 2006,
633		148, 29–42.
634	21.	Thaxter, C.B.; Buchanan, G.M.; Carr, J.; Butchart, S.H.M.; Newbold, T.; Green, R.E.; Tobias, J.A.; Foden,
635		W.B.; O'Brien, S.; Pearce-Higgins, J.W. Bird and bat species' global vulnerability to collision mortality at
636		wind farms revealed through a trait-based assessment. Proc. R. Soc. B Biol. Sci. 2017, 284.
637	22.	Wahlberg, M.; Westerberg, H. Hearing in fish and their reactions to sounds from offshore wind farms.
638		Mar. Ecol. Prog. Ser. 2005, 288, 295–309.
639	23.	Madsen, P.T.; Wahlberg, M.; Tougaard, J.; Lucke, K.; Tyack, P. Wind turbine underwater noise and
640		marine mammals: implications of current knowledge and data needs. Mar. Ecol. Prog. Ser. 2006, 309, 279–
641		295.
642	24.	Brandt, M.J.; Dragon, A.C.; Diederichs, A.; Bellmann, M.A.; Wahl, V.; Piper, W.; Nabe-Nielsen, J.; Nehls,
643		G. Disturbance of harbour porpoises during construction of the first seven offshore wind farms in
644		Germany. <i>Mar. Ecol. Prog. Ser.</i> 2018, 596, 213–232.
645	25.	Öhman, M.C.; Sigray, P.; Westerberg, H. Offshore windmills and the effects of electromagnetic fields on
646		fish. <i>Ambio</i> <b>2007</b> , <i>36</i> , 630–633.
647	26.	Bray, L.; Reizopoulou, S.; Voukouvalas, E.; Soukissian, T.; Alomar, C.; Vázquez-Luis, M.; Deudero, S.;
648		Attrill, M.J.; Hall-Spencer, J.M. Expected effects of offshore wind farms on Mediterranean Marine Life.
649		J. Mar. Sci. Eng. <b>2016</b> , 4, 18.
650	27.	Langhamer, O.; Wilhelmsson, D. Colonisation of fish and crabs of wave energy foundations and the
651		effects of manufactured holes - A field experiment. Mar. Environ. Res. 2009, 68, 151-157.
652	28.	Punt, M.J.; Groeneveld, R.A.; van Ierland, E.C.; Stel, J.H. Spatial planning of offshore wind farms: A
653		windfall to marine environmental protection? Ecol. Econ. 2009, 69, 93–103.
654	29.	Langhamer, O. Artificial reef effect in relation to offshore renewable energy conversion: State of the art.
655		Sci. World J. <b>2012</b> , 2012.
656	30.	Ashley, M.C.; Mangi, S.C.; Rodwell, L.D. The potential of offshore windfarms to act as marine protected

657		areas - A systematic review of current evidence. Mar. Policy 2014, 45, 301–309.
658	31.	den Boon, J.; Sutherland, J.; Whitehouse, R. Scour behaviour and scour protection for monopile
659		foundations of offshore wind turbines. In Proceedings of the European Wind Energy Conference &
660		Exhibition (EWEC); London, U.K., 2004; Vol. 14, pp. 1–14.
661	32.	Sumer, B.M.; Fredsøe, J. The Mechanics of Scour in the Marine Environment; River Edge, N.J., World
662		Scientific., 2002; Vol. 17; pp. 1–552.
663	33.	Atalah, J.; Fitch, J.; Coughlan, J.; Chopelet, J.; Coscia, I.; Farrell, E. Diversity of demersal and megafaunal
664		assemblages inhabiting sandbanks of the Irish Sea. Mar. Biodivers. 2013, 43, 121–132.
665	34.	Hammar, L.; Andersson, S.; Rosenberg, R. Adapting offshore wind power foundations to local
666		environment. Swedish Environ. Prot. Agency 2010, 1–86.
667	35.	Bergström, L.; Sundqvist, F.; Bergström, U. Effects of an offshore wind farm on temporal and spatial
668		patterns in the demersal fish community. Mar. Ecol. Prog. Ser. 2013, 485, 199–210.
669	36.	De Vos, L.; De Rouck, J.; Troch, P.; Frigaard, P. Empirical design of scour protections around monopile
670		foundations. Part 2: Dynamic approach. Coast. Eng. 2012, 60, 286–298.
671	37.	Liang, B.; Du, S.; Pan, X.; Zhang, L. Local scour for vertical piles in steady currents: Review of
672		mechanisms, influencing factors and empirical equations. J. Mar. Sci. Eng. 2020, 8.
673	38.	Whitehouse, R.J.S.; Harris, J.M.; Sutherland, J.; Rees, J. The nature of scour development and scour
674		protection at offshore windfarm foundations. Mar. Pollut. Bull. 2011, 62, 73–88.
675	39.	Steffen, W.; Crutzen, P.J.; Mcneill, J. The Anthropocene: Are Humans Now Overwhelming the Great
676		Forces of Nature? <i>Ambio</i> <b>2007</b> , <i>36</i> , 614–621.
677	40.	Scheffers, B.R.; De Meester, L.; Bridge, T.C.L.; Hoffmann, A.A.; Pandolfi, J.M.; Corlett, R.T.; Butchart,
678		S.H.M.; Pearce-Kelly, P.; Kovacs, K.M.; Dudgeon, D.; et al. The broad footprint of climate change from
679		genes to biomes to people. Science 2016, 354.
680	41.	Bowman, D.M.J.S.; Garnett, S.T.; Barlow, S.; Bekessy, S.A.; Bellairs, S.M.; Bishop, M.J.; Bradstock, R.A.;
681		Jones, D.N.; Maxwell, S.L.; Pittock, J.; et al. Renewal ecology: conservation for the Anthropocene. Restor.
682		<i>Ecol.</i> <b>2017</b> , <i>25</i> , 674–680.
683	42.	Maes, J.; Jacobs, S. Nature-Based Solutions for Europe's Sustainable Development. Conserv. Lett. 2017,
684		10, 121–124.
685	43.	Lenkeek, W.; Didderen, K.; Teunis, M.; Driessen, F.; Coolen, J.W.P.; Bos, O.G.; Vergouwen, S.A.;
686		Raaijmakers, T.C.; de Vries, M.B.; van Koningsveld, M. Eco-friendly design of scour protection: potential
687		enhancement of ecological functioning in offshore wind farms. Minist. Econ. Aff. Dep. Nat. Biodivers. 2017,
688		28.
689	44.	Morris, R.L.; Porter, A.G.; Figueira, W.F.; Coleman, R.A.; Fobert, E.K.; Ferrari, R. Fish-smart seawalls: a
690		decision tool for adaptive management of marine infrastructure. Front. Ecol. Environ. 2018, 16, 278–287.
691	45.	Strain, E.M.A.; Olabarria, C.; Mayer-Pinto, M.; Cumbo, V.; Morris, R.L.; Bugnot, A.B.; Dafforn, K.A.;
692		Heery, E.; Firth, L.B.; Brooks, P.R.; et al. Eco-engineering urban infrastructure for marine and coastal
693		biodiversity: Which interventions have the greatest ecological benefit? J. Appl. Ecol. 2018, 55, 426-441.
694	46.	Dafforn, K.A.; Mayer-Pinto, M.; Morris, R.L.; Waltham, N.J. Application of management tools to
695		integrate ecological principles with the design of marine infrastructure. J. Environ. Manage. 2015, 158, 61–
696		73.
697	47.	Pullin, A.S.; Stewart, G.B. Guidelines for systematic review in conservation and environmental
698		management. Conserv. Biol. 2006, 20, 1647–1656.

699 48. Papathanasopoulou, E.; Beaumont, N.; Hooper, T.; Nunes, J.; Queirós, A.M. Energy systems and their

22 of 29

700		impacts on marine ecosystem services. Renew. Sustain. Energy Rev. 2015, 52, 917–926.
701	49.	Flávio, H.M.; Ferreira, P.; Formigo, N.; Svendsen, J.C. Reconciling agriculture and stream restoration in
702		Europe: a review relating to the EU Water Framework Directive. Sci. Total Environ. 2017, 596-597, 378-
703		395.
704	50.	Moher, D.; Libersti, A.; Tetzlaff, J.; Altman, D.G.; Group, T.P. Preferred Reporting Items for Systematic
705		Reviews and Meta-Analyses: The PRISMA Statement. PLoS Med. 2009, 6, e1000097.
706	51.	Whitmarsh, S.K.; Fairweather, P.G.; Huveneers, C. What is Big BRUVver up to? Methods and uses of
707		baited underwater video. Rev. Fish Biol. Fish. 2017, 27, 53–73.
708	52.	Støttrup, J.G.; Dahl, K.; Niemann, S.; Stenberg, C.; Reker, J.; Stamphøj, E.M.; Göke, C.; Svendsen, J.C.
709		Restoration of a boulder reef in temperate waters: Strategy, methodology and lessons learnt. <i>Ecol. Eng.</i>
710		2017, 102, 468–482.
711	53.	Kristensen, L.D.; Støttrup, J.G.; Svendsen, J.C.; Stenberg, C.; Højbjerg Hansen, O.K.; Grønkjær, P.
712		Behavioural changes of Atlantic cod (Gadus morhua) after marine boulder reef restoration: Implications
713		for coastal habitat management and Natura 2000 areas. Fish. Manag. Ecol. <b>2017</b> , 24, 353–360.
714	54.	Kavano, Y. Effects of artificial reefs and the acoustic-sound feeding method on the colonization of
715	•	released grouper seedlings. <i>Fish. Sci.</i> <b>2002</b> , <i>68</i> , 1683–1686.
716	55.	Castège, L: Milon, E : Fourneau, G : Tauzia, A. First results of fauna community structure and dynamics
717		on two artificial reefs in the south of the Bay of Biscay (France). Estuar Coast Shelf Sci <b>2016</b> 179 172-
718		180
719	56	Carral L. Alvarez-Feal I.C. Tarrio-Saavedra I. Rodriguez Guerreiro M.I. Fraguela I.Á. Social interest
720	00.	in developing a green modular artificial reef structure in concrete for the ecosystems of the Galician rías
721		I Clean Prod 2018 172 1881–1898
722	57	Yang Y : Ji T : Lin X : Chen C : Yang Z Biogenic sulfuric acid corrosion resistance of new artificial reef
723	07.	concrete Constr. Build Mater 2018 158 33-41
724	58	Sánchez S: López-Gutiérrez IS: Negro V: Esteban MD Foundations in offshore wind farms:
725	00.	Evolution characteristics and range of use Analysis of main dimensional parameters in monopile
726		foundations I Mar Sci Fing 2019 7
727	59	Abecasis D. Bentes I. Lino P.G. Santos M.N. Frzini K. Residency movements and habitat use of
728	07.	adult white seabream (Diplodus sargus) between natural and artificial reefs. Estuar Coast Shelf Sci 2013
729		118 80–85
730	60	Cresson P: Ruitton S: Harmelin-Vivien M Artificial reefs do increase secondary biomass production:
731	00.	Mechanisms evidenced by stable isotopes Mar. Ecol. Prog. Ser. 2014, 509, 15–26
732	61	La Mesa G. Longobardi A. Sacco F. Marino G. First Release of Hatchery Inveniles of the Dusky
733	01.	Grouper Epipephelus marginatus (Lowe 1834) (Serrapidae: Teleostei) at Artificial Reefs in the
734		Mediterranean: Results from a Pilot Study. Sci Mar 2008 72 743–756
735	62	Hunter WR Saver MDII The comparative effects of habitat complexity on faunal assemblages of
736	02.	northern temperate artificial and natural reefs. <i>ICES I. Mar. Sci.</i> <b>2009</b> , 66, 691–698
737	63	Wilhelmsson D: Malm T Fouling assemblages on offshore wind power plants and adjacent substrata
738	00.	Fstuar Coast Shelf Sci 2008 79 459-466
739	64	Wilhelmsson $D$ : Malm T: Öhman M.C. The influence of offshore windpower on demersal fish <i>ICES</i>
740	01.	I Mar Sci 2006 63 775–784
741	65	Lindeboom HI-Kouwenhoven HI-Rergman MIN-Rouma S-Brassour S-Daan R-Fiin P.C-Da
742	00.	Haan D. Dirkson S. Van Hal R. et al Short-term ecological effects of an offebore wind form in the
742		Haan, D.; Dirksen, S.; Van Hal, R.; et al. Short-term ecological effects of an offshore wind farm in the

743		Dutch coastal zone; A compilation. Environ. Res. Lett. 2011, 6.
744	66.	De Troch, M.; Reubens, J.T.; Heirman, E.; Degraer, S.; Vincx, M. Energy profiling of demersal fish: A
745		case-study in wind farm artificial reefs. Mar. Environ. Res. 2013, 92, 224–233.
746	67.	Bergman, M.J.N.; Ubels, S.M.; Duineveld, G.C.A.; Meesters, E.W.G. Effects of a 5-year trawling ban on
747		the local benthic community in a wind farm in the Dutch coastal zone. ICES J. Mar. Sci. 2015, 72, 962-
748		972.
749	68.	Krone, R.; Dederer, G.; Kanstinger, P.; Krämer, P.; Schneider, C.; Schmalenbach, I. Mobile demersal
750		megafauna at common offshore wind turbine foundations in the German Bight (North Sea) two years
751		after deployment - increased production rate of Cancer pagurus. Mar. Environ. Res. 2017, 123, 53-61.
752	69.	van Hal, R.; Griffioen, A.B.; van Keeken, O.A. Changes in fish communities on a small spatial scale, an
753		effect of increased habitat complexity by an offshore wind farm. Mar. Environ. Res. 2017, 126, 26–36.
754	70.	Reubens, J.T.; Degraer, S.; Vincx, M. Aggregation and feeding behaviour of pouting (Trisopterus luscus)
755		at wind turbines in the Belgian part of the North Sea. Fish. Res. 2011, 108, 223–227.
756	71.	Reubens, J.T.; Pasotti, F.; Degraer, S.; Vincx, M. Residency, site fidelity and habitat use of atlantic cod
757		(Gadus morhua) at an offshore wind farm using acoustic telemetry. Mar. Environ. Res. 2013, 90, 128–135.
758	72.	Andersson, M.H.; Öhman, M.C. Fish and sessile assemblages associated with wind-turbine
759		constructions in the Baltic Sea. Mar. Freshw. Res. 2010, 61, 642–650.
760	73.	Stenberg, C.; Støttrup, J.G.; van Deurs, M.; Berg, C.W.; Dinesen, G.E.; Mosegaard, H.; Grome, T.M.;
761		Leonhard, S.B. Long-term effects of an offshore wind farm in the North Sea on fish communities. Mar.
762		Ecol. Prog. Ser. 2015, 528, 257–265.
763	74.	van Deurs, M.; Grome, T.M.; Kaspersen, M.; Jensen, H.; Stenberg, C.; Sørensen, T.K.; Støttrup, J.; Warnar,
764		T.; Mosegaard, H. Short- and long-term effects of an offshore wind farm on three species of sandeel and
765		their sand habitat. <i>Mar. Ecol. Prog. Ser.</i> <b>2012</b> , 458, 169–180.
766	75.	Langhamer, O.; Dahlgren, T.G.; Rosenqvist, G. Effect of an offshore wind farm on the viviparous eelpout:
767		Biometrics, brood development and population studies in Lillgrund, Sweden. <i>Ecol. Indic.</i> <b>2018</b> , <i>84</i> , 1–6.
768	76.	Fabi, G.; Fiorentini, L. Comparison between an artificial reef and a control site in the Adriatic Sea:
769		analysis of four years of monitoring. Bull. Mar. Sci. <b>1994</b> , 55, 538–558.
770	77.	Bombace, G.; Fabi, G.; Fiorentini, L.; Speranza, S. Analysis Of The Efficacy Of Artificial Reefs Located in
771		Five Different Areas Of The Adriatic Sea. Bull. Mar. Sci. <b>1994</b> , 55, 559–580.
772	78.	Wilber, D.H.; Carey, D.A.; Griffin, M. Flatfish habitat use near North America's first offshore wind farm.
773		J. Sea Res. <b>2018</b> , 139, 24–32.
774	79.	Reubens, J.T.; Degraer, S.; Vincx, M. The ecology of benthopelagic fishes at offshore wind farms: A
775		synthesis of 4 years of research. <i>Hydrobiologia</i> <b>2014</b> , 727, 121–136.
776	80.	Reubens, J.T.; Vandendriessche, S.; Zenner, A.N.; Degraer, S.; Vincx, M. Offshore wind farms as
777		productive sites or ecological traps for gadoid fishes? - Impact on growth, condition index and diet
778		composition. <i>Mar. Environ. Res.</i> <b>2013</b> , 90, 66–74.
779	81.	Bat, L.; Sezgin, M.; Sahin, F. Impacts of OWF installations on fisheries: A Literature Review. J. Coast. Life
780		Med. <b>2013</b> , <i>1</i> , 241–252.
781	82.	Bohnsack, J.A.; Harper, D.E.; McClellan, D.B.; Hulsbeck, M. Effects of Reef Size on Colonization and
782		Assemblage Structure of Fishes At Artificial Reefs Off Southeastern Florida, U.S.A. Bull. Mar. Sci. 1994,
783		55, 796–823.
784	83.	Methratta, E.T.; Dardick, W.R.; Methratta, E.T.; Dardick, W.R. Meta-Analysis of Finfish Abundance at
785		Offshore Wind Farms. Rev. Fish. Sci. Aquac. 2019, 27, 242–260.
		,

J. Mar. Sci. Eng. 2020, 8, x FOR PEER REVIEW

786	84.	Yu, J.; Chen, P.; Tang, D.; Qin, C. Ecological effects of artificial reefs in Daya Bay of China observed from
787		satellite and in situ measurements. Adv. Sp. Res. 2015, 55, 2315–2324.
788	85.	Fabi, G.; Manoukian, S.; Spagnolo, A. Feeding behavior of three common fishes at an artificial reef in the
789		northern Adriatic Sea. Bull. Mar. Sci. 2006, 78, 39–56.
790	86.	Aneer, G.; Nellbring, S. A SCUBA-diving investigation of Baltic herring (Clupea harengus membras L.)
791		spawning grounds in the Asko-Landsort area, northern Baltic proper. J. Fish Biol. 1982, 21, 433-442.
792	87.	Johannessen, A. Recruitment studies of herring (Clupea harengus L.) in Lindaaspollene, western
793		Norway. Fisk. Skr. Ser. Havundersøkelser <b>1986</b> , 18, 139–240.
794	88.	Kääriä, J.; Rajasilta, M.; Kurkilahti, M.; Soikkeli, M. Spawning bed selection by the Baltic herring (Clupea
795		harengus membras) in the Archipelago of SW Finland. ICES J. Mar. Sci. 1997, 54, 917–923.
796	89.	Šaškov, A.; Šiaulys, A.; Bučas, M.; Daunys, D. Baltic herring (Clupea harengus membras) spawning
797		grounds on the Lithuanian coast: current status and shaping factors. Oceanologia 2014, 56, 789–804.
798	90.	Brickhill, M.J.; Lee, S.Y.; Connolly, R.M. Fishes associated with artificial reefs: Attributing changes to
799		attraction or production using novel approaches. J. Fish Biol. 2005, 67, 53–71.
800	91.	Pickering, H.; Whitmarsh, D. Artificial reefs and fisheries exploitation: A review of the "attraction versus
801		production" debate, the influence of design and its significance for policy. Fish. Res. 1997, 31, 39–59.
802	92.	Granneman, J.E.; Steele, M.A. Fish Growth, Reproduction, and Tissue Production on Artificial Reefs
803		Relative to Natural Reefs. ICES J. Mar. Sci. 2014, 71, 2494–2504.
804	93.	Cresson, P.; Le Direach, L.; Rouanet, E.; Goberville, E.; Astruch, P.; Ourgaud, M.; Harmelin-Vivien, M.
805		Functional traits unravel temporal changes in fish biomass production on artificial reefs. Mar. Environ.
806		<i>Res.</i> <b>2019</b> , <i>145</i> , 137–146.
807	94.	Roa-Ureta, R.H.; Santos, M.N.; Leitão, F. Modelling long-term fisheries data to resolve the attraction
808		versus production dilemma of artificial reefs. Ecol. Modell. 2019, 407, 108727.
809	95.	Wilson, J.C.; Elliott, M.; Cutts, N.D.; Mander, L.; Mendão, V.; Perez-Dominguez, R.; Phelps, A. Coastal
810		and offshore wind energy generation: Is it environmentally benign? Energies 2010, 3, 1383–1422.
811	96.	Osenberg, C.W.; St. Mary, C.M.; Wilson, J.A.; Lindberg, W.J. A quantitative framework to evaluate the
812		attraction-production controversy. ICES J. Mar. Sci. 2002, 59, S214–S221.
813	97.	Scarcella, G.; Grati, F.; Bolognini, L.; Domenichetti, F.; Malaspina, S.; Manoukian, S.; Polidori, P.;
814		Spagnolo, A.; Fabi, G. Time-series analyses of fish abundance from an artificial reef and a reference area
815		in the central-Adriatic Sea. J. Appl. Ichthyol. 2015, 31, 74–85.
816	98.	Schwartzbach, A.; Behrens, J.; Svendsen, J. Atlantic cod Gadus morhua save energy on stone reefs:
817		implications for the attraction versus production debate in relation to reefs. Mar. Ecol. Prog. Ser. 2020,
818		635, 81–87.
819	99.	Gill, A.B. Offshore renewable energy: Ecological implications of generating electricity in the coastal zone.
820		J. Appl. Ecol. <b>2005</b> , 42, 605–615.
821	100.	Janßen, H.; Augustin, C.B.; Hinrichsen, H.H.; Kube, S. Impact of secondary hard substrate on the
822		distribution and abundance of Aurelia aurita in the western Baltic Sea. Mar. Pollut. Bull. 2013, 75, 224-
823		234.
824	101.	Makabe, R.; Furukawa, R.; Takao, M.; Uye, S.I. Marine artificial structures as amplifiers of Aurelia aurita
825		s.l. blooms: A case study of a newly installed floating pier. J. Oceanogr. 2014, 70, 447–455.
826	102.	De Mesel, I.; Kerckhof, F.; Norro, A.; Rumes, B.; Degraer, S. Succession and seasonal dynamics of the
827		epifauna community on offshore wind farm foundations and their role as stepping stones for non-
828		indigenous species. Hydrobiologia 2015, 756, 37–50.

- Tidbury, H.; Taylor, N.; van der Molen, J.; Garcia, L.; Posen, P.; Gill, A.; Lincoln, S.; Judd, A.; Hyder, K.
  Social network analysis as a tool for marine spatial planning: Impacts of decommissioning on connectivity in the North Sea. *J. Appl. Ecol.* 2020, *57*, 566–577.
- 104. Degraer, S.; Brabant, R. Offshore wind farms in the Belgian part of the North Sea: State of the art after
  two years of environmental monitoring. *R. Belgian Inst. Nat. Sci. Manag. Unit North Sea Math. Model. Mar. Ecosyst. Manag. unit.* 2009, 287.
- 835 105. Southward, A.J.; Hawkins, S.J.; Burrows, M.T. Seventy years' observations of changes in distribution
  836 and abundance of zooplankton and intertidal organisms in the western English Channel in relation to
  837 rising sea temperature. *J. Therm. Biol.* 1995, 20, 127–155.
- 838 106. Stachowicz, J.J.; Terwin, J.R.; Whitlatch, R.B.; Osman, R.W. Linking climate change and biological
  839 invasions: Ocean warming facilitates nonindigenous species invasions. *Proc. Natl. Acad. Sci. U. S. A.* 2002,
  840 99, 15497–15500.
- Walther, G.R.; Roques, A.; Hulme, P.E.; Sykes, M.T.; Pyšek, P.; Kühn, I.; Zobel, M.; Bacher, S.; BottaDukát, Z.; Bugmann, H.; et al. Alien species in a warmer world: risks and opportunities. *Trends Ecol. Evol.* 2009, 24, 686–693.
- 108. Coolen, J.W.P.; Lengkeek, W.; Degraer, S.; Kerckhof, F.; Kirkwood, R.J.; Lindeboom, H.J. Distribution of
  the invasive Caprella mutica Schurin, 1935 and native Caprella linearis (Linnaeus, 1767) on artificial hard
  substrates in the North Sea: Separation by habitat. *Aquat. Invasions* 2016, *11*, 437–449.
- 847 109. Leonhard, S.B.; Pedersen, J. Benthic Communities at Horns Rev Before, During and After Construction
  848 of Horns Rev Offshore Wind Farm Annual Report 2005. *Vattenfall* 2006, 134.
- 849 110. Adams, T.P.; Miller, R.G.; Aleynik, D.; Burrows, M.T. Offshore marine renewable energy devices as
  850 stepping stones across biogeographical boundaries. *J. Appl. Ecol.* 2014, *51*, 330–338.
- 851 111. Maar, M.; Bolding, K.; Petersen, J.K.; Hansen, J.L.S.; Timmermann, K. Local effects of blue mussels
  852 around turbine foundations in an ecosystem model of Nysted off-shore wind farm, Denmark. *J. Sea Res.*853 2009, 62, 159–174.
- 854 112. Wilson, J.C.; Elliott, M. The habitat-creation potential of offshore wind farms. *Wind Energy* 2009, *12*, 203–
  855 212.
- Reubens, J.T.; De Rijcke, M.; Degraer, S.; Vincx, M. Diel variation in feeding and movement patterns of
  juvenile Atlantic cod at offshore wind farms. *J. Sea Res.* 2014, *85*, 214–221.
- Kerckhof, F.; Rumes, B.; Jacques, T.; Degraer, S.; Norro, A. Early development of the subtidal marine
  biofouling on a concrete offshore windmill foundation on the Thornton Bank (southern North Sea): First
  monitoring results. *Underw. Technol.* 2010, 29, 137–149.
- 861 115. Hixon, M.A.; Brostoff, W.N. Substrate characteristics, fish grazing, and epibenthic assemblages off
  862 Hawaii. *Bull. Mar. Sci.* 1985, 37, 200–213.
- 863 116. Beisiegel, K.; Tauber, F.; Gogina, M.; Zettler, M.L.; Darr, A. The potential exceptional role of a small Baltic
  864 boulder reef as a solitary habitat in a sea of mud. *Aquat. Conserv. Mar. Freshw. Ecosyst.* 2019, 29, 321–328.
- 865 117. Bortone, S.A.; Cody, R.P.; Turpin, R.K.; Bundrick, C.M. The impact of artificial-reef fish assemblages on
  866 their potential forage area. *Ital. J. Zool.* 1998, 65, 265–267.
- 867 118. Rosemond, R.C.; Paxton, A.B.; Lemoine, H.R.; Fegley, S.R.; Peterson, C.H. Fish use of reef structures and
  868 adjacent sand flats: implications for selecting minimum buffer zones between new artificial reefs and
  869 existing reefs. *Mar. Ecol. Prog. Ser.* 2018, 587, 187–199.
- 870 119. Shen, W.Z.; Mikkelsen, R.F. Study on wind turbine arrangement for offshore wind farms. In Proceedings
  871 of the ICOWEOE-2011 (Vol. Paper 05); 2011.

J. Mar. Sci. Eng. 2020, 8, x FOR PEER REVIEW

- Leitão, F.; Santos, M.N.; Erzini, K.; Monteiro, C.C. Fish assemblages and rapid colonization after
  enlargement of an artificial reef off the Algarve coast (Southern Portugal). *Mar. Ecol.* 2008, 29, 435–448.
- 874 121. Ambrose, R.F.; Swarbrick, S.L. Comparison of fish assemblages on artificial and natural reefs off the
  875 coast of southern California. *Bull. Mar. Sci.* 1989, 44, 718–733.
- 876 122. Champion, C.; Suthers, I.M.; Smith, J.A. Zooplanktivory is a key process for fish production on a coastal
  877 artificial reef. *Mar. Ecol. Prog. Ser.* 2015, 541, 1–14.
- 878 123. Ogawa, S.; Takeuchi, R.; Hattori, H. An estimate for the optimum size of artificial reefs. *Bull. Japanese*879 Soc. Fish. Oceanogr. 1977, 30, 39–45.
- 880 124. Bohnsack, J.A.; Sutherland, D.L. Artificial Reef Research: A Review with Recommendations for Future
  881 Priorities. *Bull. Mar. Sci.* 1985, 37, 11–39.
- 125. Charbonnel, E.; Serre, C.; Ruitton, S.; Harmelin, J.G.; Jensen, A. Effects of increased habitat complexity
  on fish assemblages associated with large artificial reef units (French Mediterranean coast). *ICES J. Mar. Sci.* 2002, *59*, 208–213.
- 885 126. Hixon, M.A.; Beets, J.P. Predation, Prey Refuges, and the Structure of Coral-Reef Fish Assemblages. *Ecol.*886 *Monogr.* 1993, 63, 77–101.
- 127. Crowder, L.B.; Cooper, W.E. Habitat Structural Complexity and the Interaction Between Bluegills and
  Their Prey. *Ecology* 1982, 63, 1802–1813.
- Bavis, K.L.; Coleman, M.A.; Connell, S.D.; Russell, B.D.; Gillanders, B.M.; Kelaher, B.P. Ecological
  performance of construction materials subject to ocean climate change. *Mar. Environ. Res.* 2017, 131, 177–
  182.
- Walker, B.K.; Henderson, B.; Spieler, R.E.; Beach, M.; Walker, B.K.; Henderson, B.; Spieler, R.E. Fish
  Assemblages Associated with Artificial Reefs of Concrete Aggregates or Quarry Stone Offshore Miami
  Beach, Florida, USA. *Aquat. Living Resour.* 2002, *15*, 95–105.
- 895 130. Carr, M.H.; Hixon, M.A. Artificial Reefs: The Importance of Comparisons with Natural Reefs. *Fisheries*896 1997, 22, 28–33.
- 897 131. Brandt, M.J.; Diederichs, A.; Betke, K.; Nehls, G. Responses of harbour porpoises to pile driving at the
  898 Horns Rev II offshore wind farm in the Danish North Sea. *Mar. Ecol. Prog. Ser.* 2011, 421, 205–216.
- 899 132. Teilmann, J.; Carstensen, J. Negative long term effects on harbour porpoises from a large scale offshore
  900 wind farm in the Baltic Evidence of slow recovery. *Environ. Res. Lett.* 2012, 7.
- 901 133. Furness, R.W.; Wade, H.M.; Masden, E.A. Assessing vulnerability of marine bird populations to offshore
  902 wind farms. J. Environ. Manage. 2013, 119, 56–66.
- 903 134. Russell, D.J.F.; Brasseur, S.M.J.M.; Thompson, D.; Hastie, G.D.; Janik, V.M.; Aarts, G.; McClintock, B.T.;
  904 Matthiopoulos, J.; Moss, S.E.W.; McConnell, B. Marine mammals trace anthropogenic structures at sea.
  905 *Curr. Biol.* 2014, 24, 638–639.
- 906135.Spagnolo, A.; Fabi, G.; Manoukian, S.; Panfili, M. Benthic community settled on an artificial reef in the907Western Adriatic Sea (Italy). *Rapp. la Comm. Int. pour l'Exploration Sci. la Mer Médititerranée* 2004, 37, 552.
- 908 136. Verdura, J.; Sales, M.; Ballesteros, E.; Cefalì, M.E.; Cebrian, E. Restoration of a Canopy-Forming Alga
  909 Based on Recruitment Enhancement: Methods and Long-Term Success Assessment. *Front. Plant Sci.*910 2018, 9, 1–12.
- 911137.Fredriksen, S.; Filbee-Dexter, K.; Norderhaug, K.M.; Steen, H.; Bodvin, T.; Coleman, M.A.; Moy, F.;912Wernberg, T. Green gravel: a novel restoration tool to combat kelp forest decline. *Sci. Rep.* 2020, 10, 1–7.
- 913 138. Scheidat, M.; Tougaard, J.; Brasseur, S.; Carstensen, J.; Van Polanen Petel, T.; Teilmann, J.; Reijnders, P.
- 914 Harbour porpoises (Phocoena phocoena) and wind farms: A case study in the Dutch North Sea. *Environ*.

915		<i>Res. Lett.</i> <b>2011</b> , <i>6</i> .
916	139.	Hammill, M.O.; Stenson, G.B. Estimated Prey Consumption by Harp seals (Phoca groenlandica),
917		Hooded seals (Cystophora cristata), Grey seals (Halichoerus grypus) and Harbour seals (Phoca vitulina)
918		in Atlantic Canada. J. Northw. Atl. Fish. Sci. <b>1999</b> , 26, 1–23.
919	140.	Chouinard, G.A.; Swain, D.P.; Hammill, M.O.; Poirier, G.A. Covariation between grey seal (Halichoerus
920		grypus) abundance and natural mortality of cod (Gadus morhua) in the southern Gulf of St . Lawrence.
921		<i>Can. J. Fish. Aquat. Sci.</i> <b>2005</b> , <i>62</i> , 1991–2000.
922	141.	Harlin, M.M.; Lindbergh, J.M. Selection of substrata by seaweeds: optimal surface relief. Mar. Biol. 1977,
923		40, 33–40.
924	142.	Beserra Azevedo, F.B.; Carloni, G.G.; Vercosa Carvalheira, L. Colonization of benthic organisms of
925		different artificial substratum in Ilha Grande Bay, Rio de Janeiro. Brazilian Arch. Biol. Technol. 2006, 49,
926		263–275.
927	143.	Rilov, G.; Benayahu, Y. Vertical artificial structures as an alternative habitat for coral reef fishes in
928		disturbed environments. Mar. Environ. Res. 1998, 45, 431-451.
929	144.	De La Fuente, G.; Chiantore, M.; Asnaghi, V.; Kaleb, S.; Falace, A. First ex situ outplanting of the habitat-
930		forming seaweed Cystoseira amentacea var. stricta from a restoration perspective. PeerJ 2019, 7, e7290.
931	145.	Quaas, Z.; Harasti, D.; Gaston, T.F.; Platell, M.E.; Fulton, C.J. Influence of habitat condition on shallow
932		rocky reef fish community structure around islands and headlands of a temperate marine protected area.
933		2019, 626, 1–13.
934	146.	Jensen, A.; Wickins, J.; Bannister, C. The Potential Use of Artificial Reefs to Enhance Lobster Habitat. In
935		Artificial Reefs in European Seas. Springer, Dordrecht; 2000.
936	147.	Spanier, E. What Are the Characteristics of a Good Artificial Reef for Lobsters? Crustaceana 1994, 67, 173–
937		186.
938	148.	Cobb, J.S. The Shelter-Related Behavior of the Lobster, Homarus Americanus. Ecology 1971, 52, 108–115.
939	149.	Barry, J.; Wickins, J.F. A model for the number and sizes of crevices that can be seen on the exposed
940		surface of submerged rock reefs. Environmetrics 1992, 3, 55–69.
941	150.	Saunders, D.A.; Hobbs, R.J.; Margules, C.R. Biological Consequences of Ecosystem Fragmentation - A
942		Review. Conserv. Biol. 1991, 5, 18–32.
943	151.	Ries, L.; Fletcher, R.J.; Battin, J.; Sisk, T.D. Ecological responses to habitat edges: Mechanisms, models,
944		and variability explained. Annu. Rev. Ecol. Evol. Syst. 2004, 35, 491–522.
945	152.	Jelbart, J.E.; Ross, P.M.; Connolly, R.M. Edge effects and patch size in seagrass landscapes: an
946		experimental test using fish. Mar. Ecol. Prog. Ser. 2006, 319, 93–102.
947	153.	Selgrath, J.C.; Hovel, K.A.; Wahle, R.A. Effects of habitat edges on American lobster abundance and
948		survival. J. Exp. Mar. Bio. Ecol. 2007, 353, 253–264.
949	154.	Smith, T.M.; Hindell, J.S.; Jenkins, G.P.; Connolly, R.M. Edge effects on fish associated with seagrass and
950		sand patches. Mar. Ecol. Prog. Ser. 2008, 359, 203–213.
951	155.	Malcolm, H.A.; Jordan, A.; Smith, S.D.A. Biogeographical and cross-shelf patterns of reef fish
952		assemblages in a transition zone. Mar. Biodivers. 2010, 40, 181–193.
953	156.	McClanahan, T.R.; Graham, N.A.J.; Calnan, J.M.; MacNeil, M.A. Toward Pristine Biomass: Reef Fish
954		Recovery in Coral Reef Marine Protected Areas in Kenya. Ecol. Appl. 2007, 17, 1055–1067.
955	157.	Fukami, T.; Nakajima, M. Community assembly: Alternative stable states or alternative transient states?
956		<i>Ecol. Lett.</i> <b>2011</b> , <i>14</i> , 973–984.
957	158.	Kayal, M.; Lenihan, H.S.; Brooks, A.J.; Holbrook, S.J.; Schmitt, R.J.; Kendall, B.E. Predicting coral

958		community recovery using multi-species population dynamics models. Ecol. Lett. 2019, 22, 605–615.
959	159.	Schoonees, T.; Gijón Mancheño, A.; Scheres, B.; Bouma, T.J.; Silva, R.; Schlurmann, T.; Schüttrumpf, H.
960		Hard Structures for Coastal Protection, Towards Greener Designs. Estuaries and Coasts 2019, 42, 1709-
961		1729.
962	160.	Hammar, L.; Perry, D.; Gullström, M. Offshore Wind Power for Marine Conservation. Open J. Mar. Sci.
963		<b>2016</b> , <i>06</i> , 66–78.
964	161.	Díaz, D.; Mallol, S.; Parma, A.M.; Goñi, R. A 25-year marine reserve as proxy for the unfished condition
965		of an exploited species. <i>Biol. Conserv.</i> <b>2016</b> , 203, 97–107.
966	162.	Kough, A.S.; Belak, C.A.; Paris, C.B.; Lundy, A.; Cronin, H.; Gnanalingam, G.; Hagedorn, S.; Skubel, R.;
967		Weiler, A.C.; Stoner, A.W. Ecological spillover from a marine protected area replenishes an over-
968		exploited population across an island chain. Conserv. Sci. Pract. 2019, 1, e17.
969	163.	Burkhard, B.; Gee, K. Establishing the resilience of a coastal-marine social-ecological system to the
970		installation of offshore wind farms. Ecol. Soc. 2012, 17, 42.
971	164.	Hewitt, J.E.; Thrush, S.F. Monitoring for tipping points in the marine environment. J. Environ. Manage.
972		<b>2019</b> , 234, 131–137.
973	165.	Mavraki, N.; Degraer, S.; Moens, T.; Vanaverbeke, J. Functional differences in trophic structure of
974		offshore wind farm communities: A stable isotope study. Mar. Environ. Res. 2020, 157, 104868.
975	166.	Claisse, J.T.; Pondella, D.J.; Love, M.; Zahn, L.A.; Williams, C.M.; Williams, J.P.; Bull, A.S. Oil platforms
976		off California are among the most productive marine fish habitats globally. Proc. Natl. Acad. Sci. 2014,
977		111, 15462–15467.
978	167.	Ajemian, M.J.; Wetz, J.J.; Shipley-Lozano, B.; Dale Shively, J.; Stunz, G.W. An analysis of artificial reef
979		fish community structure along the northwestern gulf of Mexico Shelf: Potential impacts of "rigs-to-
980		reefs" Programs. PLoS One 2015, 10, 1–22.
981	168.	Todd, V.L.G.; Lavallin, E.W.; Macreadie, P.I. Quantitative analysis of fish and invertebrate assemblage
982		dynamics in association with a North Sea oil and gas installation complex. Mar. Environ. Res. 2018, 69-
983		79.
984	169.	McLean, D.L.; Taylor, M.D.; Giraldo, A.; Partridge, J.C. An assessment of fish and marine growth
985		associated with an oil and gas platform jacket using an augmented remotely operated vehicle. Cont. Shelf
986		<i>Res.</i> <b>2019</b> , <i>179</i> , 66–84.
987	170.	Inger, R.; Attrill, M.J.; Bearhop, S.; Broderick, A.C.; James Grecian, W.; Hodgson, D.J.; Mills, C.; Sheehan,
988		E.; Votier, S.C.; Witt, M.J.; et al. Marine renewable energy: Potential benefits to biodiversity? An urgent
989		call for research. J. Appl. Ecol. 2009, 46, 1145–1153.
990	171.	Dons, S.; Jensen, D.J.; Struve, A.; Nielsen, B. Vindeby Havmøllepark. DONG Energy 2016, 1–61.
991	172.	Bull, A.S.; Love, M.S. Worldwide oil and gas platform decommissioning: A review of practices and
992		reefing options. Ocean Coast. Manag. 2019, 168, 274–306.
993	173.	Fowler, A.M.; Jørgensen, A.M.; Svendsen, J.C.; Macreadie, P.I.; Jones, D.O.B.; Boon, A.R.; Booth, D.J.;
994		Brabant, R.; Callahan, E.; Claisse, J.T.; et al. Environmental benefits of leaving offshore infrastructure in
995		the ocean. Front. Ecol. Environ. 2018, 16, 571–578.
996	174.	Smyth, K.; Christie, N.; Burdon, D.; Atkins, J.P.; Barnes, R.; Elliott, M. Renewables-to-reefs? -
997		Decommissioning options for the offshore wind power industry. Mar. Pollut. Bull. 2015, 90, 247–258.
998	175.	Topham, E.; McMillan, D. Sustainable decommissioning of an offshore wind farm. Renew. Energy 2016,
999		102, 470–480.
1000		

J. Mar. Sci. Eng. 2020, 8, x FOR PEER REVIEW

![](_page_29_Picture_1.jpeg)

© 2020 by the authors. Submitted for possible open access publication under the terms and conditions of the Creative Commons Attribution (CC BY) license (http://creativecommons.org/licenses/by/4.0/).

1001