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1 **Designing spawning closures can be complicated: experience from cod in the Baltic Sea**

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16

17 **Abstract**

18 Fisheries management measures often include spatio-temporal closures during the spawning
19 period of the fish with an overarching aim of improving the stock status. The different
20 mechanisms how a spawning closure potentially can influence the stock are often not
21 explicitly considered when designing such closures. In this paper, we review and synthesize
22 the available data and knowledge on potential effects of the implemented spawning closures
23 on cod in the Baltic Sea. The Baltic cod example represents a relatively data rich case, which
24 allows demonstrating how a closure might affect different parameters of stock status via
25 different mechanisms, including potential unintended negative effects. We conclude that
26 designing relatively small area closures appropriately is highly complex and data demanding,
27 and may involve tradeoffs between positive and negative impacts on the stock. Seasonal

28 closures covering most of the stock distribution during the spawning time are more robust to
29 data limitations, and less likely to be counterproductive if suboptimally designed.

30

31 Key words: spawning closure, fisheries management, recruitment, stock structure, Baltic cod

32

33 **1. Introduction**

34 Fisheries management measures frequently include spatio-temporal closures as a supplement
35 to catch or effort limitations and gear regulations (Murawski et al., 2000; Roberts et al.,
36 2005). The Marine Protected Area (MPA) type of closures, where fishing activities are
37 restricted or forbidden all year round generally have a wider aim of preserving biodiversity
38 and meeting various nature conservation objectives (Halpern, 2003; Lester et al., 2009). To
39 enhance a particular fish stock, seasonal closures during the spawning time of the population
40 in concern are often used, the so-called spawning closures. There are numerous studies and
41 ongoing debates on the effects of MPAs (e.g., Pendleton et al., 2017), while spawning
42 closures and related benefits to fish stocks have received comparatively less attention in the
43 literature so far.

44
45 Spawning closures can reduce overall fishing mortality of target species, especially when
46 these are forming large spawning aggregations (Russell et al., 2012; Grüss et al., 2014). In
47 case of no targeted fishery for a particular stock, spawning closures could reduce bycatch
48 (O’Keefe et al., 2014). For fish stocks, where the overall fishing mortality is regulated by
49 other measures, such as total allowable catch (TAC), the benefits from spawning closures are
50 suggested to include greater reproductive output, positive effects on stock structure and
51 reduced evolutionary effects of fishing (van Overzee and Rijnsdorp, 2015 and references
52 therein). However, such effects are often difficult to demonstrate convincingly, and are
53 possibly case specific. Therefore, case specific monitoring and evaluation of the established
54 closures are needed as part of a management process to allow for possible adaptations as well
55 as communication of their effectiveness (Pomeroy et al., 2005).

56
57 In this paper, we synthesize the state-of-the-art scientific knowledge on possible effects of the
58 established spawning closures on cod in the Baltic Sea. We are specifically interested in
59 distinguishing between different mechanisms how a spawning closure potentially could affect
60 the cod. The Baltic cod provides a relatively data rich example, allowing to identify tradeoffs
61 that may be involved when a spawning closure affects various aspects of the stock status via
62 different mechanisms. Furthermore, different types of spawning closures are implemented in
63 the Baltic Sea. Therefore, the Baltic cod example allows us addressing their relative
64 advantages, especially in situations when scientific knowledge is limited, which may partly be

65 the case even for stocks generally considered as data-rich. The present study is intended to
66 guide future monitoring and research efforts as well as inform debates and decisions on the
67 design and use of spawning closures in fisheries management in the Baltic Sea and elsewhere.
68

69 **2. Background of cod stocks and fisheries management measures in the Baltic Sea**

70 There are two genetically distinct populations of cod (*Gadus morhua*) in the Baltic Sea,
71 eastern (EB) and western (WB) (Nielsen et al., 2003; 2005). Since 2004, these are managed in
72 two separate areas, corresponding to the International Council for the Exploration of the Sea
73 (ICES) Subdivisions (SD) 25–32 and 22–24, respectively (Fig. 1). Both the EB and WB cod
74 occur in SD 24 (Hüssy et al., 2016). Cod in the Baltic Sea is a target species for fisheries with
75 demersal trawls and gill-nets, and managed by total allowable catch (TAC). Technical
76 measures including various regulations on fishing gears and a minimum conservation
77 reference size are in place.
78

79 The EB cod biomass was record high in the 1970s–1980s, after which it substantially declined
80 due to a combination of unfavourable environmental conditions for reproduction and a high
81 fishing pressure (Bagge et al., 1994). In recent decade, fishing mortality is estimated
82 substantially lower, though is presently considered still above the management target and the
83 stock size has declined after a small recovery observed in the late 2000s (ICES, 2018a). The
84 fishing mortality of WB cod is estimated well above the management target in the entire time
85 series and the spawner biomass has been below the reference level since 2009, though is
86 predicted to increase in coming years (ICES, 2018a).
87

88 In 1995, a few years after the major decline of the EB cod stock in the late 1980s, a
89 prohibition to fish cod from June to August in the EB Sea was introduced (IBSFC, 1994). In
90 1998, additionally an area closure for all fishing activities was established in the essential cod
91 spawning area in the Bornholm Basin (BB) (Fig. 1) in the period from mid-May to the end of
92 August. In the following decade, the timing of these closures as well as the size and shape of
93 the area closure in the BB were modified several times. These changes were not based on a
94 well-defined biological mechanism and it is unclear how these closures were expected to
95 contribute to rebuilding of the stock (Suuronen et al., 2010). In 2006, additional area closures
96 were established in the Gdansk Deep (GD) and Gotland Basin (GB) (Fig. 1). In the first EU

97 management plan for cod (EU, 2007), effort limitation in the EB Sea included a prohibition of
 98 all cod fisheries from July 1 to August 31, from here on referred to as a seasonal closure.
 99 Additionally, the plan included a prohibition of any fishing activities in the three designated
 100 areas (Fig. 1) from May 1 to October 31, which are referred to as area closures. In the WB
 101 Sea, a seasonal closure was in effect from April 1 to 30. In the new EU Baltic multi-annual
 102 management plan (Baltic MAP) (EU, 2016), seasonal closures in both the EB and WB Sea
 103 were lifted, while the area closures in the EB Sea were maintained (Table 1). According to the
 104 Baltic MAP, supplementary measures need to be applied when the stocks are in poor state,
 105 which can include spawning closures (EU, 2016). In practice, a seasonal closure from
 106 February 15 to March 31 was implemented in the WB Sea in 2016, further extended to
 107 February 1– March 31 in 2017–2018 (Table 1). In the EB Sea, a seasonal closure in SDs 25–
 108 26 was re-introduced for 2018. There are exemptions in place for vessels <15 m (in 2017 in
 109 WB) or <12 m (in 2018 in EB and WB) if they can demonstrate that they do not fish for cod
 110 in areas deeper than 20 m.

111

112 Table 1. Overview of the seasonal (*SC*) and area closures (*AC*) for cod fisheries enforced in
 113 the Eastern Baltic (EB) and Western Baltic (WB) management areas under the present EU
 114 management plan (EU, 2016). See Fig. 1 for location of the area closures in Bornholm Basin
 115 (BB), Gdansk Deep (GD) and Gotland Basin (GB).

116

| EB management area | WB management area |
|---|---|
| 2016–2018: | 2016: |
| <i>AC</i> : May 1–October 31 (BB, GD, GB) | <i>SC</i> : February 15–March 31 (SD 22–24) |
| 2018: | 2017–2018: |
| <i>SC</i> : July 1–August 31 (SDs 25–26) | <i>SC</i> : February 1 –March 31 (SD 22–24) |

117

118 **3. Material and Methods**

119

120 **3.1 Defining potential objectives for the cod spawning closures in the Baltic Sea**

121 The overarching aim of the cod spawning closures in the Baltic Sea is improving the stock
 122 status. The legislations do not specify further, which parameters of stock status the closures
 123 are intended to improve, and through which mechanisms. According to the literature, potential

124 benefits of spawning closures as a supplementary management measure can include greater
125 reproductive output, positive effects on stock structure, reduced evolutionary effects of fishing
126 and reduced impact on spawning habitat (e.g. van Overzee and Rijnsdorp, 2015 and
127 references therein). We focused our analyses on the potential effects of the spawning closures
128 on cod recruitment, distinguishing between three different mechanisms. These included direct
129 effects of the closures on:

- 130 i) the quantity and quality of egg production by ensuring undisturbed spawning
131 activity;
- 132 ii) preserving the spawners whose offspring have a higher survival probability;
- 133 iii) increasing the proportion of larger/older individuals in the stock.

134 Further explanation of how the cod recruitment could benefit from these potential effects of
135 spawning closures is provided in section 4. We focus on these selected potential effects of the
136 closures because of their relevance for the Baltic cod and the availability of scientific
137 information, which allows for their relatively in-depth consideration. Hence, improving cod
138 recruitment through the three mechanisms described above was used as a specific objective
139 for the Baltic cod spawning closures in this study.

140

141 Concerning other potential benefits of spawning closures suggested in the literature, such as
142 avoiding evolutionary change towards earlier maturation or reducing the risk of losing genetic
143 diversity (van Overzee and Rijnsdorp, 2015), little information is available for Baltic cod.
144 Size at maturation of the EB cod has substantially declined from the late 1990s to 2010s
145 (Köster et al., 2017), when spawning closures have been enforced. This change is not fully
146 understood, though is not considered to be connected to the spawning closures, which should
147 have an opposite effect, i.e. preventing earlier maturation. Both the EB and WB cod have
148 different spawning locations in the Baltic Sea, however genetic differences between those
149 within a stock have not been demonstrated (Nielsen et al., 2003; 2005; Poćwierz-Kotus et al.,
150 2015). Spawning site fidelity on a finer spatial scale occurs in some cod stocks (e.g., Zemekis
151 et al., 2014). Such aspects as well as related potential impacts of spawning closures are
152 largely unknown for Baltic cod.

153

154

155

156 **3.2 Evaluation approach**

157 The effects of spawning closures on wild fish stocks are generally very difficult to
158 demonstrate or quantify (e.g., Arendse et al., 2007; Clarke et al., 2015). This is because of a
159 large number of factors and processes influencing fish stock dynamics, for example
160 recruitment, including that of cod in the Baltic Sea (Köster et al., 2017). The approach of
161 looking at stock parameters before and after the implementation of closures is frequently
162 applied (e.g., Russ et al., 2004; Torres-Irineo et al., 2011), however is challenged by other
163 factors influencing the stock dynamics at the same time (Davies et al., 2017). Comparable
164 control areas or seasons are often not available or meaningful, which is also the case for cod
165 in the Baltic Sea.

166

167 Recognizing this general difficulty in assessing the *realized* effects of spawning closures, we
168 took a different approach in this study, where we instead evaluated their *potential* effects. In
169 this approach, we focused on identifying whether or not there is an overlap between the
170 closure and the stock component intended to be protected, in time and space. If such overlap
171 is lacking, the closure can impossibly be beneficial. If the overlap is present, this implies that
172 the closure can *potentially* contribute to improving the stock status through a certain
173 mechanism. Corresponding to the three potential effects of the spawning closures we address
174 in this paper, we evaluated whether there is an overlap between

- 175 • the closures and the cod spawning activity,
- 176 • the closures and the spawners whose offspring has a higher survival probability,
- 177 • the closures and the largest/oldest cod.

178

179 We evaluated potential both positive and negative effects of the closures. The latter were
180 associated with the possibility of fishing effort reallocation during the time of closure,
181 compromising some aspects of the stock status. Both the area and seasonal closures enforced
182 under the present EU management plan (EU, 2016) were considered (Table 1), separately for
183 the EB and WB management areas. Our study does not cover the economic and social
184 implications, or other possible ecological effects of the closures on other species or habitats.

185

186 **3.3 Data**

187 A number of scientific publications over the past decades have addressed cod recruitment in
188 the Baltic Sea, including aspects relevant for evaluating the spawning closures. This paper
189 presents a synthesis of these findings in the context of the specific questions regarding the
190 potential effects of the spawning closures, defined in the sections above. This review is
191 supplemented by additional analyses, using data on egg abundances from ichthyoplankton
192 surveys and cod catch information from the Baltic fish stock Assessment Working Group in
193 ICES (ICES, 2018a). For the EB cod, part of the present synthesis was conducted in
194 connection with a workshop in ICES (2018b).

195

196 **4. Results**

197 **4.1 Undisturbed spawning**

198 Fishing activities may adversely affect the spawning fish and subsequently the quantity or
199 quality of the offspring (Sadovy de Mitcheson and Erisman, 2012). The disturbance can take
200 place via a number of mechanisms, including noise from fishing and interruption of
201 spawning, causing physiological stress response in the fish and disturbance of natural
202 spawning behaviour (van Overzee and Rijnsdorp, 2015). The effect of spawning disturbance
203 on reproductive output is very difficult to demonstrate or quantify for wild fish stocks, and no
204 such investigations are available for Baltic cod. A pre-requisite for a closure to ensure
205 undisturbed spawning is a spatio-temporal overlap with spawning activity, which is the only
206 aspect in relation to spawning disturbance that can presently be evaluated for Baltic cod.

207

208 **EB:** In the EB Sea, there are historically three main cod spawning grounds, in deeper areas of
209 the Bornholm Basin (BB), Gdansk Deep (GD) and Gotland Basin (GB) (Fig. 1). Due to
210 reduced salinity and oxygen, conditions for cod egg survival in GD and GB have deteriorated
211 considerably since the mid-1980s (MacKenzie et al., 2000; Köster et al., 2009), and these
212 spawning areas have presently a limited contribution to cod recruitment (Plikshs et al., 2015;
213 Köster et al., 2017). Therefore, disturbance from fishing in these areas unlikely has a
214 measurable effect on the reproductive output of the stock. In the BB, i.e. presently the main
215 spawning ground for the EB cod, spawning is restricted to areas with water depth >60 m
216 (Wieland, 1988; Hinrichsen et al., 2007; Figure 2a). The horizontal distribution of eggs within
217 the spawning area varies between years as well as within a year (Hinrichsen et al., 2007;

218 Neumann et al., 2014; Fig. S1). Thus, the area closure in the BB covers varying proportions
219 of the spawning activity, while not covering the entire spawning area.

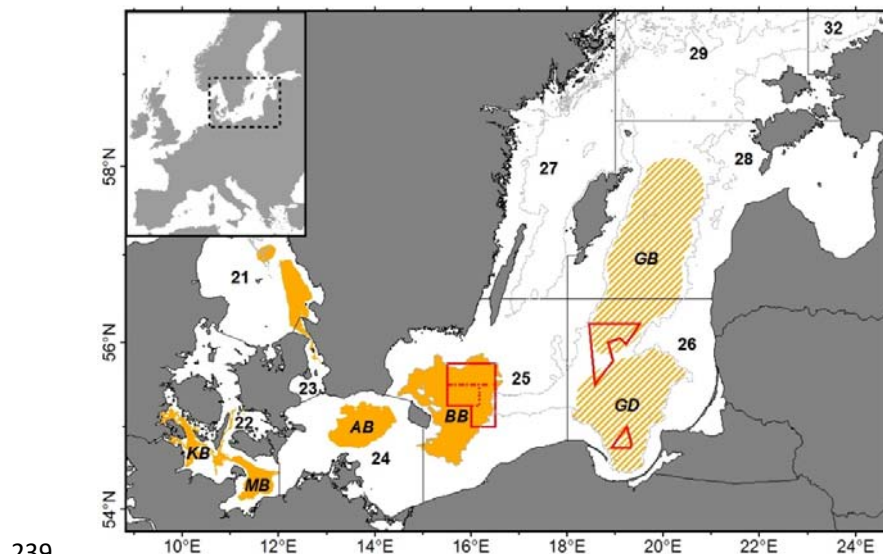
220

221 The spawning of the EB cod starts in February–March and last until October–November
222 (Wieland et al., 2000). Peak spawning occurred between the end of April and mid-June in the
223 1970s and 1980s (MacKenzie et al., 1996), and gradually changed to the second half of July
224 during the 1990s (Wieland et al., 2000). In the late 2000s, the main spawning expanded to
225 spring, covering a 4 months period from May to August (Neumann et al., 2014; Köster et al.,
226 2017). In most years since 2010, highest egg abundances have been recorded in June (ICES,
227 2018*b*), which is not covered by the seasonal closure enforced in July–August.

228

229 **WB:** WB cod spawn in areas deeper than 20 m in Kiel Bay, Mecklenburg Bay, and the
230 Danish Belts (SD 22; Fig. 1) (Bleil and Oeberst, 2002; Hüsey, 2011). Cod spawning activities
231 are also recorded in the Sound (SD 23) (Hüsey, 2011). Both the WB and EB cod spawn in the
232 Arkona Basin (AB) (in SD 24) (Bleil et al., 2009; Hüsey, 2011), in areas with water depth >40
233 m (Fig. 1) (Bleil and Oeberst, 2002). The main spawning period of female repeat spawning
234 cod in SD 22 (the core spawning area of the WB stock) is from mid-February to early April,
235 matching the timing of the seasonal closure implemented in this area in 2016–2018 (STECF,
236 2016). In the AB, the timing of cod spawning reflects a combination of different spawning
237 times of the WB and EB stock, which are mixed in the area.

238



240 Fig. 1. Cod spawning areas (filled areas on the map) in the Baltic Sea in the Sound (23), Kiel
241 Bay (KB), Mecklenburg Bay (MB), Arkona Basin (AB) and Bornholm Basin (BB); the
242 shaded areas in Gdansk Deep (GD) and Gotland Basin (GB) indicate spawning areas that
243 have had limited contribution to cod recruitment since the mid-1980s (modified from Bagge
244 et al., 1994 and Hüsey, 2011) . The bold lines show the borders of the present area closures in
245 the eastern Baltic Sea (EU, 2016), with the broken lines indicating historical borders for the
246 closure in BB. The numbers and thin lines depict the ICES Subdivisions (SD).

247

248 **4.2 Early life stage survival**

249 The effect of undisturbed spawning on reproductive output needs to be seen in conjunction
250 with survival probability of the offspring, as high egg production alone is not sufficient for
251 enhancing recruitment. If the survival of early life stages is variable in time and space,
252 spawning closures could potentially enhance the recruitment by protecting those spawners
253 whose offspring has a higher survival probability.

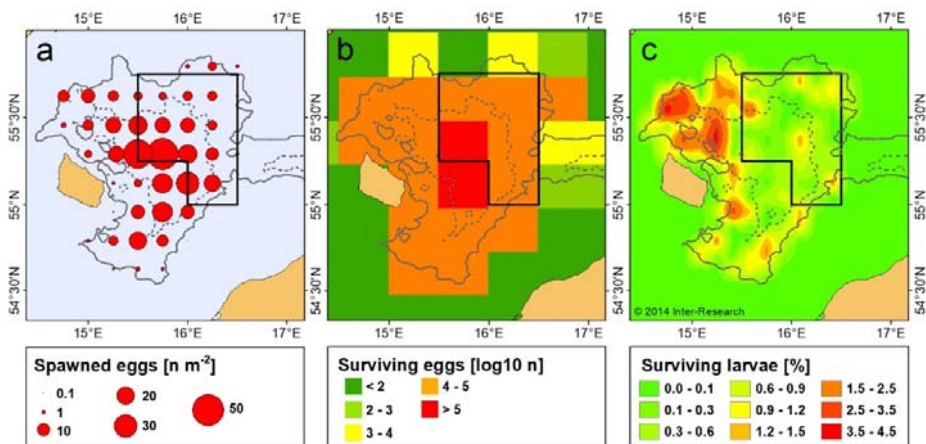
254

255 **EB:** The survival of cod eggs in the GD and GB has been low since the 1990s, due to
256 unfavourable salinity and oxygen conditions (Köster et al., 2017). Therefore, we focus this
257 section on the BB, where hydrographic conditions generally support relatively high egg
258 survival in the months and areas covered by both the seasonal and area closure. However, the
259 closures do not cover the entire window of high egg survival probability. Modelling of egg
260 survival has shown that, on average, the highest concentrations of surviving eggs originate
261 from the center of the basin (Fig. 2b), with the highest survival probability in May–August
262 (Hinrichsen et al. 2016a).

263

264 The survival of first feeding larvae, which is critical for determining recruitment success of
265 the EB cod (Köster et al., 2005; Huwer et al., 2011) is largely affected by prey availability, in
266 particular *Pseudocalanus nauplii* (Hinrichsen et al., 2002; Voss et al., 2003). A modelling
267 study suggests that, at favourable feeding conditions, larval survival is highest in the center of
268 the basin (Hinrichsen et al., 2002), i.e. including the area closure in the BB. In contrast, in
269 years with low *Pseudocalanus* abundance, larvae have better feeding opportunities and a
270 higher survival probability if retained at the slopes of the basin or transported into shallower
271 coastal regions, i.e. outside the area closure (Hinrichsen et al., 2002). This modelling result

272 was confirmed by an empirical study back-tracking hatch positions of pelagic juvenile
 273 survivors in year 2000, which revealed that the vast majority of these juveniles originated
 274 from the slopes of the BB, i.e. outside the closed area (Huwer et al., 2014; Fig. 2c). It should
 275 be noted that the absolute numbers of recruits originating from different time windows or
 276 locations have not been quantified, as this would require more extensive and systematic
 277 sampling of juveniles than is currently the case.
 278



279
 280 Fig. 2. EB cod in the Bornholm Basin: (a) Horizontal distribution of newly spawned cod eggs
 281 (stage 1A), representing the location of spawning activity (average for the years 1989–2003,
 282 from Hinrichsen et al., 2007). (b) Modelled spatial origin of first-feeding yolk-sac larvae that
 283 have survived through the egg stage (average in 1971–2010, from Hinrichsen et al., 2016a).
 284 (c) Spatial origin of pelagic juveniles that have survived through the larval stage, (example for
 285 the year 2000, from Huwer et al., 2014). The grey lines show 60 m (solid) and 80 m (dashed)
 286 depth contours. The black solid box shows the extent of the present area closure (EU, 2016).

287
 288 **WB:** The survival of WB cod eggs is affected by temperature (see Hüsey 2011 for review),
 289 which is more likely to be below the optimum (in the range of 4–8 °C) early in the spawning
 290 season, including the time of the present closure. However, no clear relation between
 291 temperature and year-class strength suggests that other factors are likely more important for
 292 determining recruitment success (Hüsey et al., 2012). The egg quality (size, fertilization
 293 success) as well as the number of eggs per batch decrease towards the end of spawning of an
 294 individual cod (Bleil and Oeberst, 1998; Vallin and Nissling, 2000). Thus, the eggs with the
 295 best quality characteristics, which may influence their survival probability, are released within

296 the period covered by the closures in 2016–2018. In terms of the spatial coverage, a closure in
297 SD 24 has likely a limited contribution to enhancing the WB cod recruitment. This is due to
298 generally low egg survival in this area related to cold winter water filling the basin during the
299 main spawning time of the WB cod (Köster et al., 2017). During the 2000s, the environmental
300 conditions for reproduction in the AB were generally more favourable for the EB than for the
301 WB cod (Köster et al., 2017), with best spawning conditions irregularly occurring from mid-
302 May to end-June (STECF, 2010; Hüsey et al., 2016).

303

304 **4.3 Size/age structure of cod catches**

305

306 Larger female cod produce higher number of eggs, and there is evidence for increased
307 offspring quality with parent age or reproductive experience (e.g., Marteinsdottir and
308 Steinarsson, 1998; Trippel, 1998). Moreover, the on average larger eggs of larger cod are
309 neutrally buoyant at a lower salinity, implying that the eggs from older/larger EB cod have a
310 greater survival probability under low salinity conditions (Vallin and Nissling 2000;
311 Hinrichsen et al., 2016*b*). The size distribution of the EB cod stock has truncated in later
312 years, with very few larger individuals in the stock (ICES, 2018*a*). Thus, protecting the
313 remaining relatively larger cod may be essentially important. If fisheries catches during the
314 spawning time contain a larger fraction of older/larger individuals than in other times of the
315 year, a spawning closure could be beneficial for the recruitment success by preserving larger
316 cod. When investigating this hypothesis, we assumed that the total annual catch amount is
317 unchanged regardless of the closures and we only focused on the potential effect of the
318 closures on the size/age structure of the catch.

319

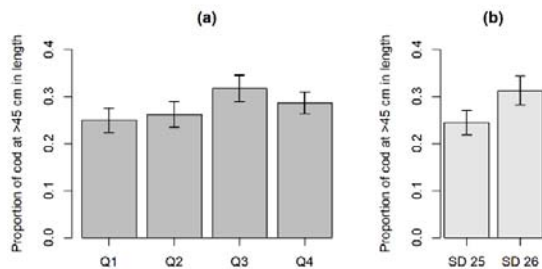
320 **EB:** Data on the amount and size structure of cod landings are available by quarter (q) and
321 ICES SDs in the Baltic Sea. Age information is not available for the EB cod for later years.
322 Seasonally, the fraction of the larger (>45 cm) EB cod in fisheries catch has generally been
323 highest in q3, although similar to q4 (Fig. 3*a*). However, the amount of cod landings in q3 has
324 been relatively low throughout the time series, i.e. including the years before the
325 implementation of the seasonal closure in 1995 (Fig. S2). This is likely due to low incentives
326 for cod fishery in this time of the year, regardless of the closure. Thus, the seasonal summer

327 closure may to some extent reduce the fraction of the largest individuals in fisheries catch,
328 though the effect is likely not substantial.

329

330 Spatially, the catches in SD 26 include a higher proportion of larger cod compared to SD 25
331 (Fig. 3*b*). There is some variability in this pattern between quarters (Fig. S3). However, the
332 fraction of the largest cod in fisheries catch in the main spawning area (in SD 25) does not
333 seem to be higher in any time of the year. The importance of SD 26 in total EB cod landings
334 has substantially increased in later years (Fig. S4), in line with the relatively larger cod found
335 in this area, making it more attractive for the fisheries. Finer scale spatial data on the
336 distribution of different size-groups of cod within a SD during the time of the closures are not
337 available. This is because research surveys are conducted in the 1st and 4th quarters, i.e.
338 outside the main spawning time, and data on size composition of the fisheries catch are only
339 available on a SD level.

340



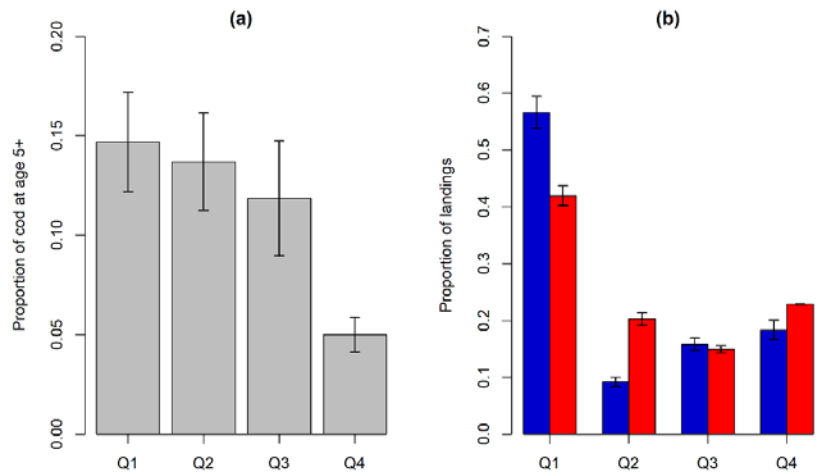
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342 Fig. 3. Average proportion of larger eastern Baltic cod (>45 cm in length) in the commercial
343 catch (a) by quarter and (b) in SDs 25 and 26 (annual average). The error bars represent
344 standard error of the mean, calculated over the years 2001–2017.

345

346 **WB:** The proportion of older WB cod in fisheries landings has been highest during the main
347 spawning season in q1, although similar to q2 (Fig. 4*a*). Thus, a fishery closure in q1 could
348 potentially reduce the proportion of the largest cod in fisheries catch. However, despite the
349 spawning closure being enforced in q1, a substantial share of the landings in 2016–2017 was
350 still taken in the period when the highest fraction of older cod occurs in the landings, i.e. in
351 the remainder of q1 and in q2 (Fig. 4*b*). This suggests that the spawning closure likely did not
352 reduce the proportion of larger cod in fisheries catch in those years substantially.

353



354

355 Figure 4. (a) Average proportion of cod at age 5+ in the landings of Western Baltic cod in
356 ICES Subdivisions 22–23, by quarter (Q). The error bars represent standard error of the mean,
357 calculated over the years 2005–2017. (b) Average relative distribution of cod landings
358 between quarters in 2005–2015 (blue bars) compared to 2016–2017 (red bars). The distinction
359 of the periods corresponds to revision of the time of the closure. Error bars show standard
360 error of the mean.

361

362 4.4 Possible tradeoffs between the different impacts of the spawning closures on cod

363

364 The present area closure in the main spawning ground (BB) of the EB cod allows part of the
365 stock to spawn undisturbed. However, this would not necessarily increase the recruitment, if
366 the offspring spawned outside the closure would have a higher survival probability due to
367 better environmental or feeding conditions. In such situations, the area closure may in fact
368 increase disturbance and fishing pressure on those spawners whose offspring would otherwise
369 have a greater chance to survive (Table 2). This is because fishing effort is likely to be
370 concentrated in the areas outside the closure, as exemplified in Fig. 5. Expansion of the area
371 closure to cover most of the spawning (defined by 60m isobaths in BB) could avoid the
372 potential negative effect of the closure in relation to offspring survival. However, an area
373 closure in SD 25 could also cause fishing effort reallocation to SD 26, increasing the fishing
374 pressure on the remaining larger cod found in this area, with negative impacts on stock
375 structure.

376

377 This example demonstrates that a spawning close that is beneficial for the stock through one
 378 mechanism may at the same time compromise other aspects of the stock status. The relative
 379 importance of these different impacts is presently not possible to quantify with the data and
 380 knowledge available for the EB cod. The tradeoffs between different impacts are more likely
 381 to occur when the closures cover small areas, causing fishing effort to concentrate on other
 382 stock components during the time of the closure. In the case of the EB cod, avoiding the
 383 potential negative effects of the area closure in BB we have identified in this study, the
 384 closure would need to cover most of the stock distribution, i.e. the entire spawning area in BB
 385 and also SD 26. The latter is because it is not possible to identify smaller areas where the
 386 largest cod occur, and these are unlikely to correspond to the present small area closures in
 387 SD 26, which are not designed for that purpose.

388

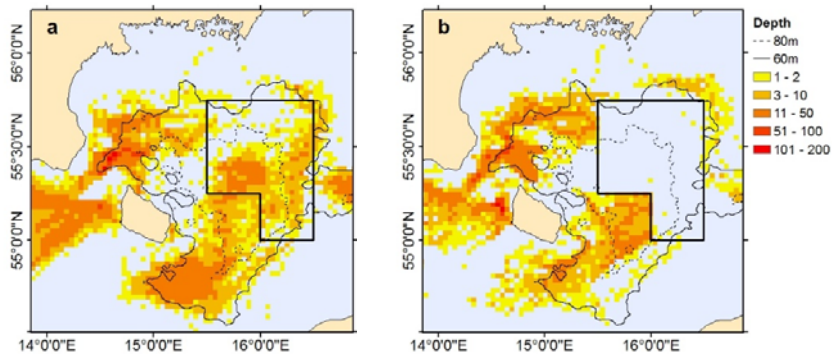
389 The seasonal closure (July 1–August 31) implemented in 2018 in the main cod distribution
 390 area (SD 25–26) does not cover the months of peak spawning (June) in recent years (ICES,
 391 2018*b*). The closure in July–August may to some extent cause temporal fishing effort
 392 reallocation to June increasing the disturbance of peak spawning in this month. However, this
 393 could be avoided simply by adjusting the timing of the closure to cover peak spawning. For
 394 WB cod, where the investigated closures covered the entire distribution area of the stock and
 395 matched the peak spawning time, potential negative effects associated with the closures were
 396 not identified.

397

398 Table 2. Summary of the potential positive and negative effects of the implemented area
 399 closure (*AC*) in Bornholm Basin (BB) and the seasonal closure (*SC*) in SDs 25–26 on the
 400 eastern Baltic cod.

| Closure | Potential positive effects | Potential negative effects |
|--|---|---|
| <i>AC</i> : BB May 1–Oct 31 | Undisturbed spawning of a fraction of the stock. | Part of the spawning, high survival of offspring, and larger cod occur in areas outside the closure, where fishing effort may reallocate. |
| <i>SC</i> : July 1- August 31(SDs 25–26) | Undisturbed spawning of a fraction of the stock; somewhat reduced fishing pressure on larger cod. | Possibly increased disturbance of peak spawning in June, due to temporal fishing effort reallocation. |

401



402

403 Fig. 5. Fishing effort in the Bornholm Basin: Example of the spatial distribution of fisheries in
 404 (a) months without the area closure (Nov–April) compared to (b) the months with the area
 405 closure in force (May–Oct). Based on Swedish and Danish VMS data for demersal fishing
 406 activities (defined by 2–4 knots speed), average for the years 2014–2016. The black solid box
 407 shows the extent of the area closure (EU, 2016).

408

409 5. Discussion

410

411 5.1 Challenges associated with evaluating the effects of spawning closures

412 Fishing closures during the spawning time of the fish are commonly applied in fisheries
 413 management worldwide, but their impacts on fish stock development are generally very
 414 difficult to demonstrate. Consequently, evaluations of the effectiveness of such closures have
 415 often remained inconclusive (e.g., Arendse et al., 2007; Clarke et al., 2015). Studies on
 416 closures in general have found that their effectiveness depends, amongst others, on their
 417 spatial and temporal properties considering the behavior and biology of the target species
 418 (Sheaves et al., 2006; van Overzee and Rijnsdorp, 2015). Size and age of the closure are
 419 important (Vandeperre et al., 2011) and what other fisheries management measures are
 420 applied in parallel (Beare et al., 2013; Clarke et al., 2015). Environmental and other
 421 influences often “mask” the effects of the closures on fish stocks complicating the evaluations
 422 of their effectiveness (Pastoors et al., 2000; Beare et al., 2013).

423

424 For the EB cod, the strongest year-classes occurred in the 1970s–early 1980s, i.e. prior to the
 425 implementation of the spawning closures, which was due to good environmental conditions in
 426 these years (Bagge et al., 1994). For the WB cod, both the strongest and the weakest year-
 427 class in the last 20 years were formed in years when most spawning was protected by the

428 closure (in 2016 and 2017) (ICES, 2018a). Although environmental and other factors
429 determine major fluctuations in the Baltic cod recruitment, spawning closures may modify the
430 recruitment possible to achieve under given ecosystem and environmental conditions.
431 However, such effects have so far not been possible to disentangle.

432

433 Recognizing the difficulty to demonstrate or quantify the *realized* effects of spawning
434 closures on fish stocks, we took a different approach in this study, where we investigated the
435 *potential* effects of the spawning closures on Baltic cod. In this approach, we evaluated
436 whether positive effects to the stock potentially can occur through specific mechanisms. This
437 approach does not verify whether a closure actually has a measurable effect on the stock.
438 Instead, it evaluates whether the design of the closure allows benefits to occur through a
439 specific mechanism, under the assumption that such benefits exist. The choice of the
440 mechanisms to consider in such evaluation depends at least partly on the specific objectives of
441 the closure, which are often not formally defined. Past reviews have identified lack of clear,
442 testable objectives as one of the basic obstacles for evaluating the effectiveness of the
443 implemented closures (STECF, 2007; Beare et al., 2013). Therefore, the first task is often to
444 define, based on knowledge and logic, some objectives for the closures that their effectiveness
445 can then be evaluated against (STECF, 2007). In this paper, we did not use the overarching
446 objective of improving the cod stock status as an evaluation criterion, because many other
447 factors influence the stock development. Instead, we focused on the specific mechanisms
448 through which the closures could potentially influence the cod stocks. This approach allows
449 considering potential both positive and negative impacts of the closures, which may be
450 important, even if not being part of the intended objectives of the closures.

451

452 The main potential benefits of spawning closures suggested in the literature (e.g., van Overzee
453 and Rijnsdorp, 2015) include the aspects of undisturbed spawning, offspring survival and
454 stock structure, which we addressed in this study. We considered these potential effects of the
455 spawning closures most relevant for the Baltic cod. However, we do not exclude that other
456 objectives could possibly be defined. For example, fishing closures can affect the total catch
457 of the species, which is often most rigorously and quantitative analysed (e.g., Kraus et al.,
458 2009; Clarke et al., 2015), in contrast to the other possible effects of the closures. STECF
459 (2016) has recently evaluated the effects of the seasonal closures on the EB cod catch,

460 suggesting that under the assumption of no effort reallocation, the closures could reduce the
461 total fisheries catch of the EB cod. However, when TAC management is in place as for the
462 Baltic cod, this can deliver the desired total catch level more directly than a closure, given that
463 TACs are effective in regulating total catch. For that reason, we focused this paper on the
464 potential benefits of the spawning closures related to recruitment, which would be difficult to
465 achieve by quota management alone.

466

467 **5.2 Lessons from the Baltic cod case study**

468 In the Baltic Sea, ichthyoplankton surveys monitoring spatio-temporal dynamics of cod eggs
469 and larvae (e.g., Hinrichsen et al., 2007) as well as studies on adult maturity (Bleil et al.,
470 2009) provide information on when and where cod spawning takes place. Regular monitoring
471 has demonstrated that the most intensive spawning activity is variable in time and space (Fig.
472 S1), implying that closures covering relatively small areas or short time periods have a low
473 chance of matching the peak spawning in all years. Furthermore, the EB cod is one of the
474 fortunate examples, where substantial process knowledge and modelling tools to evaluate
475 spatio-temporal variability in early life stage survival are available (Hinrichsen et al., 2002;
476 ICES, 2004; Kraus et al., 2009; Huwer et al., 2014). These studies have demonstrated that the
477 highest concentration of spawning activity is not always corresponding to the highest survival
478 probability up to juvenile stage, further complicating the design of spawning closures
479 covering relatively small areas.

480

481 In practice, it would likely be difficult to adjust the spawning closures to dynamic conditions
482 (Hinrichsen et al., 2007). The presently available knowledge on cod early life stage
483 production and survival in the Baltic Sea is a result of several decades of research efforts. This
484 has allowed identifying issues that may be important in relation to spawning closures,
485 however not necessarily feasible to adjust to in real time. Such investigations are often time
486 consuming and labour intensive, and therefore not regularly updated. Moreover, even in the
487 relatively data rich case of the Baltic cod, a number of knowledge gaps still exist. For
488 example, it is currently not possible to quantify the spatio-temporal origin of surviving
489 recruits in absolute terms, or describe the fine scale spatial distribution of different stock
490 components during the spawning time. Both of these questions are highly relevant especially
491 for designing smaller area closures.

492

493 The data collection for fisheries management purposes generally focuses on traditional fish
494 stock assessments, which provide the biological basis for setting annual catch limits. As
495 spawning closures can potentially affect a fish stock through various complex mechanisms,
496 data requirements for their proper evaluation are much greater, including, for example
497 information on spawning behaviour and physiology of the fish (Morgan et al., 1999; Dean et
498 al., 2014). The information relevant for designing and evaluating spawning closures, if
499 existing, is mostly produced via scientific programs, which are generally decoupled from
500 management needs. This implies that the information is often insufficient, irregularly updated,
501 or not tailored to the purpose of evaluating management measures such as spawning closures.

502

503 It is unclear whether monitoring and research on this topic can substantially increase in future.
504 Therefore, it is important to choose management measures, which are robust to data
505 limitations and related uncertainties. The Baltic cod example demonstrates that designing
506 smaller area closures properly is associated with much greater data requirements compared to
507 a closure covering most of the distribution area of the stock during its spawning time. This is
508 because small area closures cause fishing effort reallocation to other stock components with a
509 risk of unintended negative effects via the mechanisms that may not have been accounted for
510 when designing the closure. To avoid these counterproductive effects, a closure would need to
511 be sufficiently large. This is in line with experiences from other areas, suggesting that size is
512 an important feature of the closed areas in general (Edgar et al., 2014).

513

514 Quantifying the actual effects of spawning closures likely remains a challenge also in future.
515 Therefore, if spawning closures are chosen to be applied as a supplementary management
516 measure, these should be designed in a way that allows their potential benefits to occur, while
517 avoiding potential counteracting effects. The Baltic cod example suggests that the closures
518 covering most of the distribution area of the stock during its peak spawning time are better
519 suited for this purpose rather than those covering small areas.

520

521 **Conclusions**

522

523 The example of cod in the Baltic Sea illustrates the complexity of considerations that may be
524 involved in designing an appropriate spawning closure to improve fish stock status. A closure
525 and the resulting fishing effort reallocation can potentially affect the stock via a number of
526 mechanisms, which can include unintended negative effects counteracting the expected
527 benefits of the closure. Proper evaluation of the different mechanisms how a closure can
528 affect the stock has high demands for data and biological knowledge, which may not be
529 present even in data-rich cases such as the Baltic cod. Among the two types of closures we
530 have investigated, the design of smaller area closures generally involves greater complexity
531 and data requirements compared to the closures covering most of the distribution area of the
532 stock during the spawning time. Also, smaller area closures are associated with a higher risk
533 of having negative effects to the stock, if not rigorously assessed and adapted to changing
534 conditions. The spawning closures covering most of the distribution area of the stock are
535 generally more robust to the uncertainties and gaps in biological knowledge.

536

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538

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548

549 **Supplementary data**

550

551 Supplementary information is available at the online version of the article.

552

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