



Sedimentary record of plutonium in the North Yellow Sea and the response to catchment environmental changes of inflow rivers

Xu, Yihong; Pan, Shaoming; Gao, Jianhua; Hou, Xiaolin; Ma, Yongfu; Hao, Yongpei

Published in:
Chemosphere

Link to article, DOI:
[10.1016/j.chemosphere.2018.05.082](https://doi.org/10.1016/j.chemosphere.2018.05.082)

Publication date:
2018

Document Version
Peer reviewed version

[Link back to DTU Orbit](#)

Citation (APA):

Xu, Y., Pan, S., Gao, J., Hou, X., Ma, Y., & Hao, Y. (2018). Sedimentary record of plutonium in the North Yellow Sea and the response to catchment environmental changes of inflow rivers. *Chemosphere*, 207, 130-138. <https://doi.org/10.1016/j.chemosphere.2018.05.082>

General rights

Copyright and moral rights for the publications made accessible in the public portal are retained by the authors and/or other copyright owners and it is a condition of accessing publications that users recognise and abide by the legal requirements associated with these rights.

- Users may download and print one copy of any publication from the public portal for the purpose of private study or research.
- You may not further distribute the material or use it for any profit-making activity or commercial gain
- You may freely distribute the URL identifying the publication in the public portal

If you believe that this document breaches copyright please contact us providing details, and we will remove access to the work immediately and investigate your claim.

1 Sedimentary records of plutonium ~~Plutonium~~ in
2 ~~sediments of~~ the North Yellow Sea and ~~its~~ response
3 to catchment environment of inflow rivers

4 *Yihong Xu¹⁾, Shaoming Pan^{1)*}, Jianhua Gao^{1)*}, Xiaolin Hou²⁾, Yongfu Ma³⁾, Yongpei Hao¹⁾*

5 ¹⁾ The Key Laboratory of Coastal and Island Development of Ministry of Education, School of
6 Geographic and Oceanographic Sciences, Nanjing University, Nanjing 210023, China

7 ²⁾ Xi'an AMS Center, SKLLQG, Institute of Earth Environment, CAS, Xi'an 710061, China

8 ³⁾ The Key Laboratory of Environmental Radiation Monitoring of Ministry of Environmental
9 Protection; Radiation Monitoring Department, Environmental Radiation Monitoring Technical
10 Center of Ministry of Environmental Protection, Hangzhou 310012, China

11 ABSTRACT: Plutonium isotopes were first determined in surface and core sediment samples
12 collected from the northern North Yellow Sea to elucidate their source terms and deposition
13 process as well as response to the environment. ²⁴⁰Pu/²³⁹Pu atomic ratios in all sediments showed
14 the typical global fallout value of ~0.18 without any influences from the nuclear weapons tests

* Corresponding authors

E-mail: span@nju.edu.cn (S. M. Pan)

jhgao@nju.edu.cn (J. H. Gao)

15 conducted recently in the North Korea or early in the Pacific Proving Ground. The large
16 variation of $^{239+240}\text{Pu}$ activities (0.022-0.515 mBq/g) in surface sediments was mainly attributed
17 to the re-suspension and transportation of fine sediments influenced by the Liaonan Coastal
18 Current. Based on the two Pu depth profiles with easily observed onset (1952) and global fallout
19 peaks (1963), $^{239+240}\text{Pu}$ served as a valid time mark in the coastal sedimentary system. Riverine
20 input plutonium contributes only 15-27% to the total global fallout inventory (92.5-108.8 Bq/m²)
21 in the study area, much lower than that in the Yangtze River estuary (77-80%), indicating a
22 better soil conservation in the northeast China compared to the Yangtze River's drainage basin.
23 Riverine input Pu increased after 1980s reflected the increased soil erosion degree caused by the
24 land use and cover change due to the increased human activities in the northeast China. These
25 results demonstrated that plutonium is a good indicator for studying sedimentary process and its
26 response to the environment in the coastal area.

27 Key words: Plutonium, North Yellow Sea, sediment, isotopic ratio, chronological marker,
28 riverine input

30 Highlights

31 1. The atmospheric fallout is the dominant source of Pu in the North Yellow Sea.

32 2. PPG input Pu has no influence on the North Yellow Sea.

33 3. Riverine input Pu contributed about 15-27% to the northern North Yellow Sea

34 4. Riverine input Pu contributed much less in the NYS than in the Yangtze River estuary.

35 5. Pu is a good indicator to study the sediment process in coast sea.

36 1. INTRODUCTIONIntrduction

37 Plutonium isotopes (^{239}Pu and ^{240}Pu) have been substantially introduced to the marine
38 environment mainly from the global fallout of extensive atmospheric nuclear weapons tests
39 (NWT) since 1945. The global fallout has not equally dispersed on the ocean surface, but had a
40 maximum deposition over the mid-latitude in the western North Pacific Ocean due to the
41 geographic and climatological conditions [\(Gatsaud et al., 2011; Hirose et al., 2001\)](#).^{1,2} In
42 addition to the global fallout, close-in fallout from the Pacific Proving Ground (PPG), where a
43 series of U.S. nuclear tests were conducted in the 1940s-1950s, was a second significant source
44 of Pu in the North Pacific Ocean. ^{239}Pu ($T_{1/2}=24110$ yr) and ^{240}Pu ($T_{1/2}=6563$ yr) in the marine
45 environment are of great environmental concern due to their high chemical toxicity and long
46 half-lives. Besides to be good indicators for radioactive pollution, Pu isotopes can also serve as
47 useful geochemical tracers for better understanding a variety of marine processes [\(Lindahl et al.,](#)
48 [2010\)](#).³

49 With the high particle affinity, Pu isotopes entered into the oceans can be effectively
50 scavenged by settling particulate matter from the water column to the seafloor sediments. This
51 process is well known as “boundary scavenging” and generally enhanced in ocean
52 boundaries/margins, where are usually areas with high biological productivity, high particle flux
53 and sediment accumulation [\(Hong et al., 2011\)](#).⁴ Seafloor sediments are therefore the ultimate
54 sink of Pu entering the marine environment, and preserve Pu deposition history, from which
55 valuable information of impact of human activities on the marine environment and its changes
56 over time can be deduced. With the well constrained release history of the global fallout Pu from
57 the NWTs, which began in the late 1940s and had a distinct deposition maximum in 1963
58 [\(UNSCEAR, 2000\)](#),⁵ undisturbed sediment recording these two events can be identified and thus

59 used for dating purpose [\(Corcho-Alvarado et al., 2014\)](#).⁶ With the recent rapid development in
60 analytical procedures and mass spectrometric measurements [\(Qiao et al., 2009\)](#),⁷⁻⁸ $^{239+240}\text{Pu}$ has
61 been suggested as a chronostratigraphic marker for modern sediments dating in both freshwater
62 bodies (e.g. lakes and river estuaries)[\(Ketterer et al., 2004; Wu et al., 2010; Zheng et al., 2008\)](#)⁹⁻
63 ¹²-and coastal areas [\(Corcho-Alvarado et al., 2014; Hancock et al., 2011\)](#),^{6,-13} and shown some
64 advantages to the conventional ^{137}Cs method, such as more precise determination especially in
65 marine sediments[\(Hancock et al., 2011\)](#)¹³ and additional chronometers associated with changes
66 in $^{240}\text{Pu}/^{239}\text{Pu}$ atomic ratios for Pu fallout during the 1950s and 1960s [\(Koide et al., 1985\)](#)¹⁴ as
67 well as fallout from the 1986 Chernobyl accident in some European areas [\(Ketterer et al., 2004\)](#).
68 ⁹

69 In the past decades, numerous studies have been carried out to investigate source terms,
70 transport, and scavenging and deposition processes of Pu isotopes in China marginal seas [\(Hong](#)
71 [et al., 2006; Huh & Su, 1999; Lee et al., 2003; Liu et al., 2011; Nagaya & Nakamura, 1992; Pan](#)
72 [et al., 2011; Su & Huh, 2002; Wang et al., 2005; Wang et al., 2017; Wu et al., 2014; Zheng &](#)
73 [Yamada, 2006\)](#).¹⁵⁻²⁷ However, most of these studies concentrated in the [East China Sea ESC](#) and
74 the Yangtze River Estuary. Regarding Pu isotopes in the Yellow Sea (~~YS~~) sediments, only
75 several data points have been reported in the [South](#)~~thern~~ Yellow Sea (~~SYC~~) and Jiangsu tidal
76 flats [\(Hong et al., 2006; Liu et al., 2013; Nagaya & Nakamura, 1992\)](#).^{15, 21, 25} No Pu data in the
77 North Yellow Sea (NYS) sediments is available. The ~~YS~~-[Yellow Sea](#) located between mainland
78 China and the Korean Peninsula is a semi-closed shallow marginal sea. It is normally divided
79 into two parts: the ~~SYC~~-[South Yellow Sea](#) is connected to the ~~ESC~~[East China Sea](#), while the
80 NYS, considered in the present study, is adjacent to the Bohai Bay and surrounded by the
81 Liaodong Peninsula, Shandong Peninsula and Korean Peninsula. The coastal areas of the NYS

82 are densely populated areas, where are very close to North Korea. Nuclear activities including
83 NWTs in North Korea have caused a great concern on the radiation exposure to the public in the
84 past years. The Hongyanhe nuclear power plant (NPP) with 4 units located in Dalian (39.792°N,
85 121.475°E) on the Liaodong Peninsula has already being in operation from 20th September 2016.
86 ~~In addition, there are plans to build 4 more NPPs in China's three northeast provinces.~~ Potential
87 releases of radioactive materials including Pu from the NPPs to the surrounding environment and
88 the consequences are also a major concern of the local inhabitants and the authorities.

89 In the north coast of the NYS, major rivers flowing into the sea include Yalu River, Dayang
90 River and Biliu River. These rivers play important roles in the transportation of terrigenous
91 sediments to the northern NYS. The fluvial sediment discharges providing significant source of
92 sediments in the coastal sea ([Liu et al., 2013](#)), are related to the soil erosion degree in river basins
93 and can vary in response to climate changes and human activities in the catchment, which can be
94 well reflected in the coastal sedimentary system. Thus investigation of the sediment deposition
95 process in the coastal sea and its dynamics is of great significance for evaluating water and soil
96 loss in river basins and comprehensive understanding the land-ocean interaction within the
97 coastal areas.

98 In this work, surface sediments and sediment cores were collected in the northern NYS
99 (Figure 1) and analyzed for ²³⁹Pu and ²⁴⁰Pu, aiming (1) to investigate the source term, distribution
100 and deposition process of Pu in the study area; (2) to identify the contributions of direct
101 deposition and riverine input by the deposition history of Pu, as well as to evaluate the water loss
102 and soil erosion in the coastal major watersheds and its response to ~~climate change and~~ human
103 activities; (3) to provide baseline data of Pu for environmental risk assessment in relation to the
104 nuclear activities in North Korea and the Hongyanhe NPP.

105

106 2. MATERIALS AND METHODS Materials and methods

107 2.1 Study Area

108 The study area was located along the southeastern coast of the Liaodong Peninsula, in the
109 northern NYS (Figure 1). Many islands are scattered along the coast (e.g. Shicheng Island,
110 Changshan Island, etc.), resulting in the complex subwater topography in the study area ([Chen et
111 al., 2013](#)).²⁸ There are a number of rivers flowing into the NYS, among which the Yalu River is
112 the largest one with the sediment discharge of 113×10^4 tons/a ([State Oceanic Administration,
113 1998](#)).²⁹ Besides, the Dayang River and Biliu River with the sediment discharges of 69.3×10^4
114 tons/a and 53.3×10^4 tons/a also contribute significant amounts of sediments to the study area
115 ([Bao & Du, 1999](#)).³⁰ The water circulation in the study area is dominated by the Liaonan Coastal
116 Current (LCC), which is formed by the Yalu River Diluted Water Mass and flows constantly
117 along the southeast coast of the Liaodong Peninsula to the Bohai Strait. It has been reported that
118 the Yalu River derived sediments were re-suspended and then transported by the LCC
119 southwestward for a long distance and re-deposited along the southeast coast of Liaodong
120 Peninsula ([Chen et al., 2013](#)).²⁸ In addition to the LCC, the strong Yellow Sea Warm Current
121 (YSWC) in winter, the Yellow Sea Coastal Current (YSCC) and the West Korean Coastal
122 Current (WKCC) might also affect the material transport in the study area.

123 2.2 Sample collection and pretreatment

124 Sampling was conducted onboard in July 2016 and March 2017, respectively. Eight surface
125 sediments (0-2 cm depth) and two sediment cores ([NYSC-01 with water depth of 10 m and
126 NYSC-02 with water depth of 20 m](#)) were collected using the grab buckets and a gravity
127 sampler, respectively. The sediment cores were sliced in 2 cm intervals and weighed

128 immediately, 2 g of sediment was taken from each surface and interval sediment sample for the
129 analysis of the grain size composition by the Malvern Mastersize 2000 laser granulometer
130 (measuring range 0.02-2000 μm , with a duplicate measurement error of $< 3\%$). The remained
131 sediment samples were freeze-dried and reweighed to determine mass depths in the laboratory.
132 Dried samples were then ground to fine powder for Pu isotopes analysis.

133 [2.3](#) Determination of Pu isotopes

134 The method for Pu isotopes analysis is similar as described in [our previous publications Xu et](#)
135 [al., 2014](#).³¹⁻³⁴ Briefly, approximately 5 g of dried sediments for each sample were weighed to a
136 beaker and ashed at 550 °C overnight to decompose the organic matter as well as to obtain
137 organic matter (~~OM~~) content by calculating the mass loss in this process. After being spiked with
138 a known amount of ^{242}Pu (ca. 1 mBq) as a chemical yield monitor (the ^{242}Pu added to the ashed
139 aliquot does not cover the losses during the mineralization, but they should be negligible), the
140 samples were leached using *aqua regia* for 2 h at 200 °C, Pu in the leachate was co-precipitated
141 with iron hydroxides to remove the major matrix components. After centrifuge, the precipitate
142 was dissolved with concentrated HCl. Then Pu in the solution was reduced to Pu(III) by adding
143 about 300 mg of $\text{K}_2\text{S}_2\text{O}_5$ and then oxidized to Pu(IV) with conc. HNO_3 . A 15 mL AG1- $\times 4$ anion-
144 exchange column and a 2 mL TEVA column were successively used for separation and
145 purification of Pu isotopes. The final sample solution prepared in 3 mL of 0.5 mol L^{-1} HNO_3 was
146 analyzed for Pu isotopes (^{239}Pu , ^{240}Pu , ^{242}Pu) using a section field ICP-MS (SF-ICP-MS)
147 (Element 2, Thermo Scientific, Bremen, Germany). An APEX-Q high efficiency sample
148 introduction system (Elemental Scientific Inc., Omaha, NE, USA) combined with a membrane
149 desolvation unit (ACM) and a conical concentric nebulizer was used to introduce sample
150 solution to the ICP-MS. All the measurements were made in the self-aspiration mode with an

151 uptake rate of $\sim 0.2 \text{ mL min}^{-1}$ to reduce the risk of contamination from the long peristaltic pump
152 tubing. 0.1 ng mL^{-1} U standard solution was used for tuning SF-ICP-MS before the samples
153 were measured each time. The ^{242}Pu standard solution in the same media as the samples and
154 procedure blanks were measured with the samples. The concentrations of ^{239}Pu and ^{240}Pu in the
155 sample were calculated by the measured signal intensities of ^{239}Pu , ^{240}Pu and ^{242}Pu and the
156 amount of ^{242}Pu added to the sample before chemical separation. ^{242}Pu was used in this
157 procedure as both dilution isotope and chemical yield monitor.

158 3. Results and discussions ~~ESULTS AND DISCUSSION~~

159 3.1 Source terms of Pu isotopes in the northern NYS

160 Since the isotopic composition of Pu is related to its production and release, the atomic ratio
161 of $^{240}\text{Pu}/^{239}\text{Pu}$ in environmental samples could be used to identify the source term of plutonium
162 contamination. It has been well proved that the global fallout has an average $^{240}\text{Pu}/^{239}\text{Pu}$ ratio of
163 0.18 ([Kelley et al., 1999](#); [Warneke et al., 2002](#)).³⁵⁻³⁷ $^{240}\text{Pu}/^{239}\text{Pu}$ atomic ratios for weapons-grade
164 Pu were reported to be very low ($0.01 \leq 0.07$) ([Lee & Clark, 2005](#)),³⁸⁻³⁹ while much higher values
165 were reported in materials of nuclear power reactors (0.2-0.8, depending on the reactor type and
166 fuel burn-up) ([Muramatsu et al., 2000](#); [Warneke et al., 2002](#)).^{37,40} It has also been found that
167 different test series have different atomic ratio. For example, the $^{240}\text{Pu}/^{239}\text{Pu}$ atomic ratios in the
168 close-in fallout at the Nevada test site are generally very low with an average of 0.035 ([Buessler,](#)
169 [& Sholkovitz, 1987](#)),⁴¹ while close-in fallout in the PPG of the Marshall Islands in the 1950s is
170 characterized by higher ratios of 0.30-0.36 ([Buessler, 1997](#); [Koide et al., 1985](#); [Muramatsu et al.,](#)
171 [2001](#)).^{14, 42-43} Such characteristic of PPG-derived plutonium has been intensively employed to
172 trace transportation of PPG-derived pollutions in the North Pacific Ocean and its marginal seas
173 including the [ESC–East China Sea](#) and [SYSSouth Yellow Sea](#). However, it is still unclear

174 whether the PPG-derived pollution has entered to the [YSYellow Sea](#), especially the NYS due to
175 lack of information on Pu isotopic compositions in this area.

176 As presented in Table [A.S1](#), [A.S2](#) ([Supporting Supplementary material information](#)) and
177 depicted in Figure 2, the measured $^{240}\text{Pu}/^{239}\text{Pu}$ atomic ratios ranged from 0.175 to 0.190 with an
178 average of 0.184 ± 0.005 in all surface sediment samples. In the two sediment soil cores,
179 $^{240}\text{Pu}/^{239}\text{Pu}$ atomic ratios varied from 0.164 to 0.233 with an average of 0.194 ± 0.02 in NYSC-
180 01 core, and from 0.163 to 0.206 with an average of 0.185 ± 0.011 in NYSC-02 core, which were
181 relative constant and similar to those determined in surface sediments. It is clear that the atomic
182 ratios of $^{240}\text{Pu}/^{239}\text{Pu}$ obtained in all sediment samples agreed very well with the average value of
183 ~ 0.18 for the global atmospheric fallout, as well as with ~~our~~ previous reported values (~ 0.18) in
184 soils collected from Liaodong Bay coastal regions ([Xu et al., 2013](#)),^{31,33-34} which are quite close
185 to the present study area. Similar $^{240}\text{Pu}/^{239}\text{Pu}$ atomic ratios have also been reported in a sediment
186 core from the Lake Sihailongwan (42.29°N, 126.6°E) ([Wu et al., 2010](#)),⁴¹ which is also located
187 in the northeast China and not very far (<400 km) from our study area. Combined these results
188 with our previous conclusion regarding Pu isotopes in soils of northeast China that the major
189 source of Pu in northeast China is the global atmospheric fallout, and any other possible
190 contributions should be very small, Pu isotopes in the sediments collected from the present study
191 area should be also mainly from the global atmospheric fallout.

192 It is well known that several NWTs have been conducted in the North Korea in the past ten
193 years. However, Pu is a refractory element and is not easily to be released to the atmosphere
194 during underground NWT, thus contributions of Pu from the NWTs in North Korea should be
195 negligible in spite of the short distance between the North Korea and our study area. Plutonium
196 released from the Fukushima Daiichi Nuclear Power Plant (FDNPP) accident might have

197 potential impact on the surrounding areas including the [YS Yellow Sea](#). ~~Since~~ ~~However~~, the
198 FDNPP-derived Pu has been reported to have much higher $^{240}\text{Pu}/^{239}\text{Pu}$ atomic ratios (0.303 to
199 ~~0.381-330~~) ([Zheng et al., 2012](#))⁴⁴⁻⁴⁵ than the measured values in present studied sediments
200 collected in 2017, ~~and some other reports~~ ([Bu et al., 2014](#); [Hain et al., 2017](#); [Oikawa et al., 2015](#);
201 ~~have also indicated that the influence of the FDNPP-derived Pu to the marine environment~~
202 ~~would be limited.~~ ~~Thus~~ it is reasonable to believe that the contribution from the FDNPP-source
203 Pu to the total Pu inventory in our study area is negligible.

204 Numbers of studies have reported that the close-in fallout Pu from the PPG has been
205 transported from the Marshall lands through the North Equatorial Current followed by the
206 Kuroshio Current and its extensions to the [SCS-South China Sea](#) and [ESCEast China Sea](#) ([Liu et](#)
207 [al., 2011](#); [Wang et al., 2005](#); [Wang et al., 2017](#); [Wu et al., 2014](#)).^{19, 24, 26-27} Significant higher
208 $^{240}\text{Pu}/^{239}\text{Pu}$ atomic ratios (> 0.20) were usually observed in surface sediments or bottom layers of
209 sediment cores in these marginal seas due to the contributions of PPG-sourced Pu with high
210 $^{240}\text{Pu}/^{239}\text{Pu}$ atomic ratios of 0.30-0.36. But this is not the case in our study area, since no
211 abnormal higher ratios of $^{240}\text{Pu}/^{239}\text{Pu}$ were observed in both surface sediment samples and
212 bottom layers of the two sediment cores. This indicates at least that our study area did not receive
213 any PPG derived Pu from either the early direct close-in fallout or oceanic current transportation.
214 Although earlier studies regarding Pu in surface seawater samples from the central [YS-Yellow](#)
215 [Sea](#) and [SYS-South Yellow Sea](#) has shown some PPG-sourced Pu signals in surface seawater
216 with relative higher ratios (0.18-0.33 with an average of 0.23) ([Kim et al., 2004](#)),⁴⁶ the impact of
217 the PPG close-in fallout Pu on the [Yellow Sea](#) is still doubtful. Liu et al. ([2013](#))²⁵ studied the
218 Pu isotopes in the North Jiangsu tidal flats of the [YS-Yellow Sea](#) and they concluded that the
219 tidal flats did receive mainly global fallout Pu but no PPG-derived Pu in any pathways, which

220 are similar as our conclusions for the northern NYS. Therefore, the influence of the PPG-derived
221 Pu on the [YSYellow Sea](#), especially the NYS might be very limited.

222 All above results suggested that the global atmospheric fallout is the major source of Pu in
223 the northern NYS, and any other possible contributions such as releases from NWTs of the North
224 Korea, the FDNPP accident, as well as early direct deposition or later oceanic current
225 transportation of PPG derived Pu are negligible. Since Feb. 2013, the Hongyanhe NPP has
226 started in operation of four units successively. The measured Pu isotopic composition in this
227 work will serve as a reference data for the future monitoring and assessment of environmental
228 impact of this NPP.

229 [3.2](#) Spatial distribution of Pu isotopes in surface sediments in the northern NYS

230 The $^{239+240}\text{Pu}$ activities in surface sediments collected along the southeast coast of the Liaodong
231 Peninsula in the NYS are presented in Table [A.S1](#) and Figure 3. A large variation of $^{239+240}\text{Pu}$
232 concentrations from 0.022 to 0.515 mBq/g with an average of 0.122 ± 0.021 mBq/g ($n = 10$) was
233 observed in these sediment samples (Figure 2). Such $^{239+240}\text{Pu}$ activity levels are comparable to
234 the reported values observed in the NYS (0.045 to 0.901 mBq/g) ([Hong et al., 2006; Nagaya &](#)
235 [Nakamura,1992](#))-^{15,21} and the [ESC-East China Sea](#) (0.048-0.759 mBq/g) ([Liu et al., 2011; Wang](#)
236 [et al., 2017](#)).^{24,27} It can be observed from the Figure 3 that the $^{239+240}\text{Pu}$ concentrations generally
237 show a westwards increasing trend along the coast with the highest values near the northeast cost
238 of Dalian Bay, which is similar to the direction of the LCC.

239 The distribution of Pu concentrations in surface sediments might be influenced by many
240 factors such as organic matter and particle size of the sediment, coastal currents and riverine
241 input. Since the differences of organic matter content in the surface sediments excluding the
242 surface sample of the NYSC-02 sediment core are very small, the influence of the organic matter

243 contents on the $^{239+240}\text{Pu}$ activities was not significant. ~~It has been reported by~~Based on the
244 [analysis of the rare earth element compositions of the surface sediments in the southeast coastal](#)
245 [area of the Liaodong Peninsula](#), Chen et al.(2013)²⁸ [have reported](#) that after re-suspension and
246 transportation by the LCC, the Yalu River-derived fine-grained sediments were re-deposited in
247 water depth of 20-40 m in the area, where is 180-300 km away from the Yalu River mouth and
248 extends along the southeast coast of the Liaodong Peninsula between the northeast of Dalian Bay
249 and southwest of the Changshan Islands [\(as shown in Figure 1\)](#). It is obvious that the two
250 sampling sites of SS05 and SS06 in this study are just located in the reported Yalu River-derived
251 fine-grained deposition region. ~~Our previous results-Researches~~ regarding the association of Pu
252 isotopes with natural soil particles have proved that Pu isotopes tend to associate with finer
253 particles due to their bigger specific surface areas [\(Xu et al., 2017\)](#).³⁴ Therefore, significant
254 higher $^{239+240}\text{Pu}$ activities in the sampling sites SS05 and SS06 are more likely to be attributed to
255 their much smaller particle sizes [\(SS05 and SS06 with mean grain sizes of 3.2 and 2.4 \$\mu\text{m}\$](#)
256 [respectively\)](#), ~~although there was no direct linear correlation between mean grain sizes and~~
257 ~~$^{239+240}\text{Pu}$ activities in all surface sediments~~. The lowest Pu concentrations in the east part of the
258 study area might be in relation to the dilution of Pu-depleted river-borne sediments from the
259 Yalu River and Dayang River. While in the most west two sampling sites SS07 and SS08, which
260 are close to the Bohai Strait, their very low $^{239+240}\text{Pu}$ concentrations (0.025 and 0.040 mBq/g) are
261 likely to be related to the interaction influence of the YSWC and YSCC. Hence, the transport of
262 coastal currents and sediment dynamics of fine-particles should play an important role in
263 controlling the spatial distribution of Pu isotopes in the northern NYS.

264 [3.3](#) Temporal variations of Pu isotopes in the northern NYS

265 The vertical profiles of $^{239+240}\text{Pu}$ activities in the two sediment cores are illustrated in Figure 4
266 and summarized in Table [A.S2](#). The $^{239+240}\text{Pu}$ concentrations in the NYSC-01 sediment core
267 ranged from 0.006 to 0.251 mBq/g with an average of 0.141 ± 0.061 mBq/g, which are nearly 2
268 times lower than those in the NYSC-02 sediment core with a range of 0.002 to 0.576 mBq/g and
269 an average of 0.231 ± 0.052 mBq/g. This difference may be partly attributed to the effect of
270 particle size on the association of Pu with sediments. Considering that the NYSC-02 sampling
271 site with water depth of 20 m and fine size grain sediment of a mean grain size of $6.04 \mu\text{m}$ is
272 located in the fine grained deposition region, higher Pu concentrations would be expected at this
273 site due to the stronger adsorption of plutonium by finer grain sediments. While the NYSC-01
274 sampling site with water depth of 10 m and relative coarser grain sediment of a mean grain size
275 of $6.82 \mu\text{m}$ is located in the pathway of the LCC and much closer to the mouth of the Yalu River.
276 In the near shore zone, due to the strong hydrodynamic conditions, fine particles adsorbed more
277 Pu isotopes are easily re-suspended in water column and further transported westwards through
278 the LCC, but coarse sediments with weaker adsorption of Pu isotopes were deposited locally,
279 resulting in the lower Pu concentrations in the NYSC-01 site. Moreover, the difference in
280 organic matter contents may also be one reason of different Pu concentrations in these two
281 sediment cores. Although linear regression analyses showed that there was no obvious
282 correlation between Pu concentrations in the NYSC-02 core and its relative constant organic
283 matter contents from the top to the bottom layers. While, the relationship between the organic
284 matter contents and Pu concentrations in the NYSC-01 core is significant ($r^2 = 0.64$, $p < 0.05$) as
285 shown in Figure [A.S1](#). The organic matter content in the NYSC-02 (with an average of 7.32%)
286 was much higher than that in the NYSC-01 core (average of 5.84%).

287 It can be observed from the Figure 4 that $^{239+240}\text{Pu}$ activities in both sediment cores showed
288 unambiguous maximum values and background levels. Based on the previous discussions that Pu
289 isotopes in the study area are mainly from the global fallout and the well-defined global fallout
290 history, it is reasonable to relate the pronounced peaks to the historical global fallout maximums
291 in 1963 and the positions of background values to the 1952 horizons (onset fallout). Using the
292 1963 horizon as a discrete time marker, we obtained similar mean sedimentation rate of 0.46
293 cm/yr for both sediment cores, but two different sediment accumulation rates of 0.61 g/cm²·yr for
294 the NYSC-01 core and 0.52 g/cm²·yr for the NYSC-02 core, respectively. While taking the 1952
295 horizons as a chronological marker for sediment dating, the mean sedimentation rates and mean
296 sediment accumulation rates were calculated to be 0.6057 and 0.51 cm/yr and 0.774 and 0.56
297 g/cm²·yr respectively for the NYSC-01 and NYSC-02 cores, which are slightly higher but still
298 comparable to the results obtained from the 1963 marker. The difference between the results
299 obtained from the two time markers has been evaluated using a statistic *t*-test ($t_{\text{stat}} = -3.18134 < t_{\text{crit}}$
300 for $n=4$), which suggests that both 1963 and 1952 horizons are valid chronological markers for
301 the sediment dating in the present study area. In fact, our results are also within the range of the
302 results (0.06-1.18 g/cm²·yr) reported by Qi et al. (2004)⁴⁷ using $^{210}\text{Pb}_{\text{ex}}$ dating method regarding
303 the sediment cores in the NYS. The mean sediment accumulation rates obtained in the NYSC-01
304 core from the both time markers were obviously higher than those in the NYSC-02 core,
305 indicating that the sedimentary environment in the latter sampling site are relatively stable in the
306 past decades. This was reasonable and coincided well with the previous discussions.

307 Further insight into the temporal variation of Pu isotopes in the northern NYS was provided
308 by the distribution patterns of Pu concentrations from the surface to the 1963 horizon in both
309 sediment cores. According to the report of UNSCEAR (2000),⁵ the temporal variation of the

310 annual global fallout deposition of ^{137}Cs in northern hemisphere (Figure 5) showed a maximum
311 in 1963 and then significantly decreased to a low level especially after 1980. The historical
312 pattern of the global fallout deposition of $^{239+240}\text{Pu}$ is believed to be similar. According to the
313 estimation of Nakano and Povinec (2003),⁴⁸ the annual global fallout deposition of $^{239+240}\text{Pu}$ in
314 marine environment was also found to be maximal in 1963 and then decrease dramatically.
315 While the direct deposition of the global fallout Pu was recorded in undisturbed lakes, as
316 depicted in the Figure 5, typical $^{239+240}\text{Pu}$ concentration peaks corresponding to the maximum
317 deposition of 1963-1964 and followed by sustained declines were observed in lake Qinhai and
318 lake Hongfeng which are located respectively in northwest and southwest of China (Wu et al.,
319 2011; Zheng et al., 2008).^{10,12} However, A constant high $^{239+240}\text{Pu}$ activities were observed in the
320 sediment core NYSC-01 corresponding to the period of 1979- 2009 (Figure 4 and 5), and
321 significant higher $^{239+240}\text{Pu}$ activities were observed in the sediment core NYSC-02 from the
322 mass depth of 15.13 g/cm^2 to 5.99 g/cm^2 corresponding to the period of 1988 to 2005. This
323 indicates that besides the direct atmospheric fallout, there are other inputs of plutonium to the
324 sediment, especially after 1980. The riverine input sediments which are mainly sourced from the
325 soil erosion in the river catchments should be the main source of the excessive plutonium in the
326 sediment cores. Plutonium in the riverine sediments in the northeast China should also originate
327 from the global fallout according to the previous studies based on the constant $^{240}\text{Pu}/^{239}\text{Pu}$ ratios
328 in the whole soil and sediment cores (Wu et al., 2010; Xu et al., 2013).^{11,31-34} In conclusion, the
329 unusual distributions of global fallout Pu in the upper layers of the two sediment cores is
330 attributed to the riverine input sediments over ~~time,time~~ and this will be further discussed in the
331 next section.

332 3.4 Contribution of Pu from riverine input and its response to catchment environment

333 Inventory of $^{239+240}\text{Pu}$ in sediment cores can help to interpret the source function and the
334 transport of Pu, thus providing a useful indicator for assessing the sedimentary process in the
335 marine environment. In this study, how riverine input sediment influenced the global fallout Pu
336 depositions in the sediments could be further deduced from the inventories of Pu in the two
337 sediment cores. The total Pu inventories were estimated to be $92.5 \pm 1.6 \text{ Bq/m}^2$ for the NYSC-01
338 core and $108.8 \pm 1.2 \text{ Bq/m}^2$ for the NYSC-02 core, respectively. These values are nearly twice
339 the average values of global atmospheric fallout ($42\text{-}58 \text{ Bq/m}^2$) expected at the similar latitudinal
340 belts of $30\text{-}50^\circ\text{N}$ ([UNSCLEAR, 1982; 2000](#)),^{5,49} and also generally higher than those ($55.5\text{-}90$
341 Bq/m^2 , Table [A. S3](#)) obtained from the undisturbed soil and sediment sampling sites with similar
342 average annual precipitations ($600\text{-}800 \text{ mm}$) at the similar latitude of $\sim 39^\circ\text{N}$ worldwide ([Hardy](#)
343 [et al., 1973; Wu et al., 2010; Xu et al., 2013; 2015](#)),^{11,31,33,50} which are believed to receive solely
344 global fallout Pu. This indicates that besides the direct deposition from the global fallout, riverine
345 input should contribute to some Pu depositions in the study area. Since the integrated global
346 fallout values of 42 and 58.1 Bq/m^2 respectively for $30\text{-}40^\circ\text{N}$ and $40\text{-}50^\circ\text{N}$ published by
347 UNSLEAR ([1982; 2000](#))^{5,49} were average values without considering the influence of annual
348 precipitation on Pu fallout in specific sampling sites, they may not indeed represent the direct
349 deposition of global fallout in the study area. Comparing global fallout Pu inventory data
350 obtained from undisturbed soil and sediment cores with similar annual precipitation and latitudes
351 as the study area ($\sim 39^\circ\text{N}$), an average value of 78.9 Bq/m^2 (Figure 6-a) was believed to be more
352 close to the real direct deposition of global fallout Pu in the study area, and thus is taken into
353 account in the calculation of the contributions of Pu from both direct deposition and riverine
354 input.

355 The calculation results showed that the riverine input Pu contributes to about 15% and 27%
356 of the total Pu inventories in the NYSC-01 and NYSC-02 cores, respectively (Figure 6-b). Since
357 plutonium has high affinity to particle and can be easily adsorbed to soil particles after
358 deposition on the earth surface, contributions of Pu from the riverine input should directly reflect
359 the degree of soil erosion in river catchments. In order to evaluate whether this reflection is
360 reliable, we did a comparative study between the Yangtze River and the Yalu River catchments.
361 Based on the previous published Pu data from the sediment cores of SC07 and SC18 collected in
362 the Yangtze River estuary(Liu et al., 2011; Pan et al., 2011),²³⁻²⁴ the contribution of Pu from the
363 Yangtze River input to the total global fallout Pu inventory in the Yangtze River estuary was
364 estimated to be 77-80% as shown in Figure 6-b (note that the total global fallout Pu inventories
365 in the SC07 and SC08 cores were 244 and 217 Bq/m², respectively, corresponding to the total Pu
366 subtracted the PPG-derived Pu, and the direct deposition of global fallout for the Yangtze River
367 estuary was estimated to be 49.8 Bq/m² based on the data reported by Dong et al.(2010)⁵⁴).
368 Obviously, the riverine Pu contributions in the Yangtze River estuary are nearly 3 to 4 times
369 higher than that in our study area, implying that the status of soil erosion in the Yangtze River's
370 catchments might be much more serious than that of the Yalu River and other rivers in northeast
371 China.

372 The above conclusion was further supported by a comprehensive comparison of the
373 catchment area, average annual runoff and sediment discharge between these rivers as shown in
374 Table [A. S4](#). The average annual runoff per unit of catchment area in the Yangtze River
375 catchment is about 1.2 times higher than that in the three major river catchments of the study
376 area, which is reasonable and agrees with the relatively higher precipitation in the Yangtze River
377 catchment (~1000 mm) compared to that in the Yalu River catchment (~800 mm). While the

378 average annual sediment discharge of the Yangtze River was estimated to be 7.5 times the total
379 average annual sediment discharges of the three rivers in the study area, indicating the soil
380 erosion amount in the Yangtze River catchment should be much larger than that in the river
381 catchments of northeast China, which is consistent with the conclusion deduced from the
382 contributions of riverine Pu. The great difference of soil erosion between Yalu River's catchment
383 in the northeast China and Yangtze River's drainage basin in the central China should be mainly
384 attributed to the relatively higher forest coverage and less vegetation breakage by human
385 activities in the northeast China compared to the central China.

386 In addition, the sudden increase of riverine input Pu after 1980s could be generally observed
387 in both sediments cores as shown in Figure 5, ~~which might provide more insights on evaluation~~
388 ~~of the catchment environment in the northeast China. That riverine input Pu increased rapidly~~
389 ~~after 1980s should reflect this~~ might results from the increase of soil erosion degree in the
390 northeast China at the same period. ~~, which could be mainly attributed to land use and cover~~
391 ~~change (LUCC) in the area.~~ According to the studies of Li et al. (2010), due to the increased
392 human activities, the land use changed intensively in the northeast China from the 1980s to 2000.
393 During this period, cropland increased substantially, while woodland, grassland, water body and
394 marsh decreased sharply, and among all land use types, paddy fields increased
395 ~~fastest~~significantly (by 20 %), conversely, grassland and marsh decreased ~~rapidly~~ ~~fastest~~, by 17 %
396 and 12 %, respectively. Such land use and cover change ~~should~~ result in much more soil loss in
397 the northeast China.

398 4. Conclusions

399 The present work for the first time investigated the sedimentary records of ^{239}Pu and ^{240}Pu in
400 the NYS. From isotopic compositions of the two Pu isotopes in surface and core sediments, we

401 concluded that plutonium in the northern NYS is mainly from the global atmospheric fallout
402 without any significant contributions from NWTs of the North Korea, the FDNPP accident, as
403 well as PPG sourced Pu. From the depth profiles of $^{239+240}\text{Pu}$ activities with easily observed onset
404 and fallout peaks in the two sediment cores, $^{239+240}\text{Pu}$ was found to be a useful time marker in
405 sediments of the northern NYS. The riverine Pu was estimated to contribute 15-27% to the total
406 global fallout Pu inventory in the study area, which are much lower than that in the Yangtze
407 River estuary, indicating the soil erosion in the northeast China was much less than that in the
408 Yangtze River's drainage basin. The sudden increase of riverine input Pu after 1980s reflected
409 the increased soil erosion degree caused by the land use and cover change due to the increased
410 human activities in the northeast China. In addition, the activity level and the atomic ratio of Pu
411 isotopes reported in this work can serve as a baseline for future environmental monitoring and
412 risk assessment related to the operation of Hongyanhe NPP or any other potential nuclear
413 activities.

414 **Acknowledgments**

415 This work was jointly supported by the National Natural Science Foundation of China
416 (41671466, 41501286, and 41576043), Natural Science Foundation of Jiangsu Province
417 (BK20150578) and the Fundamental Research Funds for the Central Universities
418 (020914380010, 020914380001). Authors thank Dr. Zhongtang Wang and Wenna Huang for
419 their supports to sample analysis, as well as Dr. Zuo Wang for his help with map drawing.

420 **Appendix A. Supplementary material**

421 **Supplementary material related to this article can be found online at**

422 **References**

423 Bao, Y.E., Du, Z.X., 1999. Annals of Chinese Bay (Volume 3): Estuaries (Liaoning). China
424 Ocean Press, Beijing. (in Chinese).

425 Buesseler, K.O., 1997. The isotopic signature of fallout plutonium in the North Pacific. J.
426 Environ. Radioact. 36 (1), 69-83.

427 Buesseler, K.O., Sholkovitz, E.R., 1987. The geochemistry of fallout plutonium in the North
428 Atlantic: II. $^{240}\text{Pu}/^{239}\text{Pu}$ ratios and their significance. Geochim. Cosmochim. Acta. 51, 2623-
429 2637.

430 Bu, W.T., Fukuda, M., Zheng, J., Aono, T., Ishimaru, T., Kanda, J., Yang, G.S., Tagami, K.,
431 Uchida, S., Guo, Q.J., Yamada, M., 2014. Release of Pu Isotopes from the Fukushima
432 Daiichi Nuclear Power Plant Accident to the Marine Environment Was Negligible. Environ.
433 Sci. Technol. 48, 9070-9078.

434 Chen, X.H., Li, T.G., Zhang, X.H., Li, R.H., 2013. A Holocene Yalu River-derived fine-grained
435 deposit in the southeast coastal area of the Liaodong Peninsula. Chin. J. Oceanol. Limn.
436 31(3), 636-647.

437 Corcho-Alvarado, J.A., Diaz-Asencio, M., Froidevaux, P., Bochud, F., Alonso-Hernández, C.M.,
438 Sanchez-Cabeza, J.A., 2014. Dating young Holocene coastal sediments in tropical regions:
439 use of fallout $^{239,240}\text{Pu}$ as alternative chronostratigraphic marker. Quat. Geochronol. 22, 1-10.

440 Dong W., Tims, S.G., Fifield, L.K., Guo, Q. J., 2010. Concentration and characterization of
441 plutonium in soils of Hubei in central China. J. Environ. Radioact. 101, 29-32.

442 Gatsaud, J., Povinec, P.P., Aoyama, M., Hirose, K., Sanchez-Cabeza J.A., Levy, I., Roos, P.,
443 Eriksson, M., Bosc, E., Rezzoug, S., 2011. Transport and scavenging of Pu in surface waters
444 of the Southern Hemisphere Oceans. Prog. Oceanogr. 89, 92-100.

445 Hain, K., Faestermann, T., Fimiani, L., Golser, R., Gómez-Guzmán, J.M., Korschinek, G.,
446 Kortmann, F., von Gostomski, C.L., Ludwig, P., Steier, P., Tazoe, H., Yamada, M., 2017.
447 Plutonium Isotopes ($^{239-241}\text{Pu}$) Dissolved in Pacific Ocean Waters Detected by Accelerator
448 Mass Spectrometry: No Effects of the Fukushima Accident Observed. *Environ. Sci. Technol.*
449 51, 2031-2037.

450 Hancock, G.J., Leslie, C., Everett, S.E., Tims, S.G., Brunskill, G.J., Haese, R., 2011. Plutonium
451 as a chronomarker in Australian and New Zealand sediments: a comparison with ^{137}Cs . *J.*
452 *Environ. Radioact.* 102, 919-929.

453 Hardy, E.P., Krey, P.W., Volchok, H.L., 1973. Global inventory and distribution of fallout
454 plutonium. *Nature.* 241, 444-445.

455 Hirose, K., Igarashi, Y., Aoyama, M., Miyao, T., 2001. Long-term trends of plutonium fallout
456 observed in Japan. In *Plutonium in the Environment*; Kudo, A., Ed.; Elsevier Science:
457 Amsterdam, pp 251-266.

458 Hong, G.H., Chung, C.S., Lee, S.H., Kim, S.H., Baskaran, M., Lee, H.M., Kim, Y.I., Yang, D.B.,
459 Kim, C.K., 2006. Artificial radionuclides in the Yellow Sea: inputs and redistribution.
460 *Radioactivity in the Environment-International Conference on Isotopes in Environmental*
461 *Studies.* 8, 96-133.

462 Hong, G.H., Hamilton, T.F., Baskaran, M., Kenna, T.C., 2011. Applications of anthropogenic
463 radionuclides as tracers to investigate marine environmental processes. In *Handbook of*
464 *Environmental Isotope Geochemistry, Advances in Isotope Geochemistry*; Baskaran, M.,
465 Ed.; Springer-Verlag: Berlin Heidelberg, pp 367-394.

466 Huh, C.A., Su, C.C., 1999. Sedimentation dynamics in the East China Sea elucidated from ^{210}Pb ,
467 ^{137}Cs , and $^{239,240}\text{Pu}$. *Mar. Geol.* 160, 183-196.

468 Kelley, J.M., Bond, L.A., Beasley, T.M., 1999. Global distribution of Pu isotopes and ^{237}Np . *Sci.*
469 *Total Environ.* 237/238, 483-500.

470 Ketterer, M.E., Hafer, K.M., Jones, V.J., Appleby, P.G., 2004. Rapid dating of recent sediments
471 in Loch Ness: inductively coupled plasma mass spectrometric measurements of global
472 fallout plutonium. *Sci. Total Environ.* 322, 221-229.

473 Koide, M., Bertine, K.K., Chow, T.J., Goldberg, E.D., 1985. The $^{240}\text{Pu}/^{239}\text{Pu}$ ratio, a potential
474 geochronometer. *Earth Planet. Sci. Lett.* 72, 1-8.

475 Kim, C.K., Kim, C.S., Chang, B.U., Choi, S.W., Chung, C.S., Hong, G.H., Hirose, K., Igarashi,
476 Y., 2004. Plutonium isotopes in seas around the Korean Peninsula. *Sci. Total Environ.* 318,
477 197-209.

478 Lee, M.H., Clark, S.B., 2005. Activities of Pu and Am isotopes and isotopic ratios in a soil
479 contaminated by weapons grade plutonium. *Environ. Sci. Technol.* 39, 5512-5516.

480 Lee, S.H., Gastaud, J., Povinec, P.P., Hong, G.H., Kim, S.H., Chung, C.S., Lee, K.W., Pettersson,
481 H.B.L., 2003. Distribution of plutonium and americium in the marginal seas of the northwest
482 Pacific Ocean. *Deep-Sea Research Part II.* 50, 2727-2750.

483 Li, X.Y., Zhao, G.M., Li, B.Y., 2010. Analysis of Changing Situation of Land Resources in
484 Northeast China. *Res. Soil Water Conserv.* 17 (5), 68-74. (in Chinese).

485 Lindahl, P., Lee, S.H., Worsfold, P., Keith-Roach, M., 2010. Plutonium isotopes as tracers for
486 ocean processes: A review. *Mar. Environ. Res.* 69, 73-84.

487 Liu, Z.Y., Zheng, J., Pan, S.M., Dong, W., Yamada, M., Aono, T., Guo, Q.J., 2011. Pu and ^{137}Cs
488 in the Yangtze River Estuary sediments: distribution and source identification. *Environ. Sci.*
489 *Technol.* 45 (5), 1805-1811.

490 Liu, Z.Y., Zheng, J., Pan, S.M., Gao, J.H., 2013. Anthropogenic plutonium in the North Jiangsu
491 tidal flats of the Yellow Sea in China. *Environ. Monit. Assess.* 185, 6539-6551.

492 Muramatsu, Y., Rühm, W., Yoshida, S., Tagami, K., Uchida, S., Wirth, E., 2000. Concentrations
493 of ^{239}Pu and ^{240}Pu and their isotopic ratios determined by ICP-MS in soils collected from the
494 Chernobyl 30-km zone. *Environ. Sci. Technol.* 34, 2913-2917.

495 Muramatsu, Y., Hamilton, T., Uchida, S., Tagami, K., Yoshida, S., Robison, W., 2001.
496 Measurement of $^{240}\text{Pu}/^{239}\text{Pu}$ isotopic ratios in soils from the Marshall Islands using ICP-MS.
497 *Sci. Total Environ.* 278, 151-159.

498 Nagaya, Y., Nakamura, K., 1992. $^{239,240}\text{Pu}$ and ^{137}Cs in the east China and the Yellow seas. *J.*
499 *Oceanogr.* 48, 23-35.

500 Nakano, M., Povinec, P.P., 2003. Modelling the distribution of plutonium in the Pacific Ocean. *J.*
501 *Environ. Radioact.* 69, 85-106.

502 Oikawa S., Watabe T., Takata H., Misonoo J., Kusakabe M. 2015. Plutonium isotopes and ^{241}Am
503 in surface sediments off the coast of the Japanese islands before and soon after the
504 Fukushima Dai-ichi nuclear power plant accident. *J. Radioanal. Nucl. Chem.* 303, 1513-
505 1518.

506 Pan, S.M., Tims, S.G., Liu, X.Y., Fifield, L.K., 2011. ^{137}Cs , $^{239+240}\text{Pu}$ concentrations and the
507 $^{240}\text{Pu}/^{239}\text{Pu}$ atom ratio in a sediment core from the sub-aqueous delta of Yangtze River
508 estuary. *J. Environ. Radioact.* 102(10), 930-936.

509 Qi, J., Li, F.Y., Song, J.M., Gao, S., Wang, G.Z., Cheng, P., 2004. Sedimentation rate and flux of
510 the North Yellow Sea. *Marine Geology & Quaternary Geology.* 24(2), 9-14(in Chinese).

511 Qiao, J.X., Hou, X.L., Roos, P., Miró, M., 2009. Rapid Determination of plutonium isotopes in
512 environmental samples using sequential injection extraction chromatography and detection
513 by inductively coupled plasma mass spectrometry. *Anal. Chem.* 81, 8185-8192.

514 State Oceanic Administration, 1998. *Chinese Harbours and Embayments (Volume 14):*
515 *Important Estuaries.* China Ocean Press, Beijing, pp 386-432 (in Chinese).

516 Su, CC., Huh, C.A., 2002. ^{210}Pb , ^{137}Cs , and $^{239,240}\text{Pu}$ in East China Sea sediments: Sources,
517 pathways and budgets of sediments and radionuclides. *Mar. Geol.* 183, 163-178.

518 UNSCEAR, 1982. *Ionizing Radiation: Sources and Effects; United Nations Scientific Committee*
519 *on the Effects of Atomic Radiation Exposures to the Public from Man-made Sources of*
520 *Radiation; United Nations: New York, pp 211.*

521 UNSCEAR, 2000. *Sources and effects of ionizing radiation. United Nations scientific committee*
522 *on the effects of atomic radiation exposures to the public from man-made sources of*
523 *radiation; United Nations, New York, pp 214-215.*

524 Wang, J.L., Baskaran, M., Hou, X.L., Du, J.Z., Zhang, J., 2017. Historical changes in ^{239}Pu and
525 ^{240}Pu sources in sedimentary records in the East China Sea: Implications for provenance and
526 transportation. *Earth Planet. Sci. Lett.* 466, 32-42.

527 Wang, Z.L., Yamada, M., 2005. Plutonium activities and $^{240}\text{Pu}/^{239}\text{Pu}$ atom ratios in sediment
528 cores from the East China Sea and Okinawa Trough: Sources and inventories. *Earth Planet.*
529 *Sci. Lett.* 233, 441-453.

530 Warneke, T., Croudace, I.W., Warwick, P.E., Taylor, R.N., 2002. A new ground-level fallout
531 record of uranium and plutonium isotopes for northern temperate latitudes. *Earth Planet. Sci.*
532 *Lett.* 203, 1047-1057.

533 Wu, F.C., Zheng, J., Liao, H.Q., Yamada, M., 2010. Vertical distributions of plutonium and
534 ^{137}Cs in lacustrine sediments in northwestern China: quantifying sediment accumulation
535 rates and source identifications. *Environ. Sci. Technol.* 44, 2911-2917.

536 Wu, F.C., Zheng, J., Liao, H.Q., Yamada, M., Wan, G.J., 2011. Anomalous plutonium isotopic
537 ratios in sediments of Lake Qinghai from the Qinghai-Tibetan Plateau, China. *Environ. Sci.*
538 *Technol.* 45, 9188-9194.

539 Wu, J.W., Zheng, J., Dai, M.H., Huh, C.A., Chen, W., Tagami, K., Uchida, C., 2014. Isotopic
540 composition and distribution of plutonium in Northern South China Sea sediments revealed
541 continuous release and transport of Pu from the Marshall Islands. *Environ. Sci. Technol.* 48,
542 3136-3144.

543 Xu, Y.H., Qiao, J.X., Hou X.L., Pan, S.M., 2013. Plutonium Isotopes in Soils from Northeast
544 China and Its Potential Application for Evaluation of Soil Erosion. *Sci. Rep.* 3, 3506; DOI:
545 10.1038/SREP03506.

546 Xu, Y.H., Qiao, J.X., Hou X.L., Pan, S.M., Roos, P., 2014. Determination of Plutonium Isotopes
547 (^{238}Pu , ^{239}Pu , ^{240}Pu and ^{241}Pu) in Environmental Samples Using Radiochemical Separation
548 Combined with Radiometric and Mass Spectrometric Measurements. *Talanta.* 119, 590-595.

549 Xu, Y.H., Qiao, J.X., Pan, S.M., Hou, X.L., Roos, P., Cao, L.G., 2015. Plutonium as a tracer for
550 soil erosion assessment in Northeast China. *Sci. Total Environ.* 511, 176-185.

551 Xu, Y.H., Pan, S.M., Wu, M.M., Zhang, K.X., Hao, Y.P., 2017. Association of Plutonium
552 isotopes with natural soil particles of different size and comparison with ^{137}Cs . *Sci. Total*
553 *Environ.* 581-582, 541-549.

554 Zheng, J., Tagami, K., Watanabe, S., Uchida, S., Aono, T., Ishii, N., Yoshida, S., Kubota, Y.,
555 Fuma, S., Ihara, S., 2012. Isotopic evidence of plutonium release into the environment from
556 the Fukushima DNPP accident. *Sci. Rep.* 2, 304 DOI: 10.1038/srep00304.

557 Zheng, J., Wu, F.C., Yamada, M., Liao, H.Q., Liu, C.Q., Wan, G.J., 2008. Global fallout Pu
558 recorded in lacustrine sediments in Lake Hongfeng, SW China. *Environ. Pollut.* 152, 314-
559 321.

560 Zheng, J., Yamada, M., 2006. Plutonium isotopes in settling particles: Transport and scavenging
561 of Pu in the western northwest pacific. *Environ. Sci. Technol.* 40, 4103-4108.

562

563 **Figure captions**

564 Figure 1 a) Overview map of the Yellow Sea with the principal currents, plotted after Huh & Su,
565 1999 and Su & Huh, 2002. b) Detail of the Yellow Sea and its coasts with the location of the
566 study area being highlighted by a red rectangle. c) Detailed sampling sites in the North
567 Yellow Sea. [The red oval dotted frame indicates the location of the Yalu River-derived](#)
568 [clinoform according to Chen et al. \(2013\).](#) YSWC: Yellow Sea Warm Current; YSCC:
569 Yellow Sea Coastal Current; LCC: Liaonan Coastal Current; WKCC: West Korea Coastal
570 Current.

571 Figure 2 Distribution of $^{240}\text{Pu}/^{239}\text{Pu}$ atomic ratios in surface sediments and sediment cores in the
572 northern NYS. [The red dashed lines indicate the global fallout value of \$-0.18 \pm 0.02\$.](#)

573 Figure 3 Distribution of the $^{239+240}\text{Pu}$ activities in all surface sediments (including the surface
574 layer (0-2 cm) of the two sediment cores NYSC-01 and NYSC-02) along the southeast coast
575 of the Liaodong Peninsula. [The red oval dotted frame indicates the location of the Yalu](#)
576 [River-derived clinoform according to Chen et al. \(2013\).](#)

577 Figure 4 Vertical profiles of $^{239+240}\text{Pu}$ activities in sediment cores collected from the northern
578 NYS.

579 Figure 5 Abundance of annual deposition of ^{137}Cs in northern hemisphere and abundance of
580 deposition of $^{239+240}\text{Pu}$ in each layer of sediments. The annual deposition data of ^{137}Cs in
581 northern hemisphere were cited from the UNSCEAR (2000).⁵ Data of Pu in Lake Hongfeng
582 and Qinghai were respectively cited from Zheng et al.(2008)¹⁰ and Wu et al.(2011)¹¹ and
583 mainly originated from the global fallout direct deposition.

584 Figure 6 a) Comparison of $^{239+240}\text{Pu}$ inventories between the northern NYS and the Yangtze
585 River estuary. $^{239+240}\text{Pu}$ inventories for $\sim 31^\circ\text{N}$ and $\sim 39^\circ\text{N}$ were calculated as the averages of
586 the data reported respectively by Dong et al.(2010)⁵¹ and by Xu et al.(2013, 2015)^{31,33}-and
587 Hardy et al.(1973)⁵⁰ as shown in Table A.S3. b) Comparison of the relative contributions of
588 Pu from the direct global fallout deposition and riverine input between the northern NYS and
589 the Yangtze River estuary.

590

591

592

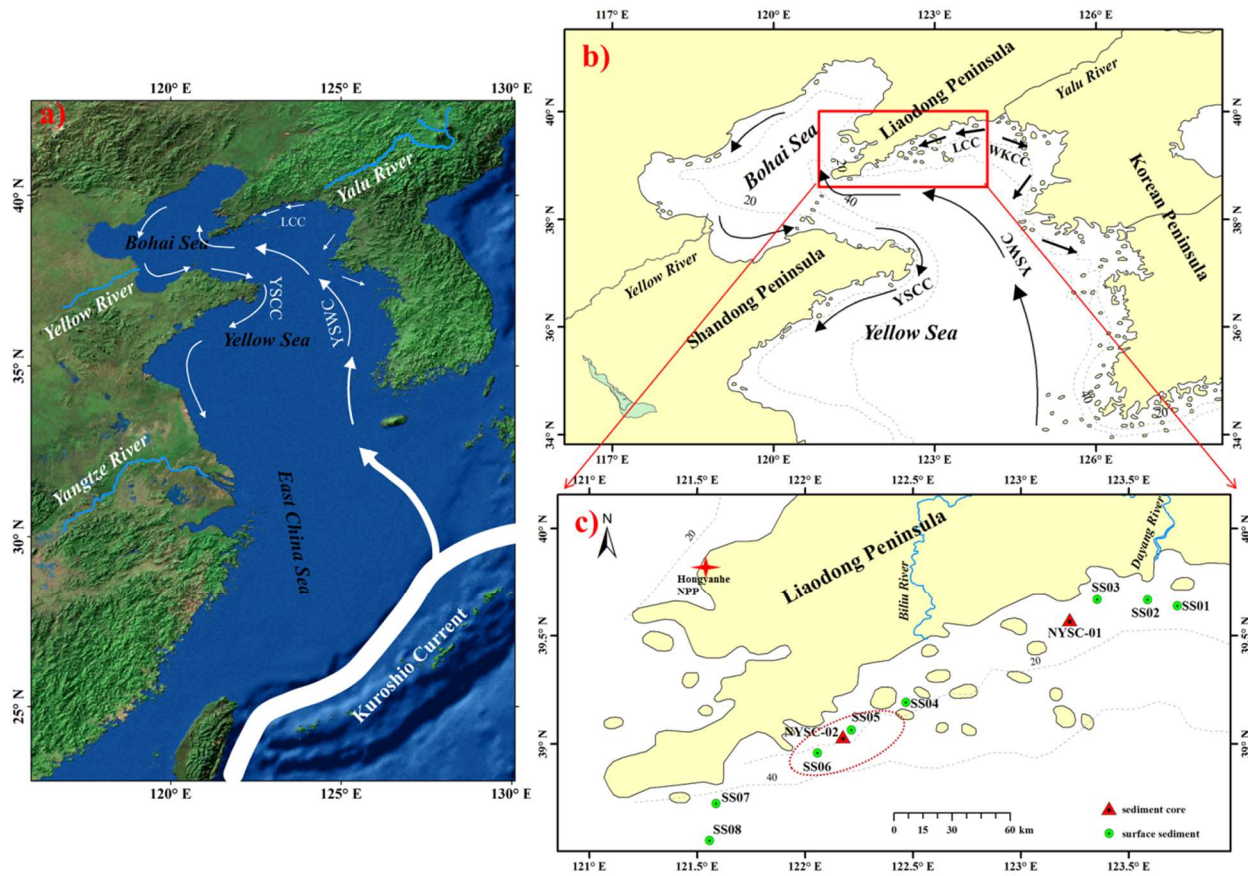
593

594

595

596 Figure 1

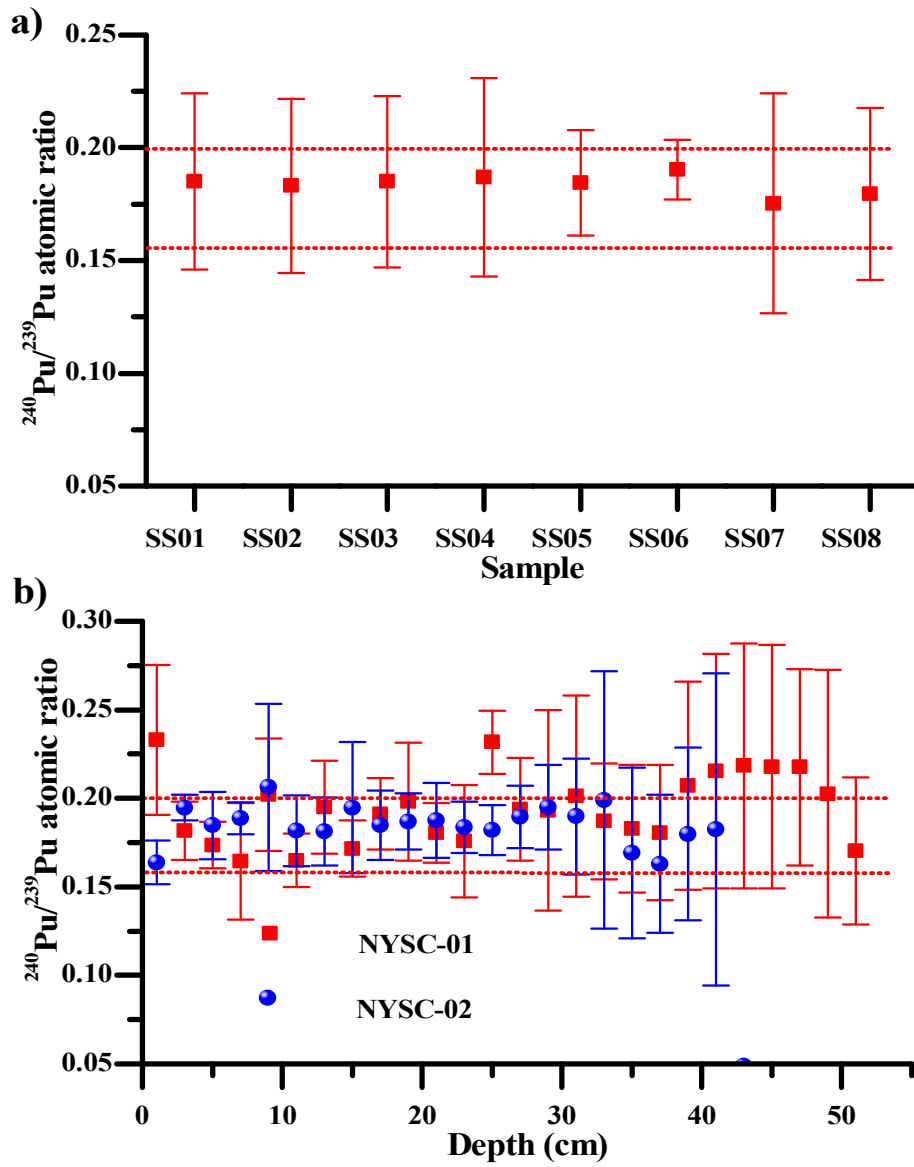
597



598

599

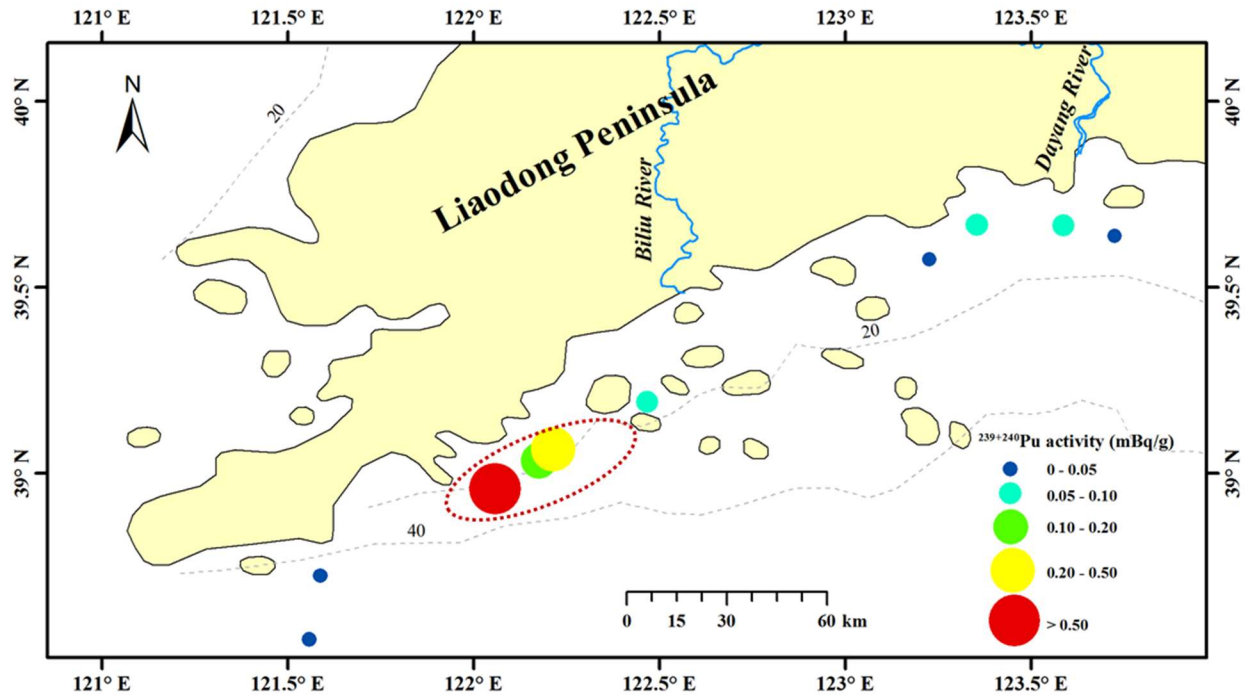
600 Figure 2



601

602

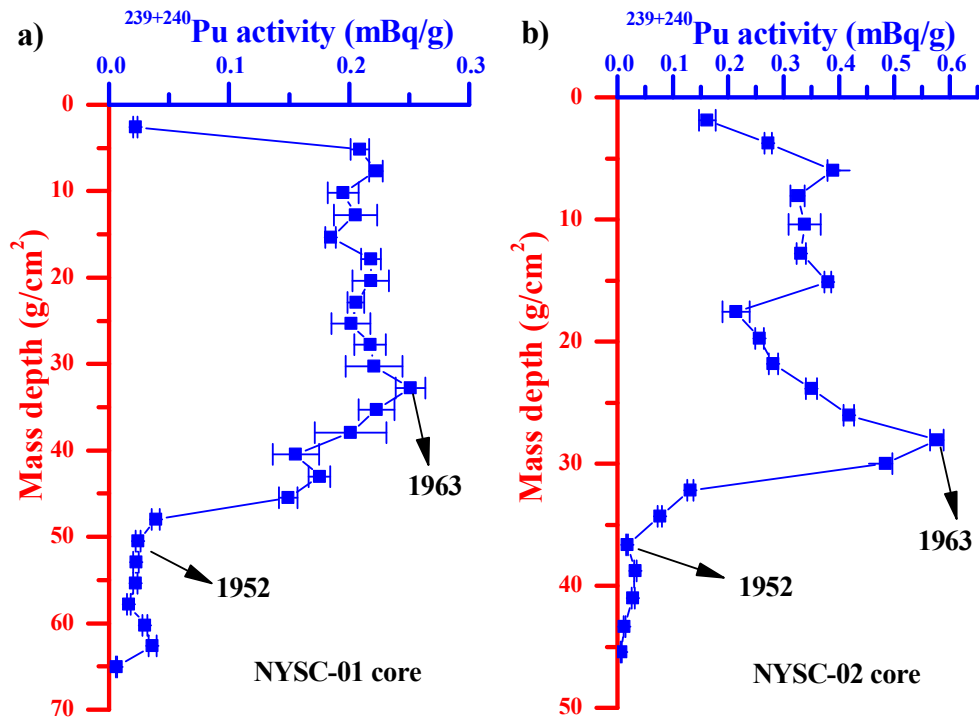
603 Figure 3



604

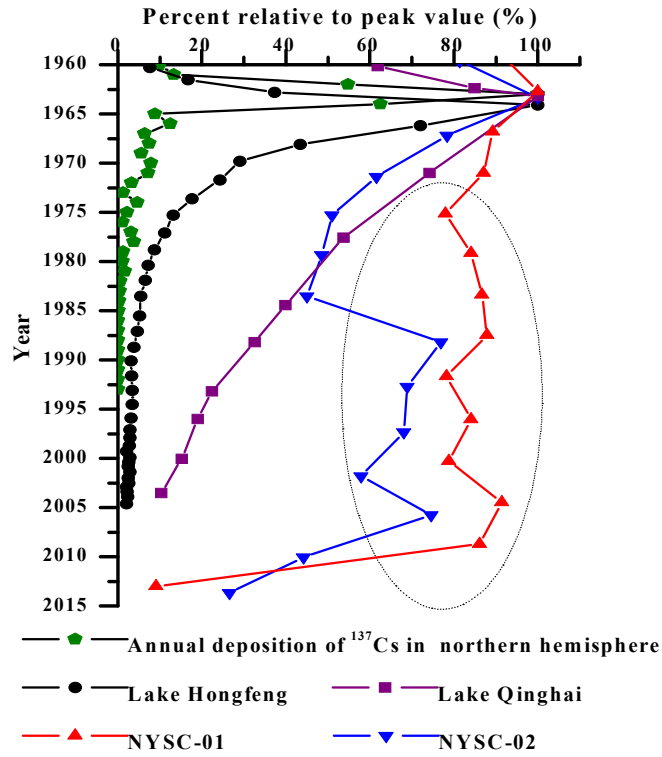
605

606 Figure 4



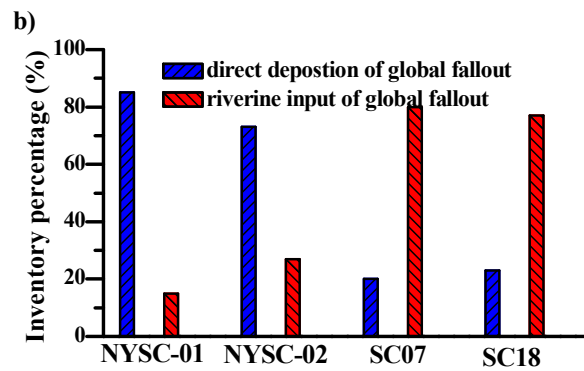
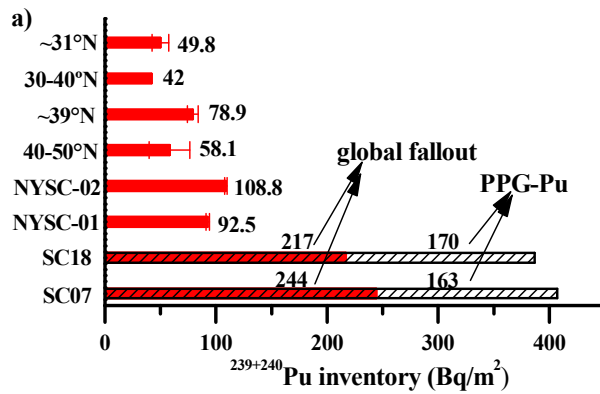
607

608 Figure 5



609

610 Figure 6



611

