



Costs and benefits of predator-induced defence in a toxic diatom

Olesen, Anna J.; Ryderheim, Fredrik; Krock, Bernd; Lundholm, Nina; Kiørboe, Thomas

Published in:
Proceedings of the Royal Society B: Biological Sciences

Link to article, DOI:
[10.1098/rspb.2021.2735](https://doi.org/10.1098/rspb.2021.2735)

Publication date:
2022

Document Version
Peer reviewed version

[Link back to DTU Orbit](#)

Citation (APA):
Olesen, A. J., Ryderheim, F., Krock, B., Lundholm, N., & Kiørboe, T. (2022). Costs and benefits of predator-induced defence in a toxic diatom. *Proceedings of the Royal Society B: Biological Sciences*, 289, Article 20212735. <https://doi.org/10.1098/rspb.2021.2735>

General rights

Copyright and moral rights for the publications made accessible in the public portal are retained by the authors and/or other copyright owners and it is a condition of accessing publications that users recognise and abide by the legal requirements associated with these rights.

- Users may download and print one copy of any publication from the public portal for the purpose of private study or research.
- You may not further distribute the material or use it for any profit-making activity or commercial gain
- You may freely distribute the URL identifying the publication in the public portal

If you believe that this document breaches copyright please contact us providing details, and we will remove access to the work immediately and investigate your claim.

1 **Costs and benefits of predator-induced defence in a toxic diatom**

2

3 Anna J. Olesen^{1*}, Fredrik Ryderheim^{2*}, Bernd Krock³, Nina Lundholm¹, and Thomas
4 Kiørboe²

5

6 ¹ Natural History Museum of Denmark, University of Copenhagen, Øster Farimagsgade 5,
7 1353 Copenhagen K, Denmark.

8 ² Centre for Ocean Life, DTU Aqua, Technical University of Denmark, Kemitorvet, Building
9 202, 2800 Kgs. Lyngby, Denmark

10 ³ Alfred Wegener Institut-Helmholtz Zentrum für Polar- und Meeresforschung, Chemische
11 Ökologie, Am Handelshafen 12, 27570 Bremerhaven, Germany.

12

13 * These authors contributed equally to this work.

14

15 We declare no competing interests.

16 **Abstract**

17 Phytoplankton employ a variety of defence mechanisms against predation, including
18 production of toxins. Domoic acid (DA) production by the diatom *Pseudo-nitzschia* spp. is
19 induced by the presence of predators and is considered to provide defence benefits, but the
20 evidence is circumstantial. We exposed eight different strains of *P. seriata* to chemical cues
21 from copepods and examined the costs and the benefits of toxin production. The magnitude
22 of the induced toxin response was highly variable among strains, while the costs in terms of
23 growth reduction per DA cell quota were similar and the trade-off thus consistent. We found
24 two components of the defence in induced cells: (1) a ‘private good’ in terms of elevated
25 rejection of captured cells, and (2) a ‘public good’ facilitated by a reduction in copepod
26 feeding activity. Induced cells were more frequently rejected by copepods and rejections
27 were directly correlated with DA cell quota and independent of access to other food items. In
28 contrast, the public-good effect was diminished by the presence of alternative prey suggesting
29 that it does not play a major role in bloom formation and that its evolution is closely
30 associated with the grazing-deterrent private good.

31

32 Keywords: Diatom defence, copepods, marine chemical interactions, defensive benefit, trade-
33 offs.

34 ***Introduction***

35 The herbivory pressure in pelagic ecosystems is intense, with up to three times higher fraction
36 of the production being consumed compared to terrestrial systems [1]. In response,
37 phytoplankton have evolved numerous defence mechanisms, varying from morphological and
38 behavioural to physiological and including the production of various chemical substances [2].
39 Through the evolution of costly defence mechanisms, grazing may be an important driver of
40 diversity in phytoplankton communities by allowing coexistence of defence and competition
41 specialists [3–5]. Coexistence requires that defence benefits comes with a cost, e.g., in the form
42 of reduced growth [6], but both benefits and costs of defences in phytoplankton have rarely
43 been convincingly demonstrated [2].

44 Toxic algal blooms are a global phenomenon. Often, the toxins accumulate in the food web
45 and may negatively affect several trophic levels [7]. Diatoms of the cosmopolitan genus
46 *Pseudo-nitzschia* can produce the neurotoxin domoic acid (DA) and form dense toxic blooms
47 with severe consequences to marine organisms and ecosystems, like mass death incidents of
48 seabirds and marine mammals [7,8]. However, the potential benefits and costs of DA
49 production have not been established. There is conflicting evidence on whether DA production
50 is costly; some studies have suggested that cells with increased toxin content have lower
51 growth rates [9,10], but other experiments have been unable to reproduce these results [11–
52 13]. Similarly, there is only circumstantial evidence that DA acts as a predator-deterrent. Some
53 studies have demonstrated reduced predator activity in zooplankton exposed to toxic *Pseudo-*
54 *nitzschia* [11,14]. Whether this was due to predators becoming intoxicated after consuming
55 some cells or being exposed to dissolved DA [15], or whether the grazers actively avoid eating
56 toxic cells is in all cases unclear (a toxic vs a deterrent effect). However, toxic effects are ‘public
57 goods’ [16] that also benefit non-toxic competitors, and ‘cheaters’ that do not pay the probable
58 cost of producing the toxin may outcompete the producers. Evolution of such public goods,

59 unless associated with private benefits, is not evolutionary supported. In contrast, the evolution
60 of toxin production as a ‘private good’, where only the cell producing the toxin benefits from
61 it, is easy to envisage [16,17].

62 Further evidence of trade-offs and a defensive value of DA is that its production is induced by
63 chemical cues from copepods, copepodamides [9,18]. Defences are expected to be inducible
64 only if threats are varying, and the defence is costly [19]. If defences did not have associated
65 costs, all organisms would evolve towards an equal state of defence [20].

66 Here, we explore the defensive value and trade-offs of DA-production in *P. seriata* exposed to
67 different concentrations of toxin-inducing copepod cues. We include eight different strains to
68 account for possible intraspecific variation. We use direct video observations of individual
69 copepod-prey interactions to quantify the defensive benefits of DA in terms of captured cells
70 being consumed or rejected, and we quantify the costs of DA-production from changes in cell
71 division rate. We demonstrate significant benefits and costs of DA-production and thus a clear
72 trade-off. We further demonstrate threefold differences in growth rate among strains, but that
73 the trade-off of DA-production is the same. We finally show that DA-production provides both
74 private and public goods, thus suggesting a potential mechanism for the evolution of apparently
75 public goods.

76 ***Materials and methods***

77 **Culturing and species identification**

78 To establish cultures of *Pseudo-nitzschia* for the experiments, live phytoplankton samples were
79 collected in Øresund, Denmark (55°45’40’’N 12°36’0’’E), in April 2020, by hauling a 20- μ m
80 plankton net through the upper (0–5 m) water column. Single cells or chains were micro-
81 pipetted and placed individually into 96-well plates with L1 medium [21] using a light
82 microscope (CKX53, Olympus, Tokyo Japan). The cultures were kept at 4 °C at a 16:8 h

83 light:dark cycle and $120 \mu\text{mol photons m}^{-2} \text{ s}^{-1}$ using cool white light bulbs. Identification was
84 performed by transmission electron microscopy (JEOL, JEM 1010, Tokyo, Japan) following
85 Hasle & Lundholm [22]. All strains included in this study were identified as *P. seriata* (Table
86 S1, S2).

87 **Induction and cost of toxin production**

88 To quantify the costs of toxin production, we performed dose-response experiments with eight
89 strains of *P. seriata*, where we measured the growth rate as a function of copepodamide
90 concentration and consequent cellular contents of DA. Cells of different strains were suspended
91 in L1 medium and added to 200 mL cell-culture flasks. The flasks were exposed to
92 copepodamides extracted from freeze-dried *Calanus* spp. [23] by coating the inside of the wall
93 with a mixture of copepodamides dissolved in methanol. The methanol was evaporated using
94 N_2 gas. Controls underwent the same treatment but received methanol without copepodamides.
95 Six strains of *P. seriata* were exposed to 0, 10, or 50 pM of copepodamides, while two strains
96 were exposed to 0 or 50 pM. The flasks were incubated for 72 h at 4 °C or 8 °C (only strain
97 SKC620) and ca $100 \mu\text{mol photons m}^{-2} \text{ s}^{-1}$ on a 14:10 light:dark cycle. Due to slow release and
98 quick degradation, the copepodamides coated on the inside of the bottles typically yield an
99 actual concentration of 1% of the added amount after 12 h, and much less after 72 h [18]. After
100 72 h incubation, the cells were harvested for enumeration of cell densities and quantification
101 of cellular DA content. Cell densities were determined by fixing three mL of sample in acidic
102 Lugol's solution (final concentration 2–3%) and counting a minimum of 400 cells in a
103 Sedgewick-Rafter chamber using an inverted microscope. Cell growth rates were calculated
104 using temporal variations in cell densities assuming exponential growth. Average cell size was
105 determined by measuring 25–30 random cells and cell volume calculated based on Lundholm
106 et al. [24]. For five of the strains, we also measured dissolved DA. Additional cells from the

107 two strains exposed to only 0 or 50 pM were harvested for use in copepod foraging
108 experiments.

109 **Copepod foraging experiments**

110 We directly observed the feeding response of copepods to the cells by video-recording
111 individual copepod-cell interactions. The feeding experiments were done in a constant
112 temperature room (8 °C) in darkness. Adult females of the feeding-current feeding
113 copepod *Temora longicornis* were glued by their dorsal surface to a human hair [25,26] and
114 starved overnight in darkness. The other end of the hair was attached to a micromanipulator
115 and the copepod was submerged in a 10×10×10 cm³ aquarium with filtered seawater (salinity
116 27). A different copepod was used for each replicate in each treatment and experiment (three
117 copepods per treatment). *Pseudo-nitzschia* cells were added to the aquarium and the copepod
118 response was immediately recorded using a high-speed camera (Phantom V210, Vision
119 Research, New Jersey, USA). The camera was equipped with lenses to yield a field of view of
120 about 1.3 x 1.0 mm². A magnetic stirrer gently mixed the water and kept the cells in suspension.

121 We used two of the strains in the feeding experiments (SKC620 and AI420). The copepods
122 were fed either non-induced control cells or cells that had been exposed to 50 pM of
123 copepodamides for 72 hours, as described above, to yield different toxin cell quotas. We
124 conducted three types of experiments: (1) In feeding experiments with strain SKC620, we
125 recorded four five-minute (0–5 min, 20–25 min, 35–40 min, 55–60 min after adding cells)
126 sequences at 50 frames per second (fps) at a cell density of 200 cells mL⁻¹. (2) In a similar
127 feeding experiment with strain AI420, we recorded six three-minute (0–3 min, 10–13 min, 20–
128 23 min, 30–33, 40–43, 50–53) sequences at 100 fps with the same cell density as above. (3) In
129 a third experiment, we mixed toxic *P. seriata* (strain AI420) with the non-toxic diatom *Ditylum*
130 *brightwellii*, 100 cells mL⁻¹ of each species, and recorded sequences (at 100 fps) as above. The

131 two species were easy to distinguish in the videos. The video recordings were analysed to
132 quantify copepod beating activity and frequency, prey capture, and the fraction of captured
133 cells that were rejected by the copepods. Additional videos of five strains were recorded using
134 the same set-up but with 1000 fps to examine potential differences in prey-handling time.

135 **Toxin analysis**

136 In all experiments, 45 or 60 mL of each replicate was harvested and divided into 15 mL
137 centrifuge tubes and centrifuged at $760 \times g$ for 15 minutes. The supernatants of five strains
138 (B2N, E9N, D2N, G3N, and H4N) were kept in 15 mL falcon tubes and stored frozen at -20
139 $^{\circ}\text{C}$ for analysis of dissolved domoic acid. For intracellular toxins, cell pellets were pooled in 2
140 mL cryotubes and centrifuged again at $1507 \times g$ for 15 minutes. The pellets were stored frozen
141 at -20 $^{\circ}\text{C}$. Toxins were extracted and analysed using liquid chromatography coupled with
142 tandem mass spectrometry [27].

143 All raw data are available in Olesen et al. [28].

144 **Statistical analyses**

145 The effect of the copepodamide treatment on the fraction of time spent beating was analysed
146 by a mixed-effects logistic regression using the *lme4* R-package [29]. ‘Treatment’ and
147 ‘Sequence’ were used as fixed effects and ‘Replicate’ as the random effect. Pairwise
148 comparisons between the different treatments (control – induced, or control – induced – mix)
149 were done using the Satterthwaite degrees-of-freedom method. The relationship between
150 growth rate or prey-handling time and cellular DA content was analysed using a linear mixed-
151 effects model (again using *lme4*) with ‘Strain’ as the random effect. All statistical models were
152 validated by visual inspection of residual plots.

153 **Results**

154 **Induction and cost of toxin production**

155 Exposure to copepodamides increased the cellular toxin content in *P. seriata*, but there was a
156 large variation among strains in terms of the magnitude of the response (Fig. 1). The relative
157 increase of cellular DA content in the induced treatments compared to the non-induced control
158 was between 40% and 35 000% (Fig. 1, Table S1). In the five tested strains (B2N, D2N, E9N,
159 G3N, and H4N), cellular and dissolved domoic acid were not significantly correlated (linear
160 regression, $F_{5,9} = 0.6$, $p = 0.72$).

161 Growth rates generally decreased significantly with increasing toxin content (Fig. 2, Table S1).
162 While the intercepts differed significantly among strains, suggesting a threefold variation in
163 growth rate, the slopes did not differ significantly from one another. Thus, the cost of DA is
164 the same for all the strains (Fig. 2).

165 **Copepod foraging response**

166 *Temora longicornis* copepods use their feeding appendages to create a feeding current from
167 which prey particles are individually perceived and captured. The copepods react to *Pseudo-*
168 *nitzschia* cells that are close to or touching the copepod feeding appendages. Depending on cell
169 orientation, the copepod briefly adjusts the captured particle and brings the cell or chain to the
170 mouth (Online videos S1–2). Prey handling by the feeding appendages concludes with the cell
171 either being ingested or rejected (Online videos S1–2). The time spent handling the ingested
172 particles increased significantly with chain length (Fig. S1) but was independent of cellular DA
173 content.

174 The two strains used in the foraging experiments both responded to the copepodamides by
175 increasing their cellular DA content (Fig. 1g, h; Two-way ANOVA, $F_{3,8} = 29.0$, $p = 0.01$), and
176 a significantly larger fraction of the induced than the control cells caught by the copepods was
177 rejected (Fig. 3a, b, c). Overall, the fraction of rejected cells increased with the cellular DA

178 content (Fig. 3c). The addition of the alternative prey (*D. brightwellii*) in the mixed experiment
179 with strain AI420 did not affect the fraction of *P. seriata* cells rejected (Fig. 3a). The *D.*
180 *brightwellii* cells were rejected at a rate like that of the non-induced *P. seriata* (Fig. 3a).

181 The copepods were initially beating their appendages 100% of the time irrespectively of
182 treatment (Fig. 4a, c). However, within minutes, exposure to the induced *P. seriata* cells
183 (AI429) significantly reduced the fraction of time that the copepods spent creating a feeding
184 current (Fig. 4a, b). Beating time was, however, not different from the controls when the
185 induced treatment was fed to the copepod in conjunction with *D. brightwellii* (Fig. 4a, b). When
186 exposed to induced *P. seriata* SKC620, a similar trend was observed (Fig. 4c, d), but the
187 difference was not statistically significant.

188 ***Discussion***

189 Our results show for the first time a convincing enhancement of survival diatoms with
190 induced toxin production when preyed upon by copepods. The survival strategy comes with
191 the costs of a significant decrease in growth rate.

192 **The defensive benefit of domoic acid**

193 The copepods individually examined captured prey before deciding to ingest or reject it. This
194 mechanism is similar to that of copepods feeding on toxic dinoflagellates [30], but it is still
195 unclear how the copepods sense the toxicity of phytoplankton cells without, presumably,
196 harming them.

197 The increased odds of cell rejection in the induced treatment is a ‘private good’ that only cells
198 that produce the toxin benefit from [16]. This is similar to the effect of toxins produced by
199 *Alexandrium* spp. [26,30]. However, in addition, we find that the feeding activity of the
200 copepods was reduced when offered only induced *P. seriata*. This would be a ‘public good’

201 [16] because the cells that are not producing the costly toxins also benefit. This effect may be
202 due to a harmful effect of the dissolved DA, which has been found to reduce feeding in krill
203 [15], or it may be due to some of the toxic cells being eaten and intoxicating the copepod. The
204 latter may explain why some time elapses before the effect is noticeable [11,25]. Public goods
205 are in most cases not considered evolutionary stable strategies [16,31]. However, if associated
206 with a private good, as suggested here, one can envisage a mechanism for their evolution. The
207 public-good effect is diminished by the presence of alternative prey and would thus only
208 become ‘public’ in the later stages of a bloom when toxic cells presumably dominate, and thus
209 not promote bloom formation [32].

210 The private-good defence is measured as the fraction of captured cells that are rejected by the
211 copepod. Up to 75% of induced cells were rejected after capture; hence, toxin production may
212 increase the fitness of the cell provided that the costs are less than the benefits. While the
213 absolute cost of toxin production is similar among the examined strains, the relative cost is
214 much larger for slowly growing strains. Thus, in slowly growing SKC620 the relative reduction
215 in growth rate of induced cells, ~75%, is of the same magnitude as the 75% reduction in
216 predation risk, while in fast growing strains the benefits outweigh the costs. This suggests that
217 there are strong trade-offs related to variation in growth rate. Otherwise, fast growing strains
218 with a very favourable cost-benefit trade-off for toxin production would rapidly outcompete
219 slowly growing cells with a very limited fitness benefit of toxin production.

220 The presence of an alternative prey may influence copepod prey selection and contribute to the
221 relative increase of toxic cells [33,34]. However, capture rate (normalized by cell density) of
222 *P. seriata*, and subsequently the potential grazing mortality, doubled with the addition of *D.*
223 *brightwellii* compared to when only toxic cells were present as prey because the copepods did
224 not reduce their beating activity. The addition of the *D. brightwellii* cells did not change the
225 fraction of induced *P. seriata* cells that were rejected. Thus, responding to predator presence

226 by increasing toxin production is still an adequate response by *P. seriata*, and presumably also
227 other toxic phytoplankton, in mixed prey suspensions. This is similar to the effect of prey
228 concentration on the rejection frequency of three species of centric diatoms that thicken their
229 shells in response to copepod cues [35]. A higher prey concentration increased the odds of
230 rejection in both defended and undefended cells, but the relative increased odds of rejection in
231 defended cells remained similar.

232 The inability of previous studies to show evidence of a private-good defence in *Pseudo-*
233 *nitzschia* spp may stem from the use of wild copepods, or as an artefact of the experimental
234 design (“black-box” incubations). The present study is the first to use direct observations of
235 copepod-*P. seriata* interactions [11,12,36]. Predators that are frequently exposed to the
236 presence of toxic cells may be more tolerant to the toxins [37,38], and the build-up of resistance
237 is quick, i.e., over a few generations [39]. Thus, collection of wild copepods, e.g., during or
238 towards the end of a bloom, may influence results. The copepods used in this study were
239 originally isolated from the same location as the *P. seriata* strains, but have since been kept in
240 culture for many generations leading, presumably, to a loss of tolerance to DA. However, our
241 results may still be ecologically relevant. *T. longicornis* generally produce ~6 generations
242 year⁻¹ [40,41], in contrast to the longer generation times of *Calanus* spp. used in several earlier
243 studies [42,43]. Jiang et al. [44] found that *Acartia tonsa* resistance to the toxic dinoflagellate
244 *Cochlodinium polykrikoides* was relaxed after just two generations of non-exposure. Thus, loss
245 of tolerance may be fast once toxic cell abundance declines [45,46], and the defence will be
246 efficient in time for the next blooming event.

247 **The cost of toxin production**

248 Defence trade-offs in phytoplankton are notably difficult to establish [2]. However, the fact
249 that many phytoplankton defences are inducible—they respond to predator-cues by

250 upregulating their defence—suggests that costs should be present [19,47]. We find here a clear
251 trade-off between predator-induced toxin production and prey growth rate, and the costs were
252 similar among the tested strains.

253 Several previous experiments have been unable to measure the costs of toxin production in
254 *Pseudo-nitzschia* [e.g., 10, 27]. Lundholm et al. [9] speculated that this could be due to low
255 levels of DA that are insufficient to imply measurable differences in growth rates. However,
256 the trade-off becomes clear in our experiments with the higher levels of DA induced by
257 copepodamide exposure (Fig. 2). Selander et al. [18] measured the *in situ* concentration of
258 copepodamides over the course of a year and found values ranging between 40 and 2000 fM.
259 Thus, the concentrations used in our experiments are within the natural range, and the reduction
260 in growth rate observed due to increased toxin production is ecologically relevant.

261 The direct metabolic cost of DA-synthesis (see Supplementary information) is low and just 1–
262 2% of the growth reduction can be accounted for by energy allocated directly to DA-production
263 (Figure S2). This suggests that other metabolic processes linked to the toxin production are
264 important. Phytoplankton exposed to predators up-or downregulate thousands of genes, e.g.,
265 genes related to signal transduction pathways, stress responses, and lipid and nitrogen
266 metabolism [48,49]. Such associated responses may be energetically costly and may account
267 for the growth reduction.

268 **Toxic blooms and ecosystem implications**

269 The formation of toxic algal blooms have implications to marine ecosystems [7,8]. The ultimate
270 cause of such blooms is still debated [50], but are frequently attributed to eutrophication. The
271 ‘private-good’ mechanism demonstrated here improve our comprehension of such toxic
272 blooms and the factors that allow them to form. A mechanistic understanding of this is key to
273 predicting their occurrence and how they may be affected by a changing climate.

274 ***Acknowledgements***

275 AJO and NL were supported by the Independent Research Fund Denmark through grand no.
276 9040-0028B. FR and TK were supported through The Centre for Ocean Life, a Villum Kahn
277 Rasmussen Centre for Excellence funded by the Villum Foundation. Additional support to TK
278 was received from the Gordon and Betty Moore Foundation through award no. 5479. BK was
279 in part financed by the Helmholtz-Gemeinschaft Deutscher Forschungszentren through the
280 research program “Changing Earth – Sustaining our Future” of the Alfred Wegener Institut,
281 Helmholtz Zentrum für Polar- und Meeresforschung. We are grateful to Jack Melbye for
282 maintaining copepods cultures, Caroline Weber for culturing *P. seriata*, Øjvind Moettrup for
283 providing access to TEM facilities, and Erik Selander and Josephine Grønning for
284 copepodamide extractions.

285 ***References***

- 286 1. Cyr H, Face ML. 1993 Magnitude and patterns of herbivory in aquatic and terrestrial
287 ecosystems. *Nature* **361**, 148–150. (doi:10.1038/361148a0)
- 288 2. Pančić M, Kiørboe T. 2018 Phytoplankton defence mechanisms: traits and trade-offs:
289 Defensive traits and trade-offs. *Biol. Rev.* **93**, 1269–1303. (doi:10.1111/brv.12395)
- 290 3. Leibold MA, Hall SR, Smith VH, Lytle DA. 2017 Herbivory enhances the diversity of
291 primary producers in pond ecosystems. *Ecology* **98**, 48–56. (doi:10.1002/ecy.1636)
- 292 4. Cadier M, Andersen KH, Visser AW, Kiørboe T. 2019 Competition–defense tradeoff
293 increases the diversity of microbial plankton communities and dampens trophic cascades.
294 *Oikos* **128**, 1027–1040. (doi:10.1111/oik.06101)
- 295 5. Ehrlich E, Kath NJ, Gaedke U. 2020 The shape of a defense-growth trade-off governs
296 seasonal trait dynamics in natural phytoplankton. *ISME J* **14**, 1451–1462.
297 (doi:10.1038/s41396-020-0619-1)
- 298 6. Winter C, Bouvier T, Weinbauer MG, Thingstad TF. 2010 Trade-Offs between
299 Competition and Defense Specialists among Unicellular Planktonic Organisms: the
300 “Killing the Winner” Hypothesis Revisited. *MMBR* **74**, 42–57.
301 (doi:10.1128/MMBR.00034-09)

- 302 7. Bates SS, Hubbard KA, Lundholm N, Montresor M, Leaw CP. 2018 *Pseudo-nitzschia*,
303 *Nitzschia*, and domoic acid: New research since 2011. *Harmful Algae* **79**, 3–43.
304 (doi:10.1016/j.hal.2018.06.001)
- 305 8. McCabe RM *et al.* 2016 An unprecedented coastwide toxic algal bloom linked to
306 anomalous ocean conditions. *Geophys. Res. Lett.* **43**. (doi:10.1002/2016GL070023)
- 307 9. Lundholm N *et al.* 2018 Induction of domoic acid production in diatoms—Types of
308 grazers and diatoms are important. *Harmful Algae* **79**, 64–73.
- 309 10. Pan Y, Subba Rao D, Mann K, Li W, Harrison W. 1996 Effects of silicate limitation on
310 production of domoic acid, a neurotoxin, by the diatom *Pseudo-nitzschia multiseries*. II.
311 Continuous culture studies. *Mar. Ecol. Prog. Ser.* **131**, 235–243.
312 (doi:10.3354/meps131235)
- 313 11. Tammilehto A, Nielsen TG, Krock B, Møller EF, Lundholm N. 2012 *Calanus* spp.—
314 Vectors for the biotoxin, domoic acid, in the Arctic marine ecosystem? *Harmful Algae*
315 **20**, 165–174. (doi:10.1016/j.hal.2012.10.004)
- 316 12. Tammilehto A, Nielsen TG, Krock B, Møller EF, Lundholm N. 2015 Induction of
317 domoic acid production in the toxic diatom *Pseudo-nitzschia seriata* by calanoid
318 copepods. *Aquat. Toxicol.* **159**, 52–61. (doi:10.1016/j.aquatox.2014.11.026)
- 319 13. Harðardóttir S, Pančić M, Tammilehto A, Krock B, Møller E, Nielsen T, Lundholm N.
320 2015 Dangerous relations in the arctic marine food web: interactions between toxin
321 producing *Pseudo-nitzschia* diatoms and *Calanus* copepodites. *Mar. Drugs* **13**, 3809–
322 3835. (doi:10.3390/md13063809)
- 323 14. Zhang S, Zheng T, Lundholm N, Huang X, Jiang X, Li A, Li Y. 2021 Chemical and
324 morphological defenses of *Pseudo-nitzschia multiseries* in response to zooplankton
325 grazing. *Harmful Algae* **104**, 102033. (doi:10.1016/j.hal.2021.102033)
- 326 15. Bargu S, Lefebvre K, Silver MW. 2006 Effect of dissolved domoic acid on the grazing
327 rate of krill *Euphausia pacifica*. *Mar Ecol Prog Ser* **312**, 169–175.
328 (doi:doi:10.3354/meps312169)
- 329 16. Driscoll WW, Hackett JD, Ferrière R. 2016 Eco-evolutionary feedbacks between private
330 and public goods: evidence from toxic algal blooms. *Ecol. Lett.* **19**, 81–97.
331 (doi:10.1111/ele.12533)
- 332 17. Driscoll WW, Pepper JW. 2010 Theory for the evolution of diffusible external goods:
333 evolution of diffusible external goods. *Evolution* **64**, 2682–2687. (doi:10.1111/j.1558-
334 5646.2010.01002.x)
- 335 18. Selander E, Berglund EC, Engström P, Berggren F, Eklund J, Harðardóttir S, Lundholm
336 N, Grebner W, Andersson MX. 2019 Copepods drive large-scale trait-mediated effects in
337 marine plankton. *Sci. Adv.* **5**, eaat5096. (doi:10.1126/sciadv.aat5096)
- 338 19. Karban R. 2011 The ecology and evolution of induced resistance against herbivores.
339 *Funct. Ecol.* **25**, 339–347. (doi:10.1111/j.1365-2435.2010.01789.x)

- 340 20. Grover JP. 1995 Competition, herbivory, and enrichment: Nutrient-based models for
341 edible and inedible plants. *Am. Nat.* **145**, 746–774. (doi:10.1086/285766)
- 342 21. Guillard RRL, Hargraves PE. 1993 *Stichochrysis immobilis* is a diatom, not a
343 chrysophyte. *Phycologia* **32**, 234–236. (doi:10.2216/i0031-8884-32-3-234.1)
- 344 22. Hasle GR, Lundholm N. 2005 *Pseudo-nitzschia seriata* f. *obtusa* (Bacillariophyceae)
345 raised in rank based on morphological, phylogenetic and distributional data. *Phycologia*
346 **44**, 608–619. (doi:10.2216/0031-8884(2005)44[608:PSFOBR]2.0.CO;2)
- 347 23. Selander E, Kubanek J, Hamberg M, Andersson MX, Cervin G, Pavia H. 2015 Predator
348 lipids induce paralytic shellfish toxins in bloom-forming algae. *Proc. Natl. Acad. Sci.*
349 *USA* **112**, 6395–6400. (doi:10.1073/pnas.1420154112)
- 350 24. Lundholm N, Hansen P, Kotaki Y. 2004 Effect of pH on growth and domoic acid
351 production by potentially toxic diatoms of the genera *Pseudo-nitzschia* and *Nitzschia*.
352 *Mar. Ecol. Prog. Ser.* **273**, 1–15. (doi:10.3354/meps273001)
- 353 25. Xu J, Hansen PJ, Nielsen LT, Krock B, Tillmann U, Kiørboe T. 2017 Distinctly different
354 behavioral responses of a copepod, *Temora longicornis*, to different strains of toxic
355 dinoflagellates, *Alexandrium* spp. *Harmful Algae* **62**, 1–9.
356 (doi:10.1016/j.hal.2016.11.020)
- 357 26. Ryderheim F, Selander E, Kiørboe T. 2021 Predator-induced defence in a dinoflagellate
358 generates benefits without direct costs. *ISME J* **15**, 2107–2116. (doi:10.1038/s41396-021-
359 00908-y)
- 360 27. Krock B, Tillmann U, John U, Cembella A. 2008 LC-MS-MS aboard ship: tandem mass
361 spectrometry in the search for phycotoxins and novel toxigenic plankton from the North
362 Sea. *Anal Bioanal Chem* **392**, 797–803. (doi:10.1007/s00216-008-2221-7)
- 363 28. Olesen AJ, Ryderheim F, Krock B, Lundholm N, Kiørboe T. 2022 Data from: Costs and
364 benefits of predator-induced defense in a toxic diatom. *Zenodo*
365 (doi:10.5281/zenodo.6346438)
- 366 29. Bates D, Mächler M, Bolker B, Walker S. 2015 Fitting linear mixed-effects models using
367 lme4. *J. Stat. Softw.* **67**, 1–48. (doi:10.18637/jss.v067.i01)
- 368 30. Xu J, Kiørboe T. 2018 Toxic dinoflagellates produce true grazer deterrents. *Ecology* **99**,
369 2240–2249. (doi:10.1002/ecy.2479)
- 370 31. Lewis WM. 1986 Evolutionary Interpretations of Allelochemical Interactions in
371 Phytoplankton Algae. *Am. Nat.* **127**, 184–194. (doi:10.1086/284477)
- 372 32. Jonsson PR, Pavia H, Toth G. 2009 Formation of harmful algal blooms cannot be
373 explained by allelopathic interactions. *Proc. Natl. Acad. Sci. USA* **106**, 11177–11182.
374 (doi:10.1073/pnas.0900964106)
- 375 33. Teegarden G. 1999 Copepod grazing selection and particle discrimination on the basis of
376 PSP toxin content. *Mar. Ecol. Prog. Ser.* **181**, 163–176. (doi:10.3354/meps181163)

- 377 34. Leitão E, Ger KA, Panosso R. 2018 Selective grazing by a tropical copepod
378 (*Notodiaptomus iheringi*) facilitates *Microcystis* dominance. *Front. Microbiol.* **9**, 301.
379 (doi:10.3389/fmicb.2018.00301)
- 380 35. Ryderheim F, Grønning J, Kiørboe T. 2022 Thicker shells reduce copepod grazing on
381 diatoms. *Limnol. Oceanogr. Letters* , lol2.10243. (doi:10.1002/lol2.10243)
- 382 36. Olson M, Lessard E, Wong C, Bernhardt M. 2006 Copepod feeding selectivity on
383 microplankton, including the toxigenic diatoms *Pseudo-nitzschia* spp., in the coastal
384 Pacific Northwest. *Mar. Ecol. Prog. Ser.* **326**, 207–220. (doi:10.3354/meps326207)
- 385 37. Colin SP, Dam HG. 2002 Latitudinal differentiation in the effects of the toxic
386 dinoflagellate *Alexandrium* spp. on the feeding and reproduction of populations of the
387 copepod *Acartia hudsonica*. *Harmful Algae* **1**, 113–125. (doi:10.1016/S1568-
388 9883(02)00007-0)
- 389 38. Gustafsson S, Rengefors K, Hansson L-A. 2005 Increased consumer fitness following
390 transfer of toxin tolerance to offspring via maternal effects. *Ecology* **86**, 2561–2567.
391 (doi:10.1890/04-1710)
- 392 39. Colin SP, Dam HG. 2005 Testing for resistance of pelagic marine copepods to a toxic
393 dinoflagellate. *Evol. Ecol.* **18**, 355–377. (doi:10.1007/s10682-004-2369-3)
- 394 40. Peterson WT. 1985 Abundance, age structure and in situ egg production rates of the
395 copepod *Temora longicornis* in Long Island Sound, New York. *Bull. Mar. Sci.* **37**, 726–
396 738.
- 397 41. Dutz J, Mohrholz V, van Beusekom J. 2010 Life cycle and spring phenology of *Temora*
398 *longicornis* in the Baltic Sea. *Mar. Ecol. Prog. Ser.* **406**, 223–238.
399 (doi:10.3354/meps08545)
- 400 42. Hirche H-J. 1997 Life cycle of the copepod *Calanus hyperboreus* in the Greenland Sea.
401 *Mar. Biol.* **128**, 607–618. (doi:10.1007/s002270050127)
- 402 43. Arnkværn G, Daase M, Eiane K. 2005 Dynamics of coexisting *Calanus finmarchicus*,
403 *Calanus glacialis* and *Calanus hyperboreus* populations in a high-Arctic fjord. *Polar Biol*
404 **28**, 528–538. (doi:10.1007/s00300-005-0715-8)
- 405 44. Jiang X, Lonsdale DJ, Gobler CJ. 2011 Rapid gain and loss of evolutionary resistance to
406 the harmful dinoflagellate *Cochlodinium polykrikoides* in the copepod *Acartia tonsa*.
407 *Limnol. Oceanogr.* **56**, 947–954. (doi:10.4319/lo.2011.56.3.0947)
- 408 45. Avery DE, Dam HG. 2007 Newly discovered reproductive phenotypes of a marine
409 copepod reveal the costs and advantages of resistance to a toxic dinoflagellate. *Limnol.*
410 *Oceanogr.* **52**, 2099–2108. (doi:10.4319/lo.2007.52.5.2099)
- 411 46. Avery DE, Altland KK, Dam HG. 2008 Sex-related differential mortality of a marine
412 copepod exposed to a toxic dinoflagellate. *Limnol. Oceanogr.* **53**, 2627–2635.
413 (doi:10.4319/lo.2008.53.6.2627)
- 414 47. Tollrian R, Harvell CD. 1999 *The Ecology and Evolution of Inducible Defenses*.
415 Princeton, NJ: Princeton University Press.

- 416 48. Amato A *et al.* 2018 Grazer-induced transcriptomic and metabolomic response of the
417 chain-forming diatom *Skeletonema marinoi*. *ISME J.* **12**, 1594–1604.
- 418 49. Harðardóttir S, Wohlrab S, Hjort DM, Krock B, Nielsen TG, John U, Lundholm N. 2019
419 Transcriptomic responses to grazing reveal the metabolic pathway leading to the
420 biosynthesis of domoic acid and highlight different defense strategies in diatoms. *BMC*
421 *Mol. Biol.* **20**, 7. (doi:10.1186/s12867-019-0124-0)
- 422 50. Anderson DM, Cembella AD, Hallegraeff GM. 2012 Progress in understanding harmful
423 algal blooms: paradigm shifts and new technologies for research, monitoring, and
424 management. *Annu. Rev. Mar. Sci.* **4**, 143–176. (doi:10.1146/annurev-marine-120308-
425 081121)
- 426

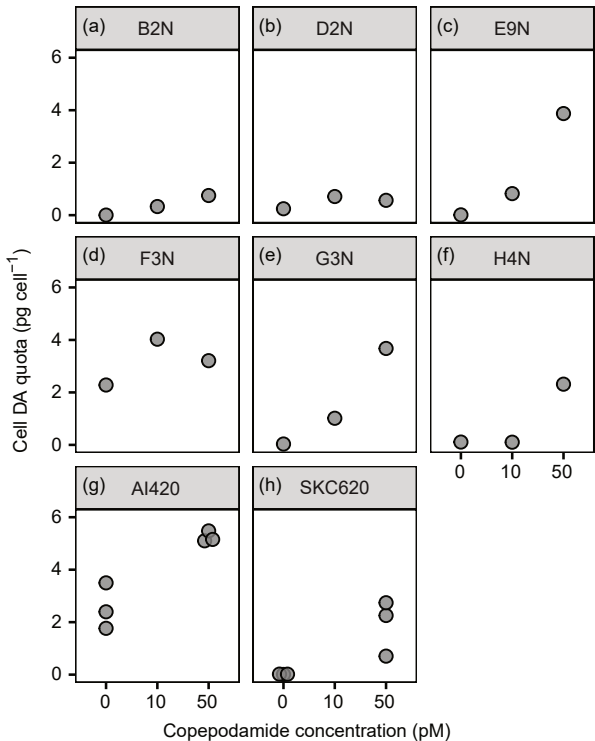
427 **Figure legends**

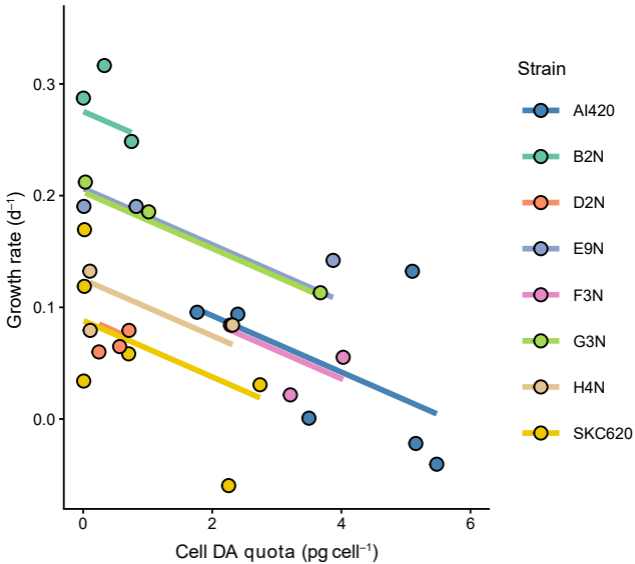
428 Figure 1. Dose-response experiments. Cellular domoic acid (DA) content (pg cell^{-1}) as a
429 function of the nominal copepodamide concentration for the eight different strains of *P. seriata*.
430 $n = 1$ for each concentration in a-f, and $n = 3$ for each concentration in g-h.

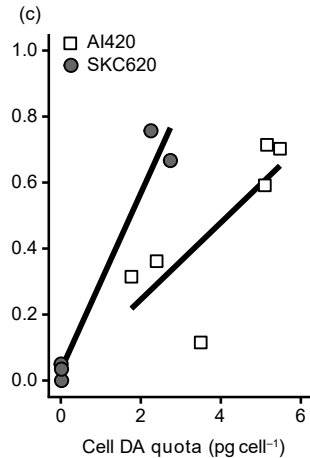
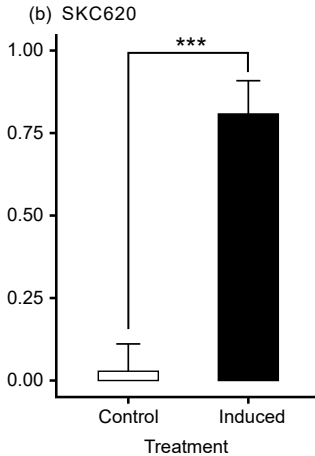
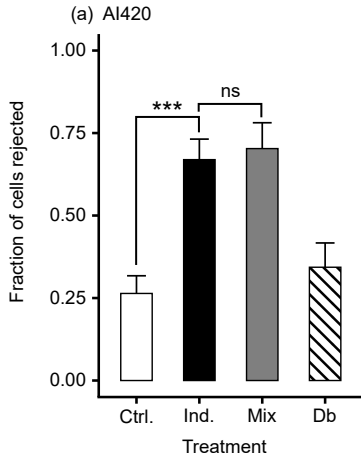
431 Figure 2. Growth-defence trade-off. Growth rate as a function of cellular DA content for all
432 strains used in this study. A linear mixed-effects model was fit to the data with one slope and
433 random intercepts. The estimated slope (with 95% confidence intervals) is $-0.025 \text{ d}^{-1} (\text{pg DA}$
434 $\text{cell}^{-1})^{-1} [-0.032, -0.018]$ ($F_{1,25.3} = 13.2, p < 0.001$).

435 Figure 3. Defence efficiency. The fraction of cells rejected in strain (a) AI420 and (b) SKC620.
436 Mix: the fraction of induced *P. seriata* cells rejected in the mixed experiment. Db: fraction of
437 *D. brightwellii* cells rejected in the mixed experiment. (c) Relationship between the fraction of
438 rejected cells and the cellular DA content. Values in (a) and (b) are the fractions from all
439 captures across several copepods recorded and error bars show 95% Wilson Score Interval (n
440 $= 48-287$). In (c) they are the fractions for each copepod and include the control and induced
441 treatments. One data point was removed in (c) due to the copepod only capturing two cells
442 (both were rejected). Asterisks indicate significant difference between treatments, **: $p < 0.01$,
443 ***: $p < 0.001$. ns: not significant. Odds ratios (with 95% confidence intervals) when
444 comparing 'Control' and 'Induced' treatments are (AI420) 5.71 [3.86, 8.54] and (SKC620)
445 83.99 [21.76, 563.50].

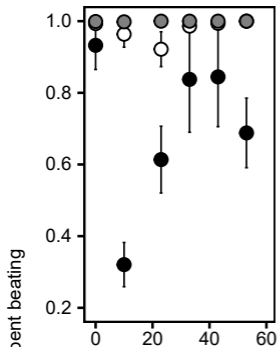
446 Figure 4. Copepod beating frequency. Fraction of time the copepods spent beating starting as
447 the cells are added (a, c) and averaged for each treatment (b, d) in the foraging experiments
448 with strains AI420 (top row) and SKC620 (bottom row). Lower case letters refer to significant
449 differences between treatments according to pairwise comparisons ($p < 0.05$). Colours referring
450 to treatments in the bar charts (b, d) apply also in (a) and (c). Error bars in all panels show
451 standard error ($n = 3$, except in 'Mix' where $n = 2$).



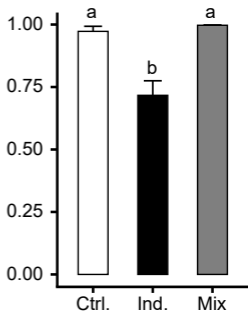




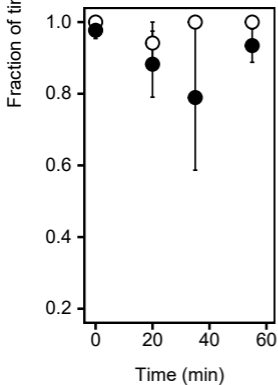
(a) AI420



(b)



(c) SKC620



(d)

